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# Continuous electron shuttling by sulfide oxidizing bacteria as a novel strategy to produce electric current

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## ABSTRACT

Sulfide oxidizing bacteria (SOB) are widely applied in industry to convert toxic  $H_2S$  into elemental sulfur. Haloalkaliphilic planktonic SOB can remove sulfide from solution under anaerobic conditions (SOB are 'charged'), and release electrons at an electrode (discharge of SOB). The effect of this electron shuttling on product formation and biomass growth is not known. Here, we study and demonstrate a continuous process in which SOB remove sulfide from solution in an anaerobic 'uptake chamber', and shuttle these electrons to the anode of an electrochemical cell, in the absence of dissolved sulfide. Two experiments over 31 and 41 days were performed. At a sulfide loading rate of 1.1 mmolS/day, electricity was produced continuously ( $3 \text{ A/m}^2$ ) without dissolved sulfide in the anolyte. The main end product was sulfate (56% in experiment 1% and 78% in experiment 2), and 87% and 77% of the electrons in sulfide were recovered as electricity. It was found that the current density was dependent on the sulfide loading rate and not on the anode potential. Biological growth occurred, mainly at the anode as biofilm, in which the deltaproteobacterial genus *Desulfurivibrio* was dominating. Our results demonstrate a novel strategy to produce electricity from sulfide in an electrochemical system.

# 1. Introduction

Hydrogen sulfide ( $H_2S$ ) is a toxic and corrosive compound which can be present in several types of gas streams, such as natural gas and biogas (Krayzelova et al., 2015; Wenhui et al., 2010). Combustion of  $H_2S$  leads to the formation of sulfur dioxide, causing acid rain (Burns et al., 2016). For sustainable control of sulfur emissions, a biotechnological gas desulfurization process has been developed, which recovers biologically formed elemental sulfur (S<sup>0</sup>) (Buisman et al., 1989; Janssen et al., 1995). This process has been applied on commercial scale since the early nineties (Janssen et al., 2009; Klok et al., 2017).

The process solution consists of a sodium carbonate/bicarbonate buffer of  $\sim 1 \text{ M Na}^+$  and the pH varies from 7.5 to 9.5. Sulfide is oxidized to (predominantly) S<sup>0</sup> by haloalkaliphilic sulfide-oxidizing bacteria (SOB), with O<sub>2</sub> as final electron acceptor. Several sustainable aspects of this process are: (i) operation under ambient pressure and temperature;

(ii) no requirement of toxic chemicals, and (iii)  $S^0$  can easily be harvested from the process solution for reuse, for example as fertilizer (Van Zessen et al., 2004). Besides  $S^0$ , sulfate ( $SO_4^{2-}$ ) and thiosulfate ( $S_2O_3^{2-}$ ) are formed, which are unwanted because it requires the addition of NaOH and, as a consequence, the formation of a bleed stream (De Rink et al., 2019; Klok et al., 2012; Van Den Bosch et al., 2007). The formation of  $SO_4^{2-}$  and  $S_2O_3^{2-}$  is stimulated at elevated  $O_2$  concentrations and therefore the  $O_2$  is controlled at very low concentrations.

To make the process more sustainable, recent studies focused on the reduction of chemical consumption (De Rink et al., 2019; Kiragosyan et al., 2020; Van Den Bosch et al., 2007). Another important aspect of the sustainability of biological gas desulfurization is the energy consumption, which is the focus of this study. In the current process, most of the energy consumption is related to the aeration of the bioreactor for  $O_2$  supply. In addition, the process solution has to be cooled due to the heat released by the exothermic oxidation of sulfide. Therefore, the enthalpy

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of the oxidation of sulfide is lost.

Recently, it has been found that the SOB taken from the biodesulfurization process are electroactive. These planktonic bacteria can use an electrode as final electron acceptor, instead of O<sub>2</sub> (ter Heijne et al., 2018). Furthermore, it was demonstrated that these SOB can remove dissolved sulfide from solution under anaerobic conditions (de Rink et al., 2020; De Rink et al., 2019; ter Heijne et al., 2018). If this principle can be applied in a continuous electrochemical process for the removal of sulfide, this would have several benefits: (i) no energy is required for aeration (O2 supply), and (ii) electrons are harvested from sulfide oxidation, which can be used to recover energy (e.g. in the form of H<sub>2</sub>, which can be formed at the cathode). Bioelectrochemical sulfide removal has been studied extensively (Ni et al., 2019; Rabaey et al., 2006; Sun et al., 2009; Zhang et al., 2009). However, one of the major drawbacks is the deposition of sulfur on the anode when the anode is contacted with dissolved sulfide directly. This deposition occurs due to an electrochemical reaction and will eventually lead to electrode passivation (Ateya et al., 2003; Dutta et al., 2008; Vaiopoulou et al., 2016). Thus, the challenge is to prevent direct contact between dissolved sulfide and the anode, which can be achieved by utilizing the shuttling capacity of SOB, as described by ter Heijne et al. (2018). This electron shuttling capacity of SOB is still poorly understood. In the batch experiments, the product formation and biomass growth could not be monitored (ter Heijne et al., 2018).

The aim of this study is to investigate a continuous bioelectrochemical process for the removal and conversion of sulfide using the sulfide shuttling capacity of SOB in the absence of  $O_2$ . SOB are continuously circulated between a 'sulfide uptake reactor', to which sulfide is continuously added and where SOB can remove the dissolved sulfide from solution anaerobically, and the anode of an electrochemical cell, where electrons are released in the absence of dissolved sulfide. By shuttling SOB between this sulfide uptake reactor and the electrochemical cell, the sulfide removal and release of electrons is spatially separated. Furthermore, biomass growth and composition are studied to understand the underlying mechanisms of the overall process.

## 2. Materials and methods

## 2.1. Experimental set-up

Experiments were performed in a system consisting of an electrochemical cell, a sulfide uptake reactor (connected to the anode side of the electrochemical cell) and a catholyte recirculation reactor (connected to the cathode side), see Fig. 1. The liquid volume of the sulfide uptake reactor and catholyte recirculation reactor were 400 mL each. The working electrode (anode) was a graphite plate and the counter electrode a titanium plate coated with Pt/IrO<sub>2</sub> (Magneto, special anodes, Schiedam, the Netherlands), both with an effective surface area of 22 cm<sup>2</sup>. A cation exchange membrane (FumaTech GmbH, Germany) was used to separate anode and cathode compartments, both with a volume of 33 mL. The anode compartment was filled with graphite granules (approximately 27 g) with a size of 2-4 mm (enViro-gran typ 514, enViro-cell, Oberursel, Germany) and the cathode compartment was filled with approximately 100 glass beads of 6-7 mm. An Ag/AgCl, 3 M KCl reference electrode (ProSense/QIS, the Netherlands) was connected to the anode chamber via a capillary. A potentiostat (nStat, multichannel, Ivium, Eindhoven, the Netherlands) was used to control the anode potential and record the current.

First, the anode and cathode, including the recirculation reactors, were filled with an NaHCO<sub>3</sub> solution of approximately 1 M and liquid circulation was started. 1 M is a typical Na<sup>+</sup> concentration of the (full-scale) biodesulfurization process. To ensure anaerobic conditions in the system, the headspaces of recirculation reactors were continuously purged with a N<sub>2</sub>/CO<sub>2</sub> gas stream of 4 L/h, dosed via Mass Flow controllers (Brooks). The pH was controlled by adjusting the ratio between the N<sub>2</sub> and CO<sub>2</sub> flow rates, which was approximately 4:1. After the O<sub>2</sub> concentration was below the detection limit (15 ppb), SOB biomass from an operating biodesulfurization installation was added to the sulfide uptake reactor. This installation is described by de Rink et al. (De Rink et al., 2019) and operated under similar conditions when the SOB were harvested. Approximately 10 mL inoculum was prepared by washing the

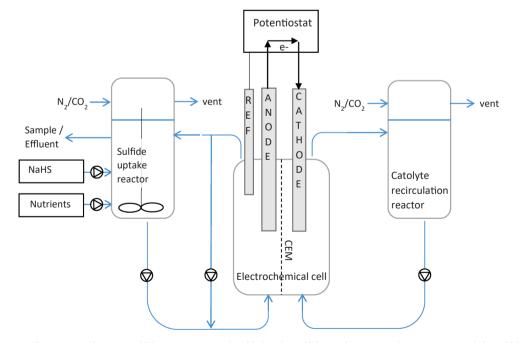


Fig. 1. Schematic drawing of experimental set-up. Sulfide was continuously added to the sulfide uptake reactor where SOB removed the sulfide from solution. The solution was circulated over the anode side of the electrochemical cell, where electricity was produced.

biomass to remove sulfur particles and (thio)sulfate. The biomass was centrifuged at 10,000 rcf and the biomass pellet was resuspended in 1 M NaHCO<sub>3</sub> solution. These steps were repeated 2 times.

Due to the anolyte liquid circulation, the planktonic SOB were continuously transferred from the sulfide uptake reactor to the anode side of the electrochemical cell and back. After about 1 h of liquid circulation, the dosing of NaHS and nutrients was initiated and the potentiostat was started. Gas bags filled with N<sub>2</sub> gas were connected to the headspaces of the NaHS and nutrient stock bottles in order to keep these anaerobic. NaHS was prepared by dissolving NaHS·xH<sub>2</sub>O (hydrosulfide hydrate pure flakes, Acros Organics) in demineralized water, which was extensively flushed with N<sub>2</sub>. The sulfide concentration was verified by potentiometric titration with 0.1 M AgNO<sub>3</sub> using a Titrino-Plus titrator (Metrohm).

## 2.2. Operation of the continuous set-up

Two biotic (experiments 1 and 2) and one abiotic (experiment 3) were performed. In each experiment, the anode potential was changed stepwise and each potential was maintained for at least 4 days. All anode potentials are reported versus Ag/AgCl, 3 M KCl.

## 2.2.1. Experiment 1

Experiment 1 had a total duration of 31 days and the applied anode potentials were -0.1, -0.2, -0.3, -0.35 and -0.4 V. The NaHS concentration in the stock solution was 20.45 gS/L and was continuously dosed into the sulfide uptake reactor with a flow rate of 1.66 g/day, resulting in a constant sulfide loading rate of 1.06 mmolS/day. The circulation flow rate over the sulfide uptake reactor was 0.6 L/h, which resulted in an HRT of 40 min in the sulfide uptake reactor. Nutrients solution consisted of macro nutrient as described by de Rink et al. (De Rink et al., 2019) and 1 mL/L trace element mix as described by Pfennig and Lippert (1966). The nutrient dosing rate was set in such way that the total residual nitrogen concentration in the supernatant was <10 mgN/L. In this way, overdosing of nutrients is prevented and the bacteria consume all nutrients. This dosing strategy is similar to full-scale biodesulfurization installations.

 $H_2$  cross-over from cathode to anode could influence the current production. To rule out effects of  $H_2$  cross-over, on day 15 of experiment 1,  $[Fe(CN)_6]^{3-}$  was added to the catholyte recirculation reactor. After addition of  $[Fe(CN)_6]^{3-}$ , the cathode potential changed, but the current did not change, indicating that  $H_2$  cross-over did not influence the anodic oxidation reaction, and thus the current, in the electrochemical cell. However, from day 16 onwards, it was noticed that total and dissolved N concentrations increased, even after nutrient dosing was stopped at day 21. Probably some  $[Fe(CN)_6]^{3-}$  leaked through the membrane into the anode compartment. From day 16 onwards, the difference between total-N and dissolved-N concentrations still represents the concentration of planktonic bacteria, but an accurate N-balance, including calculation of the amount of biofilm, could not be made.

## 2.2.2. Experiment 2

Experiment 2 had a total duration of 41 days and the applied anode potentials were -0.1, -0.2, -0.3, and -0.4 V. The NaHS concentration in the stock solution was lowered to 8.90 gS/L in order to have a higher inflow rate. Furthermore, the nutrients were diluted 2x and the dosing rate was doubled compared to experiment 1. The sulfide loading rate in experiment 2 varied in order to find the maximal sulfide uptake capacity of the bacteria. Initial loading rate was 0.79 mmolS/day. From day 12–29, the loading rate was increased to 1.16 mmolS/day; from day 30–42, the sulfide loading rate was again 0.79 mmolS/day. The

recirculation flow rate over the anolyte recirculation vessel was decreased to 0.42 L/h, resulting in an HRT of 58 min.

## 2.2.3. Experiment 3

Experiment 3 was an abiotic experiment, performed after experiment 2. Despite thorough cleaning of the system and replacement of the graphite granules, some biological activity was observed in this abiotic experiment. Therefore, approximately 0.1 M sodium azide (NaN<sub>3</sub>) was added to the system, after which the system was considered abiotic. The effect of NaN<sub>3</sub> on the current production was verified with cyclic voltammetry in a separate electrochemical cell and no effect on the current was observed in the potential range of our experiments. The abiotic experiment was performed for 10 days and the applied anode potentials were -0.1 and -0.4 V. The sulfide loading rate was 0.99 mmolS/day and no nutrients were dosed. The liquid circulation rate in experiment 3 was also 0.42 L/h, corresponding to an HRT in the anolyte recirculation vessel of 58 min.

## 2.2.4. Other experimental conditions

An overview of the flow rates, volumes in the different parts of the system and the HRT's in the different parts of the system in each experiment is provided in the Supplementary Material (A). During all 3 experiments, the pH was controlled around 8.3 by adjusting the  $CO_2$  flow rate to the headspaces of the recirculation reactors. The initial alkalinity (i.e. bicarbonate concentration) was 1.0 M and conductivity was 52 mS/cm. Due to acidification and dilution, the alkalinity and conductivity decreased during the experiments. For example, at the end of the experiment 1, the alkalinity and conductivity were 0.37 M and 31 mS/cm. Under the applied pH (~8.3), practically all sulfide is present as HS<sup>-</sup>. Based on the measured current and the product formation, the electron balance is reasonable, which confirms that release of gaseous H<sub>2</sub>S is negligible.

The thermodynamics of any (bio)electrochemical system is determined by both anode and cathode reactions and their corresponding potentials. In our experiments, the cathode potential was approximately -0.8 V to form H<sub>2</sub>. This consumed electricity, but energy is recovered in the form of hydrogen gas.

## 2.3. Analyses

Daily samples were taken from the sulfide uptake reactor. A sample from the catholyte recirculation reactor was taken at the end of the experiments, pH and conductivity were analyzed offline using a HO440d multi analyzer (Hach, Germany). Alkalinity, expressed as [HCO3<sup>-</sup>] was measured with an automated TitrinoPlus titrator (Metrohm) by titrating to pH 4.3 using a 0.1 M HCl solution.  $SO_4^{2-}$  and  $S_2O_3^{2-}$  concentrations were determined on the samples supernatant (after centrifuging for 10 min at 14,000 g) using a Dionex ICS-2100 Ion Chromatograph (ThermoScientific) with a Thermo Fisher Scientific IonPac AG17 Guard (Thermo Fisher Scientific, Waltham, MA, USA) and Thermo Fisher Scientific IonPac AS17 column (Thermo Fisher Scientific) at 30 °C. The eluent was KOH at a flowrate of 1.0 mL/min. The sample injection volume was 10  $\mu L.$  The biomass concentration was measured as the amount of total organic N using the Dr. Lange cuvette test LCK138 (Hach Lange, Germany), as described by De Rink et al. (2019). The difference between the supernatant (i.e., a sample centrifuged for 10 min at 14, 000 g) and a non-centrifuged sample indicated the total amount of N present in the (planktonic) biomass. To exclude interference by salts and biologically produced S<sup>0</sup>, the samples were diluted at least 5 times. Based on the amount of N added with the nutrients and the measured N concentrations, an N-balance was made in order to calculate the amount of biomass attached to the electrode (biofilm), see Eq. (7). The dissolved sulfide concentration in the sulfide uptake reactor was measured using the Hach Lange cuvette test LCK635 (Hach Lange, Germany). Furthermore, lead acetate paper (Merck) was used to check presence/absence of dissolved sulfide. The  $O_2$  concentration in the headspaces of the anode and cathode chambers was measured with oxygen sensor spots and Fibox 4 trace meter (PreSens, Regensburg Germany).

During each run, the experiment was occasionally interrupted to make polarization curves (determine current production at a range of anode potentials). The applied potentials were -0.5, -0.45, -0.4, -0.35, -0.3, -0.2 and -0.1 V (vs Ag/AgCl) and current was measured each second. Each potential was maintained for 5 min and the last measurement for each potential. During the polarization curves, the sulfide dosing was maintained. The microbial communities of the planktonic bacteria and biofilm were analyzed using 16S amplicon sequencing. Samples were taken from the liquid and the graphite granules at the end of each run. Also a sample of the inoculum (planktonic biomass) was analyzed. Details of the analysis can be found in the

$$P_{S_8-S} = I_{H_2S} - P_{SO_4^{2-}-S} - P_{S_2O_3^{2-}-S}$$
(3)

Here,  $P_{S_8-S}$ ,  $P_{SO_4^{2^-}-S}$  and  $P_{S_2O_3^{2^-}-S}$  are the productions of S<sup>0</sup>, SO<sub>4</sub><sup>2-</sup>, and S<sub>2</sub>O<sub>3</sub><sup>2-</sup>, respectively, in mol S-product and  $I_{HS^-}$  is the volumetric HS<sup>-</sup> influent in mol S. Overall product selectivities were:

$$Overall S_{SO_4^{2-}-S}\left(\%\right) = \frac{Total P_{SO_4^{2-}-S}\left(mol S\right)}{Total I_{HS^-}(mol S)}$$
(4)

$$Overall S_{S_2 O_3^{2^-} - S}(\%) = \frac{Total P_{S_2 O_3^{2^-} - S} (mol S)}{Total I_{HS^-} (mol S)}$$
(5)

$$Overall S_{S_8-S}\left(\%\right) = \frac{Total P_{S_8-S} (mol S)}{Total I_{HS^-}(mol S)}$$
(6)

The amount of biofilm, expresses as mg-N per liter of the total system volume was calculated as the difference between the measured total-N concentration and the theoretical total-N concentration. The theoretical total-N concentration ( $N_{theor}$ ) was calculated according to Eq. (7):

$$N_{\text{theor. day }x}\left(\frac{\text{mgN}}{L}\right) = N_{\text{theor. day }x-1} + \left(\frac{\text{N added with nutrients} - (\text{measured total} - \text{N} \cdot \text{Effluent})}{\text{V}}\right)$$
(7)

Supplementary Material (B). The EMBL-EBI accession number for the presented 16S rRNA sequencing set is PRJEB44164.

At the end of the experiments, presence of elemental sulfur on the electrode was verified using XRD analysis. Details can be found in the Supplementary Material (C).

#### 2.4. Calculations

The production of  ${SO_4}^2\mbox{-}S$  (Eq. 1) and  $S_2O_3^{\ 2}\mbox{-}S$  (Eq. 2) were calculated as follows:

$$P_{SO_4^{2-}-S} = effluent \cdot \overline{[SO_4^{2-}-S]} + V \cdot \Delta \left[SO_4^{2-}-S\right]$$
(1)

$$P_{S_2O_3^{2^-}-S} = effluent \cdot \overline{[S_2O_3^{2^-}-S]} + V \cdot \Delta \left[S_2O_3^{2^-}-S\right]$$
(2)

Here, *effluent* is the effluent of the system (i.e. the sample volume from the sulfide uptake reactor) (L),  $[\overline{SO_4^{2^-} - S}]$  and  $[\overline{S_2O_3^{2^-} - S}]$  are the average concentrations (mol-S L<sup>-1</sup>) of two consecutive samples, *V* is the total liquid volume of the system and  $\Delta[SO_4^{2^-} - S]$  and  $\Delta[S_2O_3^{2^-} - S]$  are

centration in the nutrients x the amount of dosed nutrients between  $day_{x-1}$  and  $day_x$ . *Effluent* is the amount of effluent between  $day_{x-1}$  and  $day_x$ . Based on the formed products, the total amount of released electrons

In Eq. (7), N added with nutrients is calculated as the total-N con-

was calculated. The amount of  $e^-$  released per S<sup>0</sup> formed, is 2 (Eq. 8); for each mol of SO<sub>4</sub><sup>2-</sup> produced, an additional amount of 6 mol  $e^-$  is released, i.e. a total of 8 mol  $e^-$  (Eq. 9).

$$HS^{-} \rightarrow S^{0} + 2 e^{-} + H^{+}$$
 (8)

$$HS^{-} + 4H_2O \xrightarrow{\rightarrow} SO_4^{2-} + 9H^+ + 8 e^-$$
 (9)

The amount of electrons harvested at the anode was determined based sum of the current (as logged by the potentiostat). Based on the  $SO_4^{2-}$  formation, the amount of e<sup>-</sup> used for biomass growth was calculated based on the work of Klok et al. (2012). Based on this electron balance, the coulombic efficiency (CE) was calculated excluding biomass growth (Eq. 10) and including biomass growth (Eq. 11).

(10)

(11)

$$CE_{excl biomass} \left(\%\right) = \frac{e^{-} harvested at anode}{e^{-} released by S^{0} production + e^{-} released by SO_{4}^{2} - production}$$

$$CE_{incl biomass} \left(\%\right) = \frac{e^{-} harvested at anode + e^{-} to biomass growth}{e^{-} released by S^{0} production + e^{-} released by SO_{4}^{2} - production}$$

the concentration changes (mol-S L<sup>-1</sup>) between the samples. At the end of the experiments, some  $SO_4^{2^-}$  was detected in the catholyte (<5% of the concentration in the anolyte). Hence this was not taken into account in the calculation for the  $SO_4^{2^-}$  production.

As no products other than  $S_8$ ,  $SO_4^{2-}$  and  $S_2O_3^{2-}$  were measured in the reactor, and it was not possible to quantify the amount of  $S_8$  (Van Den Bosch et al., 2007), production of  $S_8$  was calculated from the mass balance:

# 3. Results and discussion

Two experiments were performed, in which planktonic bacteria continuously removed the supplied sulfide from solution in the sulfide uptake reactor and subsequently released electrons at the anode of the (bio)electrochemical cell. The results of these two experiments are

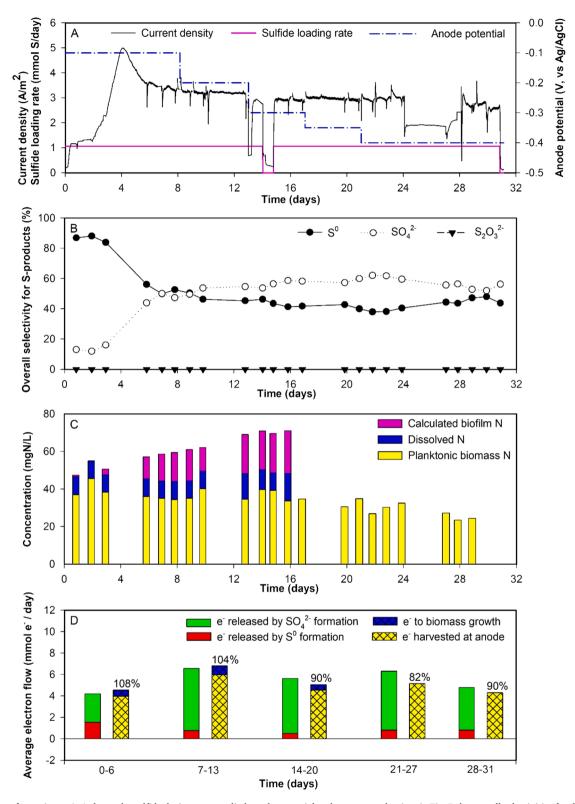


Fig. 2. Results of experiment 1. A shows the sulfide dosing rate, applied anode potential and current production; in Fig. B the overall selectivities for the S-products is shown; C shows the measured planktonic biomass concentration and the calculated concentration of biofilm (based on the total system's volume); in D the electron balance is shown.

shown in Fig. 2 (experiment 1) and 3 (experiment 2), and the performance of the reactors will be described and discussed in the following chapters.

# 3.1. Continuous electric current production in absence of dissolved sulfide

Figs. 2A and 3A show the sulfide loading rate, the applied anode potential and the resulting current density. In both experiments, current density increased during the first days of the experiment and reached a

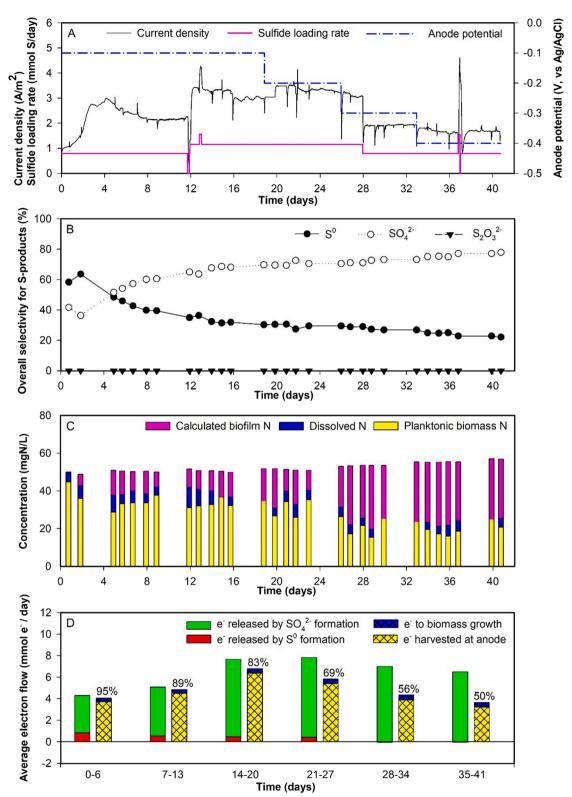


Fig. 3. Results of experiment 2. A shows the sulfide dosing rate, applied anode potential and current production; in Fig. B the overall selectivities for the S-products is shown; C shows the measured planktonic biomass concentration and the calculated concentration of biofilm (based on the total system's volume); in D the electron balance is shown.

peak at day 5. Thereafter, current decreased and stabilized.

In experiment 1, the sulfide load was constant at 1.06 mmol/day (see Fig. 2A). Throughout the experiment, all supplied sulfide was converted. In the first 3 days, the current density was approximately 1 A/m<sup>2</sup>. In this period, sulfide was converted into predominantly S<sup>0</sup> (see Fig. 2B). After

the third day, the current density increased and eventually stabilized around 3 A/m<sup>2</sup> at day 6. On day 3–6, the selectivity for SO<sub>4</sub><sup>2-</sup> increased and consequently, the selectivity for S<sup>0</sup> decreased. Since SO<sub>4</sub><sup>2-</sup> formation releases more electrons than S<sup>0</sup> formation, the increase in the current density was related to the increase in SO<sub>4</sub><sup>2-</sup> formation. The product

formation is discussed in more detail in Section 3.2. The theoretical dissolved sulfide concentration in the sulfide uptake reactor was 0.07 mM, based on sulfide loading and recirculation flow rate. However, no dissolved sulfide was detected in the analyte during the entire run, except for days 3 and 14 where some dissolved sulfide was detected (see Supplementary Material D). Thus, sulfide was removed by SOB in the sulfide uptake reactor, and the average sulfide uptake was 2.06 µmol/ mgN, based on an average biomass concentration of 34 mgN/L. At day 14, an unstable current was measured, and dissolved sulfide was detected in the anolyte. Therefore, the sulfide dosing was stopped overnight and the cable connections between the potentiostat and the three electrodes were verified. The following day, the sulfide dosing was restarted and no more dissolved sulfide was detected in the anolyte. In addition, the current density stabilized at previously obtained value of 3 A/m<sup>2</sup>. From day 24–27 the measured current density was lower, due to an offset of the reference electrode. After replacement of the reference electrode, the measured current density returned to  $3 \text{ A/m}^2$ . In this period, sulfide uptake in the sulfide uptake reactor continued, and no dissolved sulfide was detected in the sulfide uptake reactor.

Also throughout experiment 2, all supplied sulfide was converted. Experiment 2 was started with a sulfide loading rate of 0.79 mmolS/day. The initial current density was 1  $A/m^2$ , which stabilized around 2  $A/m^2$ at day 8 (see Fig. 3A). In the first days of experiment 2, the same pattern as in experiment 1 was observed: Initially, the selectivity for S<sup>0</sup> formation was high. When  $SO_4^{2^2}$  formation increased, the current density became higher. No dissolved sulfide was detected in the sulfide uptake reactor until day 12. Therefore, at day 13 the sulfide load was increased to 1.16 mmolS/day, which resulted in an increase in the current density, and still no dissolved sulfide was detected in the anolyte. Hence, the sulfide uptake was 3.5 µmol/mgN during this period, which was 32% higher than in experiment 1. After a second increase to 1.56 mmolS/day, the current density reached a value of  $3.84 \text{ A/m}^2$ . However, after the second increase, dissolved sulfide was detected with lead acetate paper (not quantified) and the sulfide load was reduced to 1.16 mmolS/day. At this dosing rate, the measured current density was around  $3 \text{ A/m}^2$ , similar to the current density measured in experiment 1. From day 20 onwards, detectable quantities of dissolved sulfide were present in the anolyte. However, the majority of sulfide was still being removed (Supplementary Material D). At day 28, the sulfide loading rate was decreased to the initial value of 0.79 mmolS/day, resulting in a current of 1.8 A/m<sup>2</sup>.

In both runs, the initial applied anode potential was -0.1 V vs Ag/ AgCl. From previous batch experiments, this anode potential was found to be sufficiently high to recover electrons from 'charged' SOB (ter Heijne et al., 2018). During the experiment, polarization curves were recorded (results discussed in Section 3.5). From these curves, it appears that current production is possible at anode potentials <-0.1 V vs Ag/AgCl. Theoretically, a lower anode potential results in a lower driving force for electron transfer. When the driving force is too low, no current can be generated. To study the effect of anode potential, the applied anode potential was stepwise decreased to -0.4 V in both experiments. Surprisingly, this decrease did not have an effect on the current density. In our experiments, decreasing the anode potential from -0.1 V to -0.4 V did not result in an increase in the selectivity for S<sup>0</sup> formation. Since the current density at constant sulfide loading rate is dependent on the product formation (i.e.  $S^0$  and  $SO_4^2$ ), the current density did not change when the anode potential was changed. The current density did change when the sulfide loading rate was changed (in experiment 2). Thus, in addition to the product selectivity, also the substrate supply determined the measured current density.

# 3.2. Product formation

Figs. 2B and 3B show the overall product selectivities in experiments 1 and 2. In both experiments, no thiosulfate  $(S_2O_3^{2-})$  was detected. While  $S_2O_3^{2-}$  is a reaction product of chemical oxidation of dissolved sulfide

with oxygen (O'Brien and Birkner, 1977; Van den Bosch et al., 2008), its absence can be explained by the absence of dissolved sulfide and/or the absence of  $O_2$  in the anolyte. At the start of both experiments, the selectivity for S<sup>0</sup> formation was higher than for  $SO_4^{2-}$  formation. However, after a few days of operation,  $SO_4^{2-}$  rather than S<sup>0</sup> became the main end product of sulfide oxidation. Bacteria prefer  $SO_4^{2-}$  formation over S<sup>0</sup> formation due to the higher Gibbs free energy. In the conventional biological desulfurization process,  $SO_4^{2-}$  formation is minimized by limiting the  $O_2$  supply (Klok et al., 2012; Van Den Bosch et al., 2007). The product formation also depends on the microbial community, which is discussed in Section 3.6. In this section, the mechanisms of  $SO_4^{2-}$  formation are further discussed.

For experiment 1, the overall selectivity for SO<sub>4</sub><sup>2-</sup> formation was 56% and the selectivity for S<sup>0</sup> formation 44%. For experiment 2, the selectivity for SO<sub>4</sub><sup>2-</sup> was 78% and selectivity for S<sup>0</sup> 22%. From day 28 onwards, all HS dosed was converted into SO42-. In none of the experiments, S<sub>8</sub>-sulfur was detected on the granules or graphite plate, as confirmed with XRD analysis. Occasional analysis of the total sulfur concentration in the anolyte with inductively coupled plasma optical emission spectrometry showed that the total sulfur concentration was higher than the sum of the  $SO_4^{2-}S$  and  $S_2O_3^{2-}S$  concentration as analyzed with ion chromatography, indicating that also elemental sulfur was present in the anolyte. The overall obtained selectivities for S<sup>0</sup> formation are similar to the results of Ni et al. (Ni et al., 2019) who obtained a selectivity for S<sup>0</sup> formation of 40% in a fed-batch reactor set-up with a similar electrochemical cell. In the conventional biological desulfurization process a selectivity for sulfur formation of 97% has been achieved (De Rink et al., 2019).

## 3.3. Biomass growth

The amount of biomass in the system is shown in Fig. 2C (experiment 1) and 3C (experiment 2). The yellow bars represent the biomass of planktonic cells. The blue bars show the measured dissolved N concentration, which is most likely resulting from non-consumed nutrients in the form of urea and/or ammonium. The concentration of planktonic biomass slightly decreased over the course of both experiments (i.e. the yellow bars in Figs. 2C and 3C). This means that the loss of planktonic bacteria via the effluent is higher than the increase of planktonic bacteria via bacterial growth. Based on the nutrient dosing rate, a higher total nitrogen concentration in the effluent was expected. For example, at the end of run 2, the measured biomass was about 50% lower than expected based on the supply of nutrients. Thus, there is another nitrogen sink, which is most likely the formation of a biofilm on the anode. The amount of biofilm could be calculated based on the addition of nutrients and the measurement of the nitrogen concentration in the anolyte. This amount, expressed as mgN/L of the total system, is represented by the purple bars in the Figs. 2C and 3C. Hence, the height of the bars represents the total amount of N in the system, which consist of biofilm, planktonic bacteria and non-consumed nutrients.

From Figs. 2C and 3C it appears that growth of sulfur metabolizing bacteria occurred, i.e. the bulk of the dosed nutrients were consumed. The majority of the growth occurred at the anode in the form of biofilm. At the end of experiment 2, the amount of biofilm-N on the electrode was determined by measuring the total N concentration of the granules, and was found to be 10.85 mgN. This translates to 31.46 mgN/L based on the total volume of the system and is in agreement with the amount of biofilm calculated based on the N-balance (31.28 mgN/L).

#### 3.4. Electron balance

Based on the average values for produced current, product formation and biomass growth, electron balances were made for periods of a week of operation. Results are shown in Figs. 2D and 3D.

For experiment 1 (Fig. 2D), during day 0–6, the average rate of electrons released was 4.2 mmol e /day, of which 37% was coming from

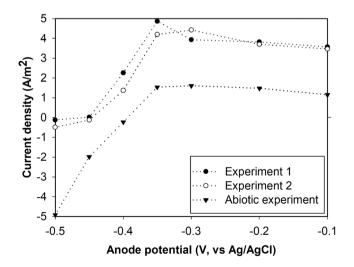
oxidation of HS<sup>-</sup> to S<sup>0</sup> and 63% from the oxidation of HS<sup>-</sup> to SO<sub>4</sub><sup>2-</sup>. In this period, the rate of harvesting electrons at the anode was 4.0 mmol e<sup>-</sup>/ day, which was 94% of the e<sup>-</sup> released from HS<sup>-</sup>. When taking into account the calculated e<sup>-</sup> used for biomass growth (0.16 mmol e<sup>-</sup>/day), the coulombic efficiency was 108%. The coulombic efficiencies varied over the time periods. The overall coulombic efficiency during experiment 1 was 87%.

During experiment 2, the coulombic efficiencies were around 90% (up to day 20). Towards the end of the experiment the coulombic efficiency decreased for unknown reasons and was 56% in the period of day 35–41. In this period it was noticed that the formation of  $SO_4^{2-}$  was slightly higher than the amount of HS<sup>-</sup> supplied. This can be explained by  $SO_4^{2-}$  production from S<sup>0</sup> which has been formed in the initial stage of the experiment. As a result, from day 28–41, the net production of S<sup>0</sup> formation is negative. Hence, the e<sup>-</sup> released by S<sup>0</sup> formation (the red bar in Fig. 3D) in this period is expressed as a negative number. The overall coulombic efficiency during experiment 2 was 77%. The coulombic efficiencies obtained in the experiments are higher than reported by Ni et al. (48.4%), who used a similar medium, but higher sulfide concentrations (Ni et al., 2019).

## 3.5. Biotic vs abiotic operation

To study the effect of the presence of SOB on HS' removal, an abiotic control experiment was performed at anode potentials of -0.1 V and -0.4 V. In this experiment, sodium azide (NaN<sub>3</sub>) was added to the system. Azide is a general agent to block growth of aerobic bacteria by the inhibition of the terminal cytochrome c oxidases of their respiratory chain (Lichstein and Soule, 1944; Solioz et al., 1982). Addition of NaN<sub>3</sub> had a direct effect on the current density. After each addition, the current density dropped. After the first addition (1.7 g/L), the current density dropped, but started to increase after several hours. Therefore, more NaN<sub>3</sub> was added (total 6.8 g/L) until no direct effect on current density was observed anymore; at this moment the current was almost zero. After the 4th addition, the experiment was considered abiotic. It was confirmed in a separate batch test that azide did not react with the electrode under the applied process conditions.

Under the abiotic conditions, dissolved sulfide was present in the sulfide uptake reactor at all times. For the abiotic experiment the system was operated at an anode potential of -0.1 V for 4 days. During this period, a current density of 0.46 A/m<sup>2</sup> was obtained at a sulfide dosing



**Fig. 4.** Typical polarization curves for the biological runs 1 and 2 and the abiotic control experiment. Polarization curves were made by stepwise increasing the anode potential (from -0.5 to -0.1 V) and measuring the current. Each potential was maintained for 5 min to obtain a stable current at the applied potential. The plotted values are the last measured value.

rate of 0.99 mmolS/day. This is considerably lower than the 3 A/m<sup>2</sup> that was measured in the presence of bacteria. At the same time, sulfide levels increased in the anolyte during the abiotic experiment, indicating that sulfide uptake in the sulfide uptake reactor only occurred with planktonic bacteria. Both  $SO_4^{2-}$  and  $S_2O_3^{2-}$  were produced at rates of 0.14 mmol/day and 0.08 mmolS/day. Interestingly, S<sub>2</sub>O<sub>3</sub><sup>2-</sup> was not detected in the biotic experiments because no dissolved sulfide was present. At an anode potential of -0.4 V, the current density was negative  $(-0.1 \text{ A/m}^2)$ . This means that electrons flowed in reverse direction, i.e. from counter electrode to working electrode. Biotic and abiotic polarization curves are shown in Fig. 4. For the biotic experiments, a slightly negative current was observed at anode potentials of -0.5 and -0.45 V. For higher anode potentials, current density was positive. Under abiotic conditions, the current density was still negative at an anode potential of -0.4 V and positive at potentials of -0.35 V and higher. A further increase in anode potential did not result in higher current density. Between -0.35 and -0.1 V the anode potential did not influence the current. Under these conditions, the current density was determined by sulfide dosing and not by the anode potential.

## 3.6. Microbial community

The microbial community of the planktonic bacteria and the biofilm was analyzed at the end of each run. The inoculum, obtained from a pilot-scale biodesulfurization plant, was also analyzed. Results are shown in Table 1. The inoculum contains *Thioalkalivibrio* (relative abundance 32%) and *Alkalilimnicola* (18%). Both are haloalkaliphilic SOB. *Thioalkalivibrio sulfidiphilus* has previously been found to be the dominant SOB in both full-scale and lab scale biodesulfurization installations (Sorokin et al., 2012, 2008b). Genome analysis indicates that it oxidizes HS<sup>-</sup> to S<sup>0</sup> via flavocytochome *c* (Fcc) sulfide dehydrogenase and S<sup>0</sup> to SO4<sup>2-</sup> via the reversed DSR pathway (Muyzer et al., 2011). *Alkalilimnicola* species are facultatively autotrophic and facultatively anaerobic SOB (Hoeft et al., 2007; Sorokin et al., 2006). It has been suggested that some *Alkalilimnicola* members are limited to partial sulfide oxidation to S<sup>0</sup> (De Rink et al., 2019).

At the end of each experiment, the microbial community of both the planktonic and the biofilm biomass had undergone considerable changes. In both experiments, the biofilm was dominated by bacteria belonging to the genus of Desulfurivibrio, with a relative abundance of 37.0% in experiment 1% and 43.3% in experiment 2. Desulfurivibrio is also present in the planktonic biomass (21.6% and 11.1%), while the relative abundance of Desulfurivibrio in the inoculum was only 1.3%. The genus includes a single species: Desulfurivibrio alkaliphilus (Sorokin and Merkel, 2020). Until now, Desulfurivibrio alkaliphilus was only found in soda lakes (Sorokin et al., 2020, 2008a). It is an obligately anaerobic deltaproteobacterium originally described as sulfur-thiosulfate reducer. However, more recently it was found that Desulfurivibrio alkaliphilus is a chemolithoautotrophic sulfur disproportionator (Poser et al., 2013). This means it is using  $S^0$  as both electron acceptor and donor, dismutating it to 1 mol  $SO_4^{2-}$  and 3 mol HS<sup>-</sup>. Furthermore, it can also couple the oxidation of sulfide to sulfate to the reduction of nitrate/nitrite to ammonium (Müller et al., 2016; Thorup et al., 2017). In our experiments, the N-source was urea and NO322 and NO222 were not supplied. Therefore, Desulfurivibrio alkaliphilus could have been responsible for the formation of  $SO_4^{2-}$ , but it is not clear whether it was contributing to the observed bioelectrochemical HS<sup>-</sup> oxidation to elemental sulfur. Desulfurivibrio alkaliphilus is related to the long-distance electron transferring cable bacteria (Müller et al., 2016; Thorup et al., 2017). Furthermore, it possess genes for the expression of conductive pili, which are used for external electron transfer (Walker et al., 2018). The dominance of Desulfurivibrio alkaliphilus in the biofilm was also observed by Ni et al. (2019).

It must be noted that electricity was produced immediately from the start of the experiment, i.e. with only planktonic bacteria. Since the relative abundance of *Desulfurivibrio* in the inoculum was only 1.28%,

#### Table 1

Overview of the key bacterial species in the solution (planktonic biomass) and attached to the electrode (biofilm) as analyzed via 16 S rRNA amplicon sequencing. An extensive table of the present species is provided in the Supplementary Material B.

	Relative abundances					Key physiology	Occurrence
	Inoculum	End of exp 1 (planktonic)	End of exp 1 (biofilm)	End of exp 2 (planktonic)	End of exp 1 (biofilm)		
Sulfide oxidizing bact	teria (SOB)						
Alkalilimnicola	18.48%	18.58%	4.02%	6.36%	1.94%	facultatively autotrophic and facultatively denitrifying SOB	Soda lakes and biodesulfurization plants
Thioalkalispririla	0.22%	1.21%	0.12%	3.34%	0.23%	microaerophilic lithoautotrophic SOB	Soda lakes
Thioalkalivibrio	31.88%	2.54%	0.75%	16.40%	9.93%	Lithoautotrophic SOB	Soda lakes and Thiopaq reactors
Thioalkalimicrobium	3.00%	2.90%	0.34%	0.73%	0.99%	Lithoautotrophic SOB	Soda lakes and Thiopaq reactors
Rhodobacteraceae / other	28.75%	2.59%	0.42%	9.38%	2.43%	Might be photoheterotrophic anaerobic SOB	Might be in soda lakes
Arcobacter Anaerobic sulfur bact	0.06%	3.08%	0.08%	14.30%	0.86%	Microaerophilic sulfide oxidation	Marine
Desulfurivibrio	1.28%	21.59%	36.97%	11.11%	43.31%	Elemental sulfur disproportionation or sulfur reduction with H <sub>2</sub>	Biodesulfurization plants
Desulfurispirillum	0.35%	1.10%	6.03%	0.35%	1.38%	Sulfur reduction; dissimilatory ammonification	Soda lakes
Fermentative bacteria	a						
Acholeplasma	0.58%	6.10%	0.67%	3.44%	1.99%	Fermentative	Found in soda lake metagenomes
Anoxynatronum	0.83%	0.92%	1.30%	0.28%	2.73%	Fermentative	Soda lakes
Tindallia	0.40%	0.14%	1.89%	0.16%	1.51%	Acetogenic, can use H <sub>2</sub> and reduce thiosulfate	Soda lakes
Uncultured (cloacimonetes) Hydrolytic bacteria	0.00%	2.12%	2.84%	0.10%	0.40%	Fermentative	Found in soda lake metagenomes
Lentimicrobiaceae / uncultured	0.24%	15.69%	3.69%	16.13%	8.02%	Polysaccharidolytic	Soda lakes
Natronoflexus	0.00%	0.03%	7.31%	0.01%	0.45%	sugar fermentation	Soda lakes

*Desulfurivibrio* is not the only species that can exchange electrons with the electrode and that a biofilm is not required for the electricity production. This is in agreement with the experiments of ter Heijne et al. (2018), in which no biofilm was present.

At the end of run 1, the relative abundance of *Alkalilimnicola* in the solution was still 18%, while the relative abundance of *Thioalkalivibrio* decreased to 2.5%. In run 2, relative abundances of both *Alkalilimnicola* (6%) and *Thioalkalivibrio* (10%) in the solution had decreased.

## 3.7. Considerations

In this paper we demonstrated a continuous electrochemical process to simultaneously remove sulfide, produce elemental sulfur and recover energy. In this process, the removal of sulfide and release of electrons was spatially separated by applying the electron shuttling capacity of SOB. Our results show that continuous release of electrons by SOB took place in absence of dissolved sulfide at the anode. Furthermore, it was shown that biomass growth occurred.

However, there is still room for improvement, e.g. with respect to product formation. In both experiments, the main product was sulfate. Sulfate formation is unwanted, because it requires NaOH addition to prevent acidification of the process solution. In experiment 1, the overall selectivity for sulfur formation was 44% and in experiment 2, the selectivity for sulfur formation was 22%.

Even though the aim was to growth planktonic biomass, which can be recirculated between sulfide uptake reactor and the anode compartment, it was found that biomass was accumulating on the electrode, while the amount of planktonic biomass slightly decreased during the experiments. This suggests that most of the metabolic energy was obtained by the biofilm, even without dissolved sulfide in the solution to the anode compartment. Therefore, it is likely that the sulfate was formed by bacteria in the biofilm (e.g. *Desulfurivibrio*, which was the dominant species in the biofilm). Furthermore, at the first days of each experiment, while no biofilm was present yet, the selectivity for sulfur was much higher (see Figs. 2 and 3). Therefore it is hypothesized that the planktonic bacteria oxidize sulfide to sulfur and that the biofilm oxidize sulfur to sulfate. Hence, it remains a challenge to be able to control the product formation towards sulfur (without dissolved sulfide in solution). Preventing the formation of a biofilm might be an important step to obtain a higher selectivity for sulfur formation.

We have demonstrated that bacteria can take up sulfide in the sulfide uptake reactor and subsequently discharge electrons to the anode of an electrochemical cell without dissolved sulfide present in the electrochemical cell. This reduces the risk of electrode passivation resulting from sulfur deposition on the anode via electrochemical oxidation. This electrode passivation is problematic in (bio)electrochemical systems for direct sulfide removal. We have used XRD to determine the presence of elemental sulfur on the electrode and no elemental sulfur was detected. Future research should focus on the sulfur deposition to confirm that the electron shuttling approach prevents electrodeposition of elemental sulfur.

Another advantage is that, due to the application of haloalkaliphilic bacteria, the system can be operated at high salt concentrations (1 M Na<sup>+</sup>) and alkaline conditions (pH 8–9). This reduces ohmic resistances, resulting in a higher energy efficiency of the bioelectrochemical system compared to systems operated at neutrophilic conditions and low salt concentrations. Compared to conventional biodesulfurization processes, no energy is required for e.g. aeration of the bioreactor and energy from the oxidation of sulfide is recovered. This will improve the energy efficiency of biological gas desulfurization under haloalkaline conditions.

## CRediT authorship contribution statement

**Rieks de Rink** planned and performed experiments, analyzed the results and wrote the manuscript. **Micaela Lavender** and **Dandan Liu** supported experimental work and data analysis. **Jan Klok**, **Annemiekt**  ter Heijne and Cees Buisman supported study organization and data analysis and revised the manuscript. Dimitry Sorokin supported data analysis and revised the manuscript. All authors read and approved the manuscript.

## **Declaration of Competing Interest**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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# Appendix A. Supporting information

Supplementary data associated with this article can be found in the online version at doi:10.1016/j.jhazmat.2021.127358.

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