Update on the 2012-2020 trends in the St. Eustatius fisheries

Author(s): Morgane Amelot[1] Thomas Brunel[1], Dolfi Debrot[1], Kimani Kitson-Walters[2]

1. Wageningen Marine Research

2. Caribbean Netherlands Science Institute

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Contents

Summary

This report is an update of previously published reports presenting an overview of the trends in St. Eustatius fisheries based on the fisheries data collected on the island from 2012 to 2020.
The fisheries of St. Eustatius remain mostly conducted by small open boats with outboard engines. The number of fishing trips carried out by the fleet increased over 2014 peaking in 2015 with an average of 79 trips per months, and subsequently decreased in the following years to reach a minimum average of 25 trips per month in 2019. In 2020, the fishing effort per month increased to an average of 42 trips per month.

The main activity is a Caribbean spiny lobster (*Panuliris argus*) fishery using traps, also catching a mix of reef fish. This fishery is responsible for nearly 70% of the lobster landings on St. Eustatius. The trend in the annual landings in this fishery broadly follows the trends in the fishing effort, with landings reaching 30 tonnes in 2015 and since 2017 dropping to values comprising between 7 and 9 tonnes per year. Landings of lobsters from the trap fishery show a strong seasonality with higher landings during September-March, and low landings during June-July. The abundance index (derived by modelling the landings per trip) indicates an overall increase in lobster abundance from 2012 to 2020, with temporary declines in 2015 and 2018. The average carapace length (CL) shows interannual variations without any specific trend, but is on average 94.5 mm for females and 102 mm for males. This means that an average of 42% of the lobsters are landed below the legal size limit (95 mm). This problem is especially acute for females of which 55% of the landings are of sublegal size. Estimations of F/F_{MSY} proxies based on the length distribution over time suggest an overexploitation of this stock with values of F/F_{MSY} between 1.25 and 1.375 for females and 1.125 and 1.25 for males.

The species composition of the bycatch of reef fish in the lobster traps is very diverse, and is dominated by Acanthuridae (Blue Tang, Doctorfish, Surgeonfish), Ostraciidae (Honeycomb and Scrawled Cowfish), Serranidae (Coney and Red Hind) and Holocentridae (Squirrelfish). The trends in the reef fish bycatch in the lobster traps decreased from 2014 with values ranging from 9.9 tonnes in 2015 to 1.5 tonnes in 2019 and 1.6 tonnes in 2020. The biomass index calculated from the landings per trip suggests a decrease in the combined reef fish stock size from 2014 to 2020. Length frequency data for the main fish species caught in the lobster traps do not show any notable changes over the period studied. F/F_{MSY} proxies were estimated for the 7 most landed fish species.

The second most important fishing activities after trap fishing are scuba diving and free diving. Both activities catch spiny lobster and fish though they both are largely dominated by lobster catches. Scuba and free diving fleets composition and reporting varied considerably during the time period considered making it impossible to extract a year effect from a GLM approach. Consequently, this approach was not considered in this fishery. Scuba divers also conduct a fishery targeting Queen conch (*Lobatus gigas*), representing roughly 34% of the trips. Estimates of the annual conch landings are highly variable, and likely to be fairly uncertain due to the lack of information from logbooks in some years. The mean length of the conch landed appears to be stable over time, at 24.5 cm and 23.7 cm for females and males, respectively.

Next to the traps and diving fisheries, different line fisheries are conducted around St. Eustatius. A handline fishery on reef fish produced landings of between 1.7 and 4.3 tonnes of fish per year in the period 2014-2017, but with much lower estimates during 2018 and 2019, mainly due to a drop in effort. In 2020, landings increased to 1.2 tonnes. Large pelagic fish are also caught by trolling, with landings varying between 0.6 and 2.4 tonnes per year (2014 and 2016 respectively), 2020 landings were estimated at 1.8 tonnes.

Our main recommendations in terms of both management and research and monitoring are as follows:

- Improve control of and compliance with lobster size-limit regulations.

- Develop a FAD (Fish Aggregation Device) fishery management plan as part of a St. Eustatius fisheries development plan.
- Improve port sampling monitoring and subsampling intensity to cover at least one third of the trips dedicated to each fishing métier.
- Keep collecting data on reef fish species to estimate their status. This can be done best by combining more intensive port sampling with fisheries-independent studies on the distribution and abundance of these species around St. Eustatius.
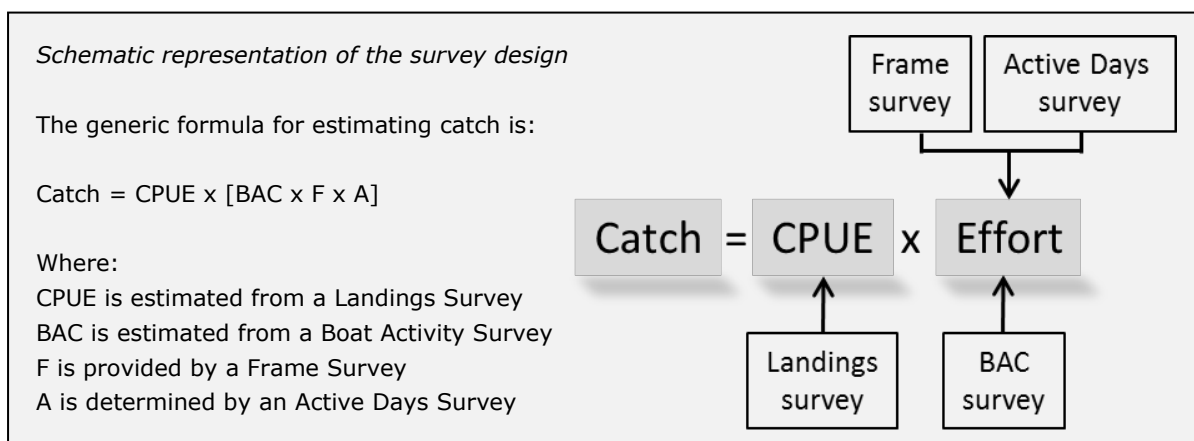
1 Introduction

A fleet of about 20 fishing boats operates on St. Eustatius (de Graaf et al. 2015), however, the number of boats that actively fish full-time is ≤5. All other vessels do so on a part-time basis. All boats are open boats ranging between 5 and 10 m in length and powered by single or twin outboard engines. The most commonly used fishing gear is the Antillean-style arrowhead trap, usually built from chicken wire around a wooden or steel frame. This fishery targets spiny lobster, and has a bycatch of mixed reef fish. In addition to trapping, both mixed reef fish (spear gun) and lobster (snare) are caught using the different methods of either free diving or scuba diving. The fleet also has different line fisheries, among which trolling for large pelagic fish is the main activity.

In 2012, WMR (then named IMARES) was commissioned by the Dutch Ministry of Economic Affairs to start working with local organizations on baseline marine resource studies and the development and implementation of robust, efficient and (internationally) standardized monitoring programs of coral reef health indicators in the Caribbean Netherlands. As part of this activity, a fisheries research programme was initiated on St Eustatius, and has since then been continued. Fisheries data collection is currently carried out on both Saba and St Eustatius by Data Monitoring Officers funded by the Dutch government. A previous WMR report (de Graaf et al. 2015) made an assessment of the status of the coral reefs in St. Eustatius, based on a number of health indicators (Fish biomass, Coral and algae cover, Coral recruitment and health, Water quality, Status of sharks, Status and trends in the fishery). As part of that assessment, a first analysis of the fisheries data collected on St. Eustatius was presented. A first update of these analysis was published in 2020 (Brunel et al. 2020). The present report gives a new update of these analyses on the St. Eustatius fisheries including 2 additional years of data, now based on data collected between 2012 and 2020. In addition to estimates of landings for the main species groups and effort for the main fishing gears, abundance indices are calculated based on the average landings per trip, to indicate the relative developments in abundance of the main species groups over the last 9 years. Finally, F/F_{MSY} proxies are estimated based on length distributions for lobster and the main reef fish species landed in order to assess the likelihood of these species experiencing overfishing.

2 Methods

2.1 Data collection

A sample-based fishery survey (Stamatopoulos, 2002) was implemented in 2012 to collect basic data on catch, effort, species composition and length frequency of the fisheries on St Eustatius. Rather than directly counting all catches, the total catch for each boat for each of the fishing methods was estimated by using data on the number of boats, the activity level of the boats and the average catches per boat per day.

Schematic representation of the survey design

The generic formula for estimating catch is:

Catch = CPUE x [BAC x F x A]

Where:
CPUE is estimated from a Landings Survey
BAC is estimated from a Boat Activity Survey
F is provided by a Frame Survey
A is determined by an Active Days Survey

Frame Survey: A frame survey is a census-based approach to collate a list of homeports and boat/gear categories which is used as the basis for the Active Days, Boat Activity and Landings surveys. This entails generating a list of all fishing vessels on the island and associated gear types used by that vessel The frame survey is conducted at the start of each year and is updated monthly throughout the year.

Active Days Survey: Active Day Surveys are conducted at the end of each month to determine the number of active fishing days for each stratum in the survey design (e.g. homeport, boat).

Boat Activity Survey: Boat Activity Surveys are conducted at homeports separately for each boat to determine how many boats are active on a given day. These surveys are conducted 7 days per week by contacting each vessel to ascertain their activity on each day based on the results of the Active Days Survey. Note that these surveys are not always successful and is dependent on the willingness of the vessel owner to participate.

Landings Survey: Landings Surveys consist of collecting fishing logbooks, filled in by part of the fleet, based on fishermen's participation, which provide data per trip on landings (in number or weight depending on the species categories), gear types used, number of each gear deployed, and other operational information, and, for a sub-sample of the trips, conducting biological sampling to estimate the species composition and the length frequency of the fish landed (See Annex 1 for Logbook forms used in data collection). The gear types used by fishing vessels are fish/lobster traps, scuba diving, free diving, handline, trolling, dropline and longline. The logbooks are recovered at the same time as the Boat Activity Surveys and are also affected by the willingness of the vessel owner to participate.

2.2   Effort and landings estimation

Estimates of Number of trips :
The number of trips carried out per month was calculated from the activity survey only. To take into account the fact that there was not a 100% coverage of the vessel activity (i.e. several days of the month with no observation on vessel activity), a first raising of the number of trips was operated as follows:

raised number of trips = number of trips observed / survey coverage rate

The survey coverage rate is typically around 40% for the period before 2018, and closer to 100% since 2018 (Figure 1).

Some inconsistencies between the logbooks and the activity survey were noticed, with occasionally boats having more trips registered in certain months in the logbooks database than estimated based on the activity survey. This can be explained caused by the assumption of the raising procedure that the number of trips is evenly spread throughout the month, and that this repartition is the same than those of the logbooks. Of course, this is not necessarily true and means that when raising the activity to cover the whole month, differences can be expected between "estimated" and "observed". In cases where more trips were actually registered in the logbooks than estimated based on the activity survey, the number of trips in the logbook database was used instead of the raised number of trips from the activity survey.

From May 2017 to April 2018, the activity survey was discontinued, and therefore, there was no direct observation of the number of trips conducted. For these months, the number of fishing trips was inferred from the number of fishing trips registered in the logbooks database. The average ratio of this number of trips from the logbooks and the estimated monthly number of trips per boat was calculated for each fishing method, and used to raise the number of trips based on the logbooks to the total monthly number of trips for this period without activity survey.
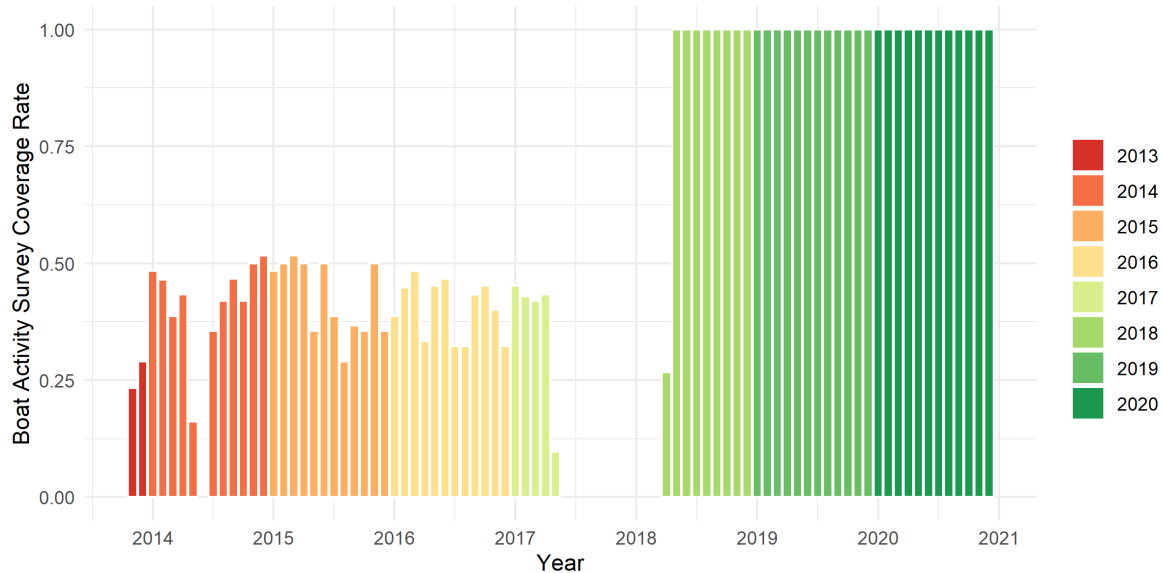

Figure 1. Rate of coverage of the boat activity survey (fraction of the days in a month when boat activity was recorded).

Number of trips per fishing method:

For the part of the fleet delivering logbooks, it was possible to split the effort between fishing methods (trapping, free diving, scuba diving, handlining, trolling). First, at the boat/month level, the proportion of each fishing method is calculated from the logbook and multiplied by the raised number of monthly trips (from the activity survey) for the corresponding boat to get an estimate of the number of trips per fishing method per boat. Summing over the boats having delivered logbooks, gives an estimate of the effort per fishing method for the part of the fleet delivering logbooks.

By comparing the active boats each month (activity) with the boats delivering logbooks, the % of logbook coverage can be estimated (Figure 2), and is used as a raising factor to estimate the effort per fishing method at the scale of the entire fleet. Again, for the period with no activity survey, the mean % of logbook coverage was used (grey bars on Figure 2). The 2014-2018 period was used as an 'historical period' to get the raising factor for 2017 and 2018 missing values.
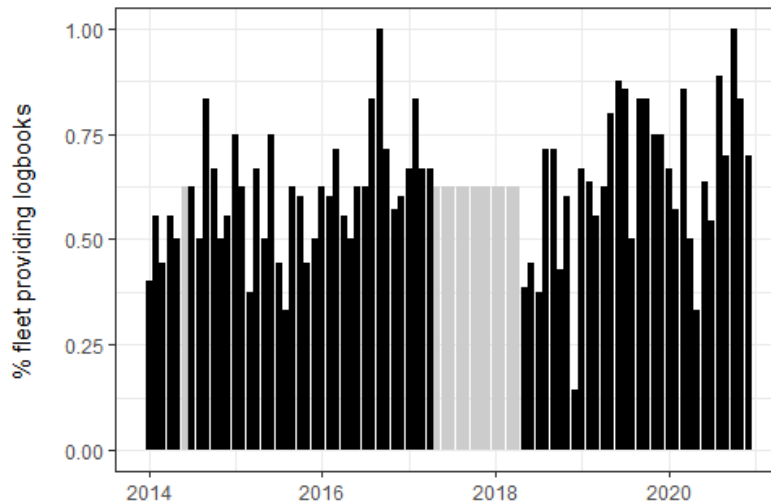

Figure 2. Proportion of the active boats providing logbooks each month (in grey: for months without activity log, the mean value is assumed).

Annual landings per species category

Landings per trip are reported by species categories. The categories considered in this report are the lobster, reef fish, pelagic fish and conch. Since not all vessels deliver logbooks, the calculation of the landings (i.e. multiply CPUE by effort) was not done on a per vessel basis, but using average CPUE across those boats that delivered logbooks in a given month and total effort per fishing method calculated as described above.

Conch is a relatively limited fisheries, and after consultation of local experts, it appeared that all the vessels involved in this fisheries (between 1 and 5 depending on the years) were actually delivering logbooks. For this reason, the total effort and catches for the conch fishery were only based on the data of those boats providing logbooks, and no raising to the total fleet was done.

Figure 3 summarizes the successive steps and raising procedures to estimate monthly landings and effort.
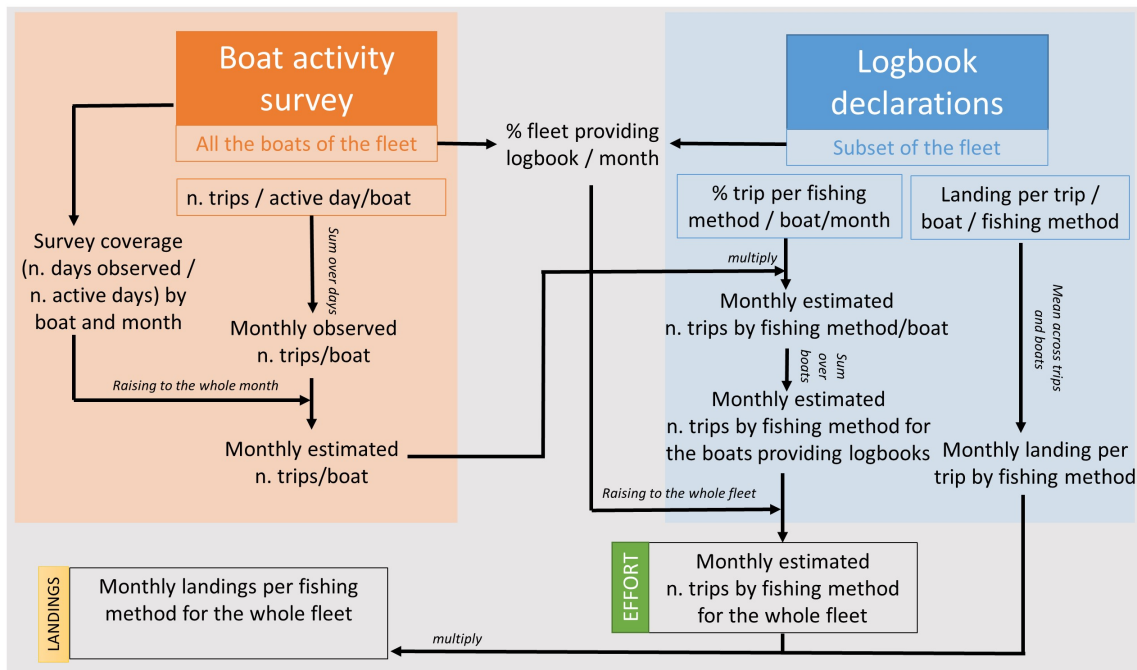
Figure 3. Estimation method for the monthly landings per species category and "métier" (fishing method) and monthly effort by métier.

2.3 Landings, species and length composition

For a subset of the trips, sampling of the landings was conducted both for species and length composition. For some fishing methods (free diving and handlines), the number of samples was very low. For the length measurements, there were only few samples of fish available for the trolling and the free diving activities, and only few samples of lobsters from the scuba diving fishery. For almost half of the lobster length trip Id were not reported in the database which made the attribution of those length to a particular gear impossible, this record were identified as NA gear (NA gear in Table 1).

Table 1. Number of samples available for the period 2011-2020 per fishing method for the length composition of the landings. FD : free diving, HL : handline, PT : lobster traps, SD : scuba diving, TR : trolling, NA: unknown gear.

Species composition	FD	HL	PT	SD	TR	NA	Total
Number of trips	3	13	155	72	148		391
Number of fish	31	398	5 257	1 148	993		7 827
Length measurements	**FD**	**HL**	**PT**	**SD**	**TR**		**Total**
Number of fish	39	776	14 318	952	226		16 311
Number of lobsters	46	-	1742	304	7	1537	3 636
Number of conch				3 166			3 166

Lobster vs. fish

Some methods target both lobster and a mix of reef fish. The proportion of these two species categories was estimated based on the landings per trip as declared in the logbooks. Lobster landings are declared in numbers while fish landings are declared in weight. Available lobster length measurements were used to compute the average size of the lobsters in the landings, which was subsequently converted to average individual lobster weight. Landings per trip in numbers were then converted into landings in weight using the average lobster weight.

Proportion of fish species

Landings species composition data was available for a subset of the trips for which port sampling was conducted (see Table 1). This data was used to analyze the yearly species composition of the landings for the different fishing methods.

2.4   Estimation of abundance indices

In order to get annual abundance indices for the main species groups, landings and effort data from the logbooks were analysed for each of the main species groups for each of the fishing methods. Part of the variations in the annual mean landings per trip reflects other factors than changes in abundance, such as changes in overall effort, monthly or spatial repartition of the effort or even different contributions of different vessels to the annual effort. All these factors were considered in the modelling approach that was apply to obtain abundance indices. In order to standardise the landings per trip and extract an annual abundance index, the landings per trip where modelled for each of the main species groups and for each of the fishing methods using a GLM with a negative binomial distribution. The formulation of the full model is as follows:

Log(landings per trip)   =   intercept
                                            + Year effect
                                            + Month effect
                                            + Boat effect
                                            + soaking time
                                            + log(trap number)

Model selection was done by fitting the full model, fitting sub-models dropping each of the explanatory variables one by one, and then conducting a likelihood ratio test to assess whether each explanatory variables from the full model is significant. In case not all model terms are significant, the explanatory variables with the highest p-value is then removed from the model. This procedure is repeated until all remaining explanatory variables have a significant effect.

The model formulation above corresponds to the model applied to the lobster landings, in which the response variable is the number of lobsters caught in one trip, and the explanatory variables representing the fishing effort deployed during the trip are soaking time (number of days of immersion of the traps that were lifted during the trip) and number of traps. For other species, different units may be use for the landings (e.g. kg), and for different fishing methods, different variables may represent fishing effort (e.g. number of divers, number of fishing lines …). In cases where the response variable was in kilograms, a GLM with a log-normal distribution was used.

In this formulation of the model, one parameter is estimated for the intercept and one for each of the levels of the different effects (year, month, boat and area). One parameter is also estimated for the linear regression of log landings (number) against the log of the fishing effort (in trap numbers, hours of diving). The chosen model formulation implies a power function between the landings and the effort, which is the formulation typically used for ad hoc standardisation of catch rates in trap fisheries. The parameter estimated for this variable corresponds to the exponent of the power function, which, in case saturation indeed occur, is expected to be smaller than 1. The year effects estimated by this method correspond to the variations in the landings per trip which are explained by the year, when all other

sources of variation have been taken into account (including the changes in effort). These year effects can therefore be interpreted as abundance indices.

For the lobster, the GLM used a negative binomial error distribution, as the data are counts and the fit of a GLM with a Poisson distribution indicated overdispersion. As the response variable is landings in numbers, the estimated year effect are referred to as abundance index. For the GLM on fish species, the landings are reported in weight, and the model uses normally distributed errors. The index is referred to as a biomass index.

GLM approaches were only used when a set of 3 or more boats covered more than half of the time period (2012-2020). No boat were consistently present and fishing using scuba or free diving. The changes in boat composition of the scuba and free diving fleet from 2012 to 2020 made it impossible to differentiate year effects due to change in fleet, from year effects due to changes in stock abundances. Consequently, this approach was unsuitable to obtain standardised abundance indices for this fleet.

2.5 Length composition analyses and F_{MSY} proxy

The length frequency distribution was inspected for the main species and their interannual variations were quantified. A number of length-based approaches have been developed to provide indicators of the state of the stock from landings length-frequency data. The general underlying idea is that applying a size selective fishing mortality to a population will result in a shrinkage of the population length composition towards smaller sizes. Measuring the extent to which the length composition has shrunk towards smaller sizes (compared to an unfished state) can provide an indication of the level of the fishing pressure.

Here, the length-based indicator for the ratio of F/F_{MSY}, part of the ICES Data Limited Stock methods (ICES, 2018), has been applied to length frequency data for the lobster and the main reef fish species. This approach requires the estimation of the length at recruitment in the fishery, L_c, defined as the length for which the frequency (number of fish measured for this length bin) corresponds to 50% of the frequency of the mode of the distribution (fig 2.4.1). Then, assuming that 1) the stock is at equilibrium (constant total mortality and constant recruitment) and 2) fishery selectivity follows a logistic curve, the theoretical mean length of landings larger than L_c when the stock is exploited with a fishing mortality, F, equal to natural mortality, M, can be approximated by :

$$L_{F=M} = 0.75\, L_c + 0.25\, L_{inf}$$

where L_{inf} is the asymptotic length from the von Bertalanffy growth equation. This length $L_{F=M}$ is used as a length-based proxy for F_{MSY}. Computing L_{mean}, the observed mean length of the landings (individuals larger than L_c) and comparing L_{mean} with $L_{F=M}$ gives an indication of the state of the stock with respect to MSY. In the example of fig 2.4.1, the value of L_{mean} is very close to the value of $L_{F=M}$, which correspond to a fishing mortality close to F_{MSY}. In practice, the ratio $L_{F=M}/L_{mean}$ can be used as a proxy for the exploitation status F/F_{msy}, with values above 1 indicating overfishing and values below 1 sustainable fishing mortality.

The values of the parameter L_{inf} are presented in table 2.4.1 and were taken from Leocádio and Cruz (2008) for lobster and from literature on reef fish (Annex 2). The reef fish L_{inf} values did not come from the same area and might be considered carefully.

Table 2.4.1 : L_{inf} **values used for the calculation of the length based indicators.**

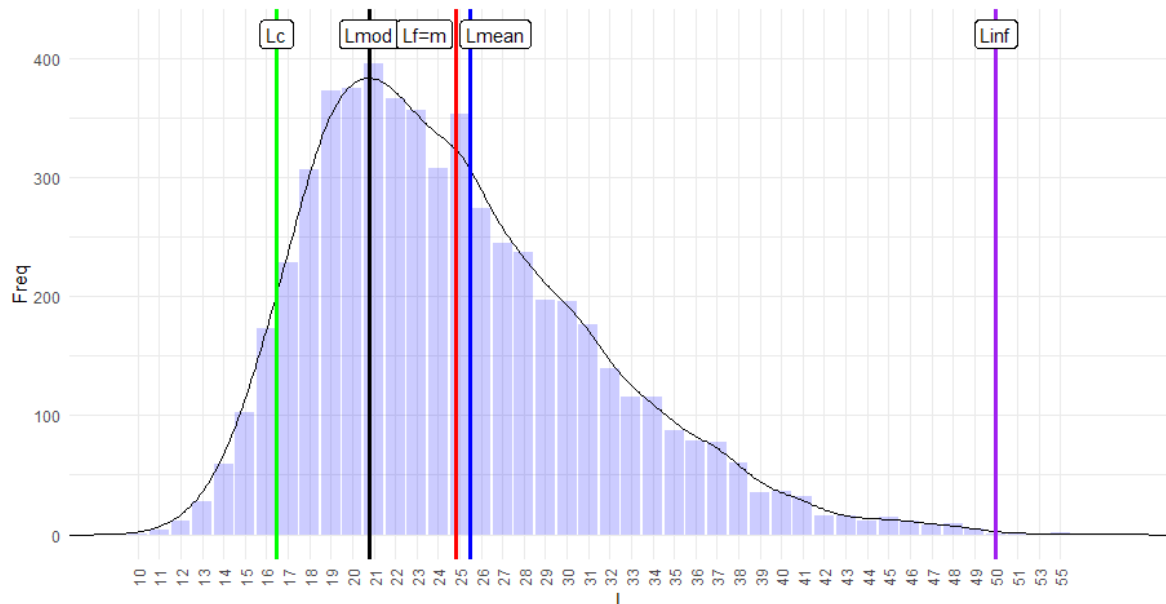| Species	L_{inf}
Spiny lobster male	239mm
Spiny lobster female	245mm
Queen triggerfish	45cm
Red hind	39cm
Blue tang	28cm
Doctorfish	27cm
Squirrelfish	27cm
Coney	38cm
Honeycomb cowfish	50cm

Fig 2.4.1. : illustration of the ICES length-based indicator method applied to simulated length frequency distribution data. L_{mod} : modal length, L_c : length at recruitment to the fishery (length at 50% of modal value), L_{inf} : asymptotic length from the von Bertalanffy growth equation, $L_{f=m}$: mean length of individuals above L_c when the population is exploited at F_{MSY}, L_{mean} : mean length of the landings of individuals above L_c.

3   Results

3.1   Fishing activity and landings

3.1.1   Effort

The monthly effort (number of trips) of the fleet was estimated based on the data from the boat activity survey (and when the survey was not carried out, extrapolated from the logbook data). The monthly estimated number of trips was very variable with no unidirectional trend over the period 2014-2020. Between the last months of 2014 and mid-2016 the number of trips was constantly higher than 50 per month (Figure 4). For the period before October 2014 and after June 2016, the monthly effort was lower, at around 25 trips, with even lower values in the first half of 2017. The effort then increases between July 2017 and March 2018, with estimates up to 250 trips per month. However, this period does correspond to months without activity survey and for which the effort is inferred from the logbook data and should therefore be considered with caution. After that, effort was again at a lower level (around 30 monthly trips), with a slight increase after mid-2020 (42 to 80 trips per month). The proportion of the trips for which no logbook data was available (white part of the bars on Figure 4) is variable, with months of good logbook coverage (e.g. August/September 2016, April/December 2019, July/December 2020) to months with a low coverage (August 2014, December 2018).
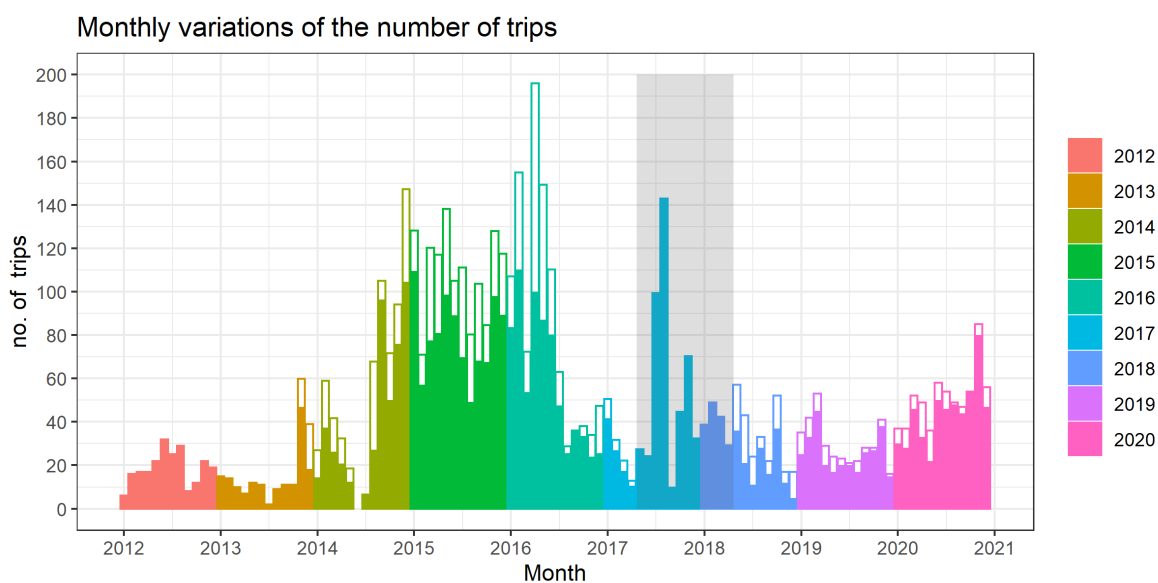

Figure 4. Estimated number of trips per month. The white part of the bars represents the number of trips registered in the activity survey corresponding to boats not providing any logbook data. Note that there is no data available for June and July 2014, nor for the period May 2017 to May 2018 (grey area). For these periods, the number of trips is inferred from the logbook information only (no activity survey, see section 2.2).

3.1.2   Fishing methods

Based on the logbooks that were available, it was possible to estimate the relative importance of the different métiers in the activity of the fleet. The importance of each métier was quite variable from one month to the next, and there was no clear temporal trend in the data (Figure 5). The lobster traps appear to be consistently the most-used gear, representing overall 40% of the trips. Free diving was

the second most important fishing practice (26%), followed by handlines and scuba diving (22% each). Other fishing practices represented smaller percentages of the activity (18% for trolling).
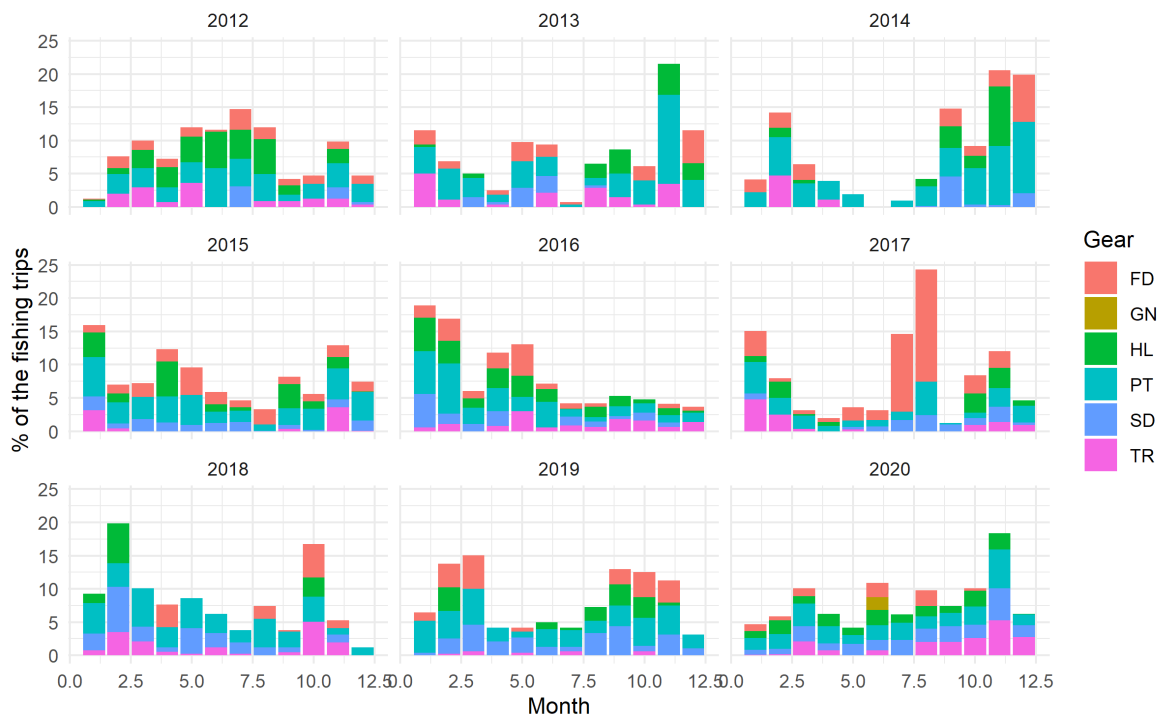

Figure 5. Monthly proportion of the fishing trips for each métier. FD: free diving, HL: handline, PT: lobster traps, SD: scuba diving, TR: trolling.

3.1.3 Landings per species category

The main species of interest is the Caribbean Spiny Lobster, *Panulirus argus*, which is caught mainly using lobster traps (72.5%) and in free and scuba diving (17.5 and 10% respectively). Monthly estimates of the number of trips, catch rates and landings of lobster per métier are given in annex 3, and the yearly total landing values are given in Table 2. The estimated annual landings of lobster were the highest in 2015 (47.5 tonnes) and were relatively low in 2019 (around 9.5 tonnes). This trend is quite similar to the variations observed in the number of trips (which were the highest in 2015) and shows no correlation with the changes in the average landings per trip, which were not higher in 2015 than in other years (annex 3).

Similar to the lobster, most of the reef fish landings were taken by the trap fishery (43%), with also a significant contribution from free diving (21%) and handlines (25%) and a smaller share from scuba diving (10%). Annual landings were higher during the period 2014-2016 (between 10 and 17 tonnes annually) and decreased to lower levels (3 to 7 tonnes) between 2017 and 2020. Again, this decrease seems to be more related to the changes in fishing effort than to any change in the catch rates of the different métiers (annex 3).

Landings of pelagic fish caught by trolling were on a lower scale, ranging annually between 0.6 tonnes in 2014 to 2.4 tonnes in 2016. For this category as well, there were no marked differences between years in the catch rates (annex 3), and the differences in annual landings values seem to be more related to the variations in the number of trips carried out.

Finally, estimates of landings of conch in the scuba diving fishery were highly variable, ranging from 1420 individuals in 2014 to 8257 in 2020. The particularly low value in 2014 is explained by the fact that landings of conch were only registered in the logbook data for 4 months of the year, and it is unknown whether it was because there were no trips targeting conch during the rest of the year, or because no logbooks from conch trips were available.

For all species categories, the estimated landings for 2017 and 2018 were based on monthly effort inferred from logbook data but only for the months without activity survey. This means that the estimates for these two years have additional uncertainty compared to other years, due to the crude assumption made that the number of trips for which logbooks were provided is proportional to the total number of trips carried out.

Table 2. Estimation of the annual landings per métier in tonnes per species category (PT = lobster traps, FD = free diving, SD = scuba diving, HL = handlines, TR = trolling). All gears include PT, SD, FD, HL and TR.

	Lobster				reef fish					pelagic fish	All gears	conch
	(tonnes)				(tonnes)					(tonnes)	(tonnes)	(number)
Year	PT	FD	SD	total	FD	HL	PT	SD	total	TR	total	SD
2014	26.4	2.8	0.9	30.1	1.1	4.3	6.3	0.1	11.8	0.6	42.5	1420
2015	30.7	12.4	4.4	47.5	5.2	1.7	9.9	0.6	17.4	0.7	65.6	7750
2016	14.1	4.6	2.5	21.2	2.6	4.1	3.1	0.5	10.3	2.4	33.9	6016
2017	8.8	5.8	1.5	16.1	1.8	2.4	1.6	0.7	6.5	0.9	23.5	5315
2018	8.8	1.7	1.1	11.6	1.6	0.3	2.6	0.4	4.9	1.7	18.2	3043
2019	7.3	1.0	1.3	9.6	0.5	0.5	1.5	0.7	3.2	0.2	13	5916
2020	9.2	0.5	1.7	11.4	0.4	1.2	1.6	0.8	4	1.8	17.2	8257

3.2 Trap fishery

3.2.1 Lobster landings

Proportion of lobster in the landings by weight

Reef fish and lobsters each represent roughly half of the landings by weight in the trap fishery (Figure 6). These proportions are relatively stable across the years, except for 2014 and 2016 when lobster landings corresponded only to 25% of the landings.
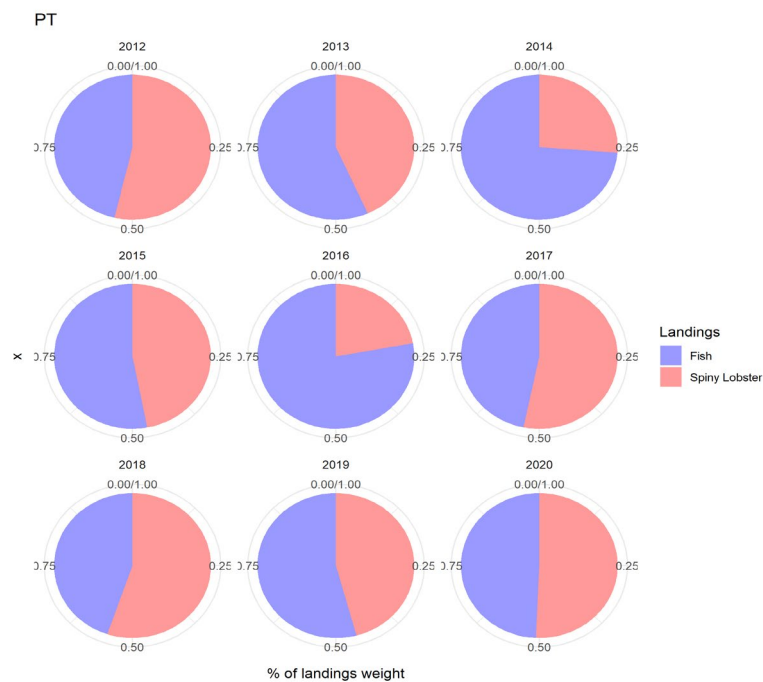

Figure 6. annual proportions of lobster and reef fish in the landings in weight of the trap fishery over the period 2012-2020.

LPUE standardisation using a GLM model

The unstandardized lobster LPUE (lobster landings per trip) of the trap fishery are shown in the annex 3. There is a clear seasonality, with higher landings per trip for the winter months than for the summer months. The GLM model was fitted to estimate the influence of different factors on the LPUE, and to extract an abundance index (year effect). This analysis was based on a subset of four boats selected on the basis of their presence during more than half of the time period (Table 3), and which were consistently active throughout the period studied.

Table 3. Number of years of the selected boats that were observed fishing from 2012 to 2020.

Boat	1	2	6	12
Coverage	9/9	8/9	6/9	5/9

All the factors included in the model were found to have a significant effect (Table 4). The landings per trip were positively related to both the number of traps set during the trip and the soaking time. There were also clear differences in fishing effectiveness between the vessels used for the analyses, with a ratio of around 1:2 between the least and the most effective boats. The estimated month effect (Figure 7) clearly illustrate the seasonality in the catchability of the traps, with lower values from April to August and higher values the rest of the year. Finally, the estimated year effects indicated an increase from low abundances in 2012-2013 to intermediate levels in 2014-2016, a further increase to a higher level in 2017, decreasing in 2018 to similar values as in 2014-2016 and increasing again 2019 and 2020, back to the 2017 level (Figure 7).

Table 4. Significance of each model term tested by removing them stepwise and comparing to the full model (GLM model of lobster catches per trip). AIC stands for Akaike information criterion (the lower the better). * : significant at the 5% risk level, ** : significant at the 1% risk level.

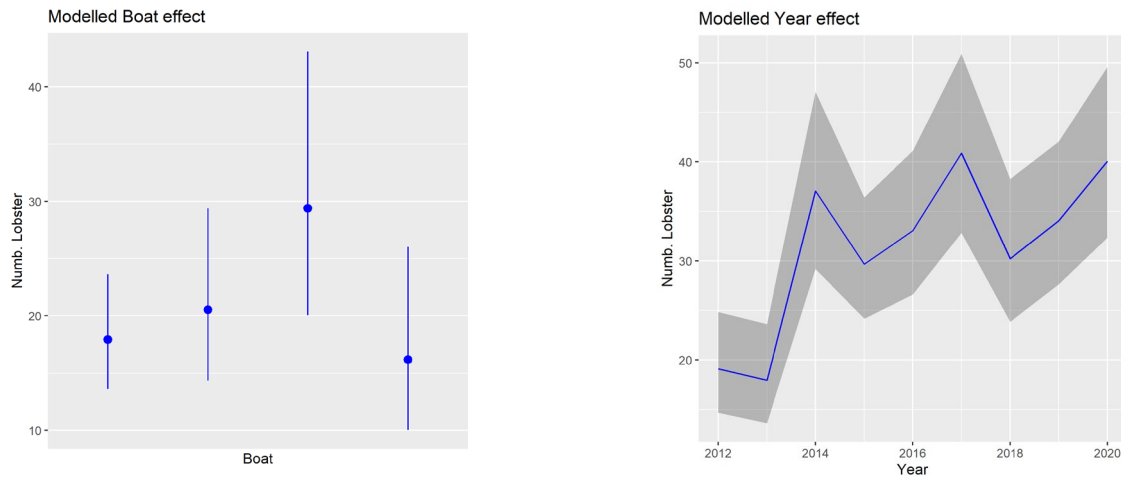| Model term removed	Number of parameters	AIC	Log-Likelihood ratio	p.value	Signif. level
<none>		5292.5			
factor(Year)	8	5348.9	72.460	1.588e-12	***
factor(Month)	11	5402.8	132.312	< 2.2e-16	***
factor(Boat_name)	3	5302.2	15.727	0.00129	**
Soaking_time_days	1	5341.8	51.329	7.811e-13	***
logTraps	1	5536.9	246.416	< 2.2e-16	***
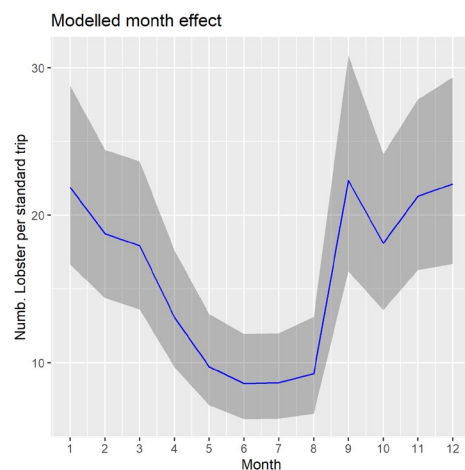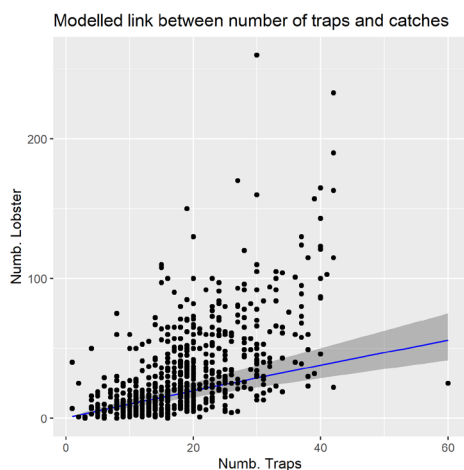
Figure 7. Modelled trap number, month, boat and year effects on the landings of lobsters per trips (in numbers). Blue lines represent the modelled effect (and associated 95% uncertainty in grey).

Length frequency distribution of the landings

Length measurements of lobster were available since 2011 (Table 5). The number of individuals measured varied from more than 1,000 in 2012 to less than 100 in 2016.

The mean length of the females landed is lower than for the males, and for both sexes length displayed similar variations in time. Annual mean length is comprised between 91 and 97mm for females and 97 and 107mm for males (Table 5). Annual median length is comprised between 90 and 97mm for females and 97 and 104mm for males. The higher values were observed mainly for males in 2017 and 2018. The sex ratio indicated a predominance of the males in the landings until 2018, and a predominance of females for the last 2 years. The proportion of landed individuals below the minimum size (95 mm) was in the range of 34-49%. Females were consistently caught at a smaller size and a higher percentage below legal size than males (55% below legal size, versus 32% below legal for males). Sub-legal size capture more greatly affects females than males.

Table 5. Annual mean length, sex ratio and proportion of undersized individuals in the lobster landings (samples mainly from the trap fishery).

| Year | Mean length (mm) | | | Sex ratio (males per female) | Proportion <95mm | | | Number lobsters measured |
	Females	**Males**	**Combined**		**Females**	**Males**	**Combined**	
2011	95	102	98	1.23	55%	42%	48%	165
2012	93	103	98	1.04	60%	22%	41%	1055
2013	96	103	100	1.06	41%	30%	36%	200
2014	95	97	96	1.44	50%	40%	44%	122
2015	95	101	98	1.43	55%	31%	41%	216
2016	91	101	97	1.87	83%	30%	48%	66
2017	97	107	102	1.09	41%	28%	34%	340
2018	96	106	101	1.11	44%	25%	34%	660
2019	92	98	95	0.79	59%	36%	49%	375
2020	94	101	98	0.95	57%	30%	44%	434
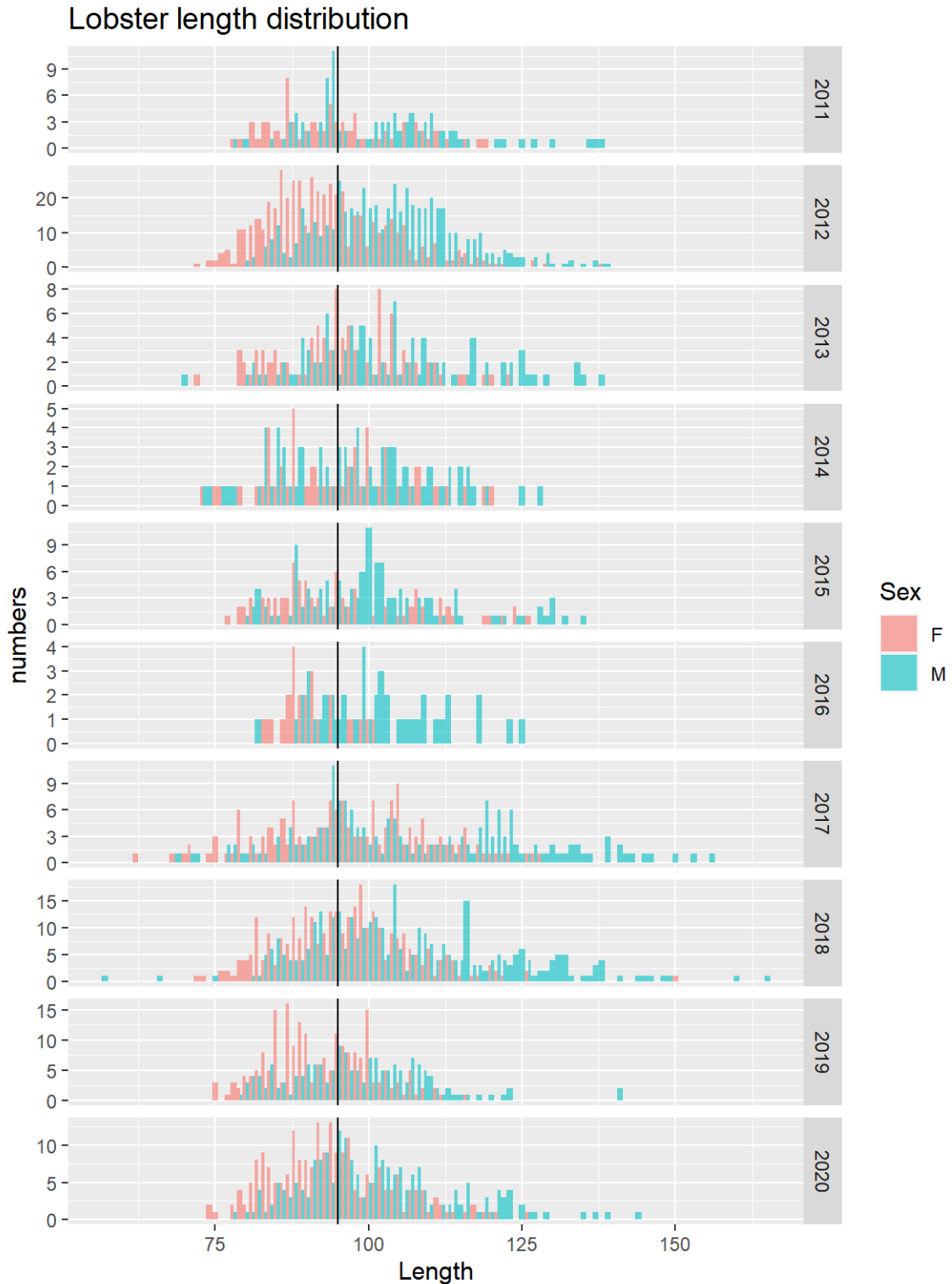
Figure 8. Yearly length frequency distribution of lobster by sex, black vertical line is the minimum landing length (95mm)

F$_{MSY}$ proxies

The length at recruitment to the fishery (necessary to compute the length based indicator for F/F_{MSY}) was estimated based on the landings length-composition for all years combined and for males and females, at L_c= 81 mm and L_c=87 respectively. Using the values for L_{inf} of 239 mm and 245 mm for males and females respectively, the corresponding values of $L_{F=M}$ were of 120.5 mm and 126.5 mm. The F/F_{MSY} proxy values indicate that the stock has been overexploited over the whole study period. In all years, the F/F$_{MSY}$ proxy indicates that females were more heavily exploited than males (since females were of smaller size in the landings, and have a higher $L_{F=M}$ value than males due to their larger L_{inf}).
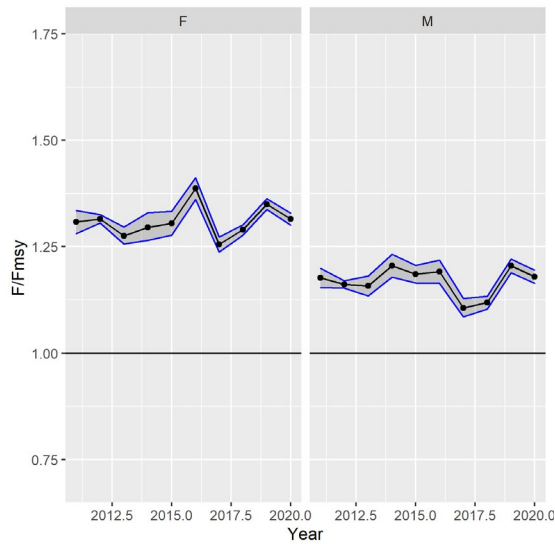
Figure 9. Length based indicator for F/F$_{MSY}$ calculated per sex on the annual length frequency data, blue lines and shaded area represent the confidence intervals obtained by bootstraps

3.2.2   Mixed fish landings

Species composition

The species composition of the reef fish landings is very diverse, as shown by the large proportion of the species representing less than 5% of the landings ("other"). Acanthuridae (Blue tang, *Acanthurus coeruleus*, Doctorfish, *A. chirurgus*, Surgeonfish, *A. oceanus*), Ostraciidae (Honeycomb cowfish *Acanthostracion polygonius* and Scrawled filefish, *Aluterus scriptus*) and Serranidae (Coney, *Cephalopholis fulva* and Red hind, *Epinephelus guttatus*) and Holocentridae (Squirrelfish, *Holocentrus adscensionis*) are the most abundant in the landings, representing on average, 40% , 20% and 23% and 11% of the landings, respectively (Figure 10).
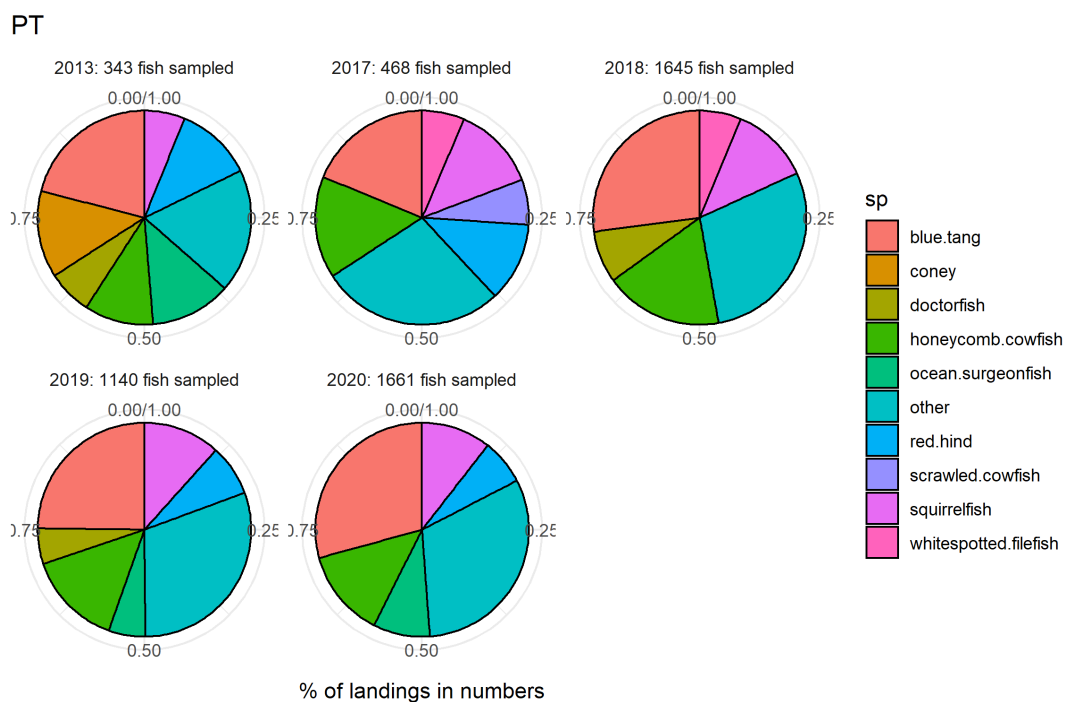

Figure 10. Species composition in number of the reef fish landed from the lobster trap fishery.

LPUE standardisation using a GLM model

The low LPUE for the reef fish in the trap fishery is shown in annex 3. Unlike for the lobster, no clear seasonal pattern was visible. The GLM model was fitted to estimate the influence of different factors on the LPUE, and to extract an biomass index. For reef fish, there was no significant effect of the number of traps set, but the effect of soaking time was significant (decreasing number of fish caught with increasing soaking time, Table 6). The month effect was also significant, but mainly due to the lower CPUE for the month of September compared to the rest of the year (Figure 11). The year effect indicates an increase in the biomass in 2013-2014, followed by an overall declining trend in biomass of reef fish.

Table 6. Significance of each model term tested by removing them step-wise and comparing to the full model (GLM model of reef fish landings per trip). See also Table 3.

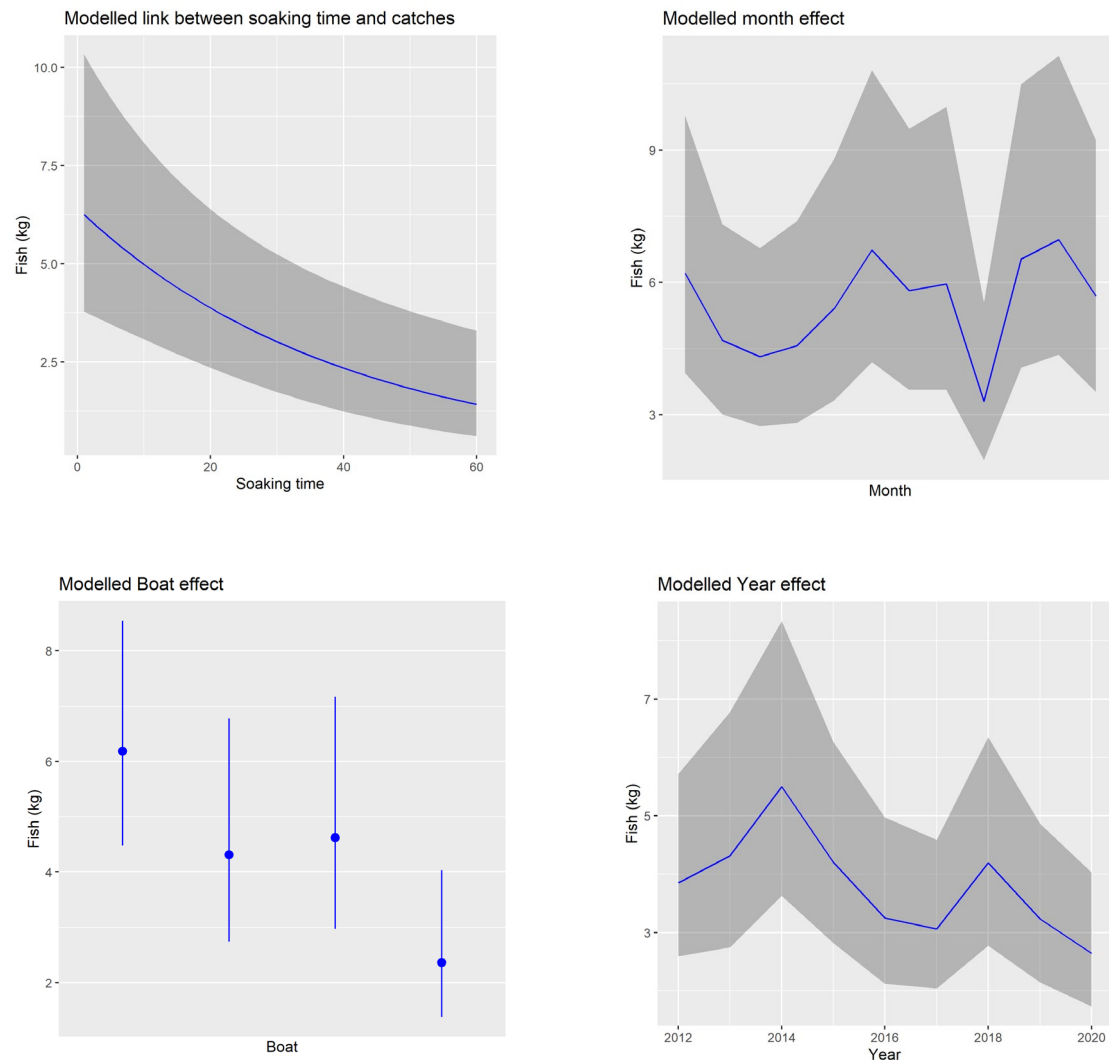| Model term removed	Number of parameters	AIC	F value	p.value	Signif. level
<none>		1509.7			
factor(Year)	8	1536.6	5.3395	1.699e-06	***
factor(Month)	11	1523.4	3.2152	0.0002858	***
factor(Boat_name)	3	1529.0	8.2943	2.058e-05	***
Soaking_time_days	1	1520.9	12.8215	0.0003703	***

Figure 11. Modelled effect of soaking time, month, boat and year on the landings of reef fish from the trap fishery per trips (in kg). Blue lines represent the modelled effect (and associated 95% uncertainty in grey).

Length-frequency distribution of the landings

The length-frequency composition of the landings of the reef fish caught in the traps are given in annex 4 for the main species. Although some length measurements are taken from the landings of other gears, the majority of the samples were from the trap fishery. Data for the other gears are not displayed in this report as they were too scarce to give an accurate representation of landings length composition per species.

For most species the mean length was stable over time. The Stoplight parrotfish (*Sparisoma viride*) is an exception, with a notably lower mean length in 2012 than for the rest of the period.

F$_{MSY}$ proxies

All fish are sold at the same price per pound. The seven species presented are the most landed ones. All F/F$_{MSY}$ proxies indicated stocks exposed to a fishing mortality closed to F$_{MSY}$ (Figure 12).
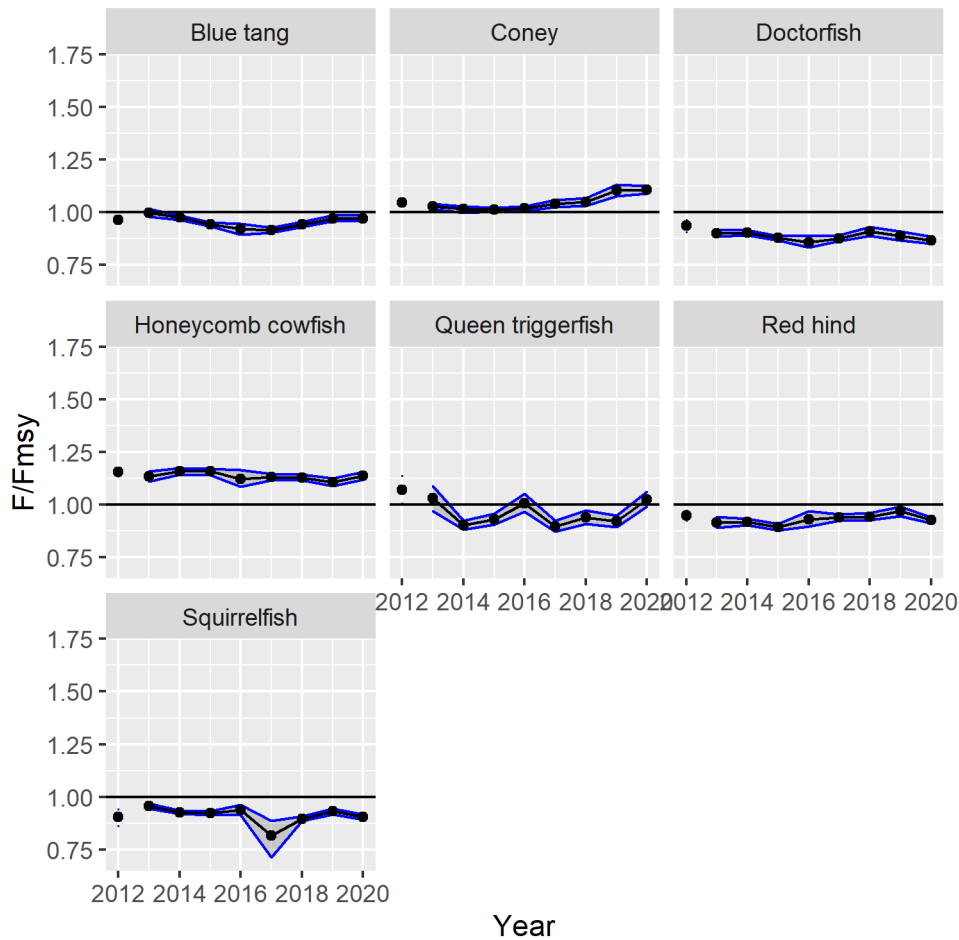

Figure 12. length-based indicator for F/F$_{MSY}$ calculated per species for the 7 most landed mixed reef fish species on the annual length frequency data.

3.3 Diving

3.3.1 Fishing activity

The activity of boats conducting scuba diving trips is divided between trips targeting lobster and reef fish, and trips targeting conch. Conch represents the totality of the landings for 34% of the trips (flat line right part of Figure 13), while it is absent from the landings for 47% of the trips (flat line left part of Figure 13). For the remaining 21% of the trips with a mix of conch, lobster and reef fish(increasing curve middle of Figure 13), the proportion of conch in the landings is generally high. The proportion of trips with conch only increased during the last two years, 2019-2020.
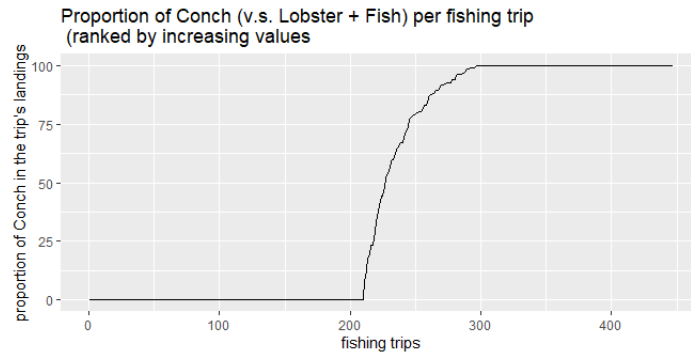
Figure 13. Distribution of the proportion of conch in the landings of scuba diving trips.

3.3.2 Mixed Lobster / reef fish fishery

Species composition

Lobster represents the vast majority of the landings for free diving (Figure 14), while lobster and reef fish contribute almost equal shares in the landings for the scuba diving fishery (for the trips not targeting conch).
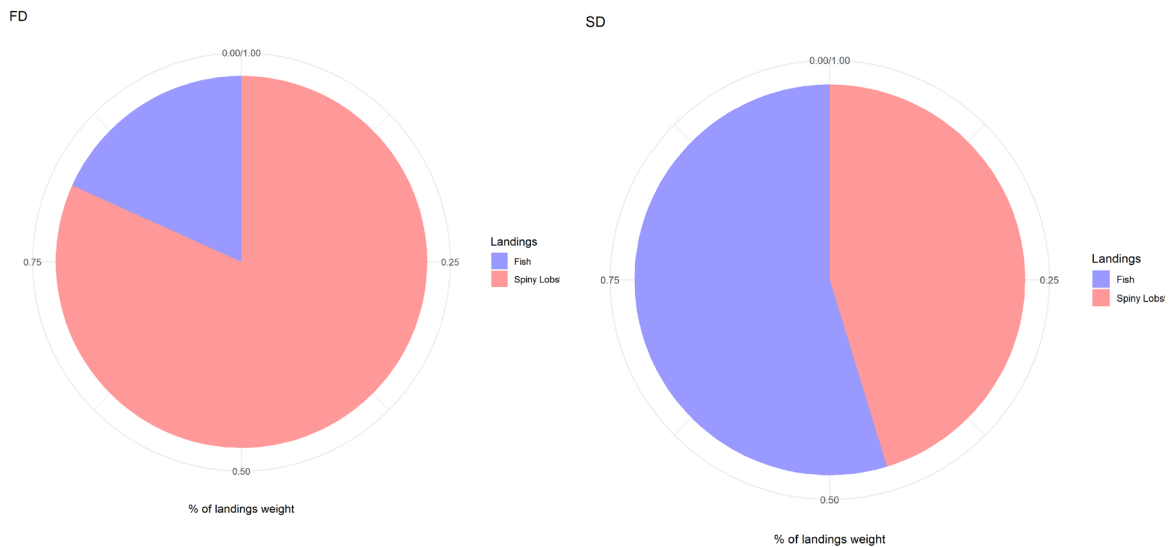

Figure 14. Average proportion of lobster and reef fish in the landings in weight of the free diving (FD) and scuba diving (SD) fishery over the period 2012-2020.

Landings from scuba diving were well-sampled in 2017, 2018 and 2020. The main species caught were Serranidae (Coney and Red hind), representing nearly half of the landings (Figure 15). The fishery also lands a fair proportion of Lionfish, *Pterois volitans*, and of different parrotfish (Scaridae). The category "other" (species that represent less than 5%) is also large, indicating that a large number of species has a small contribution to the total landings. Very little information was collected on the composition of the reef fish landings from free diving, and these are not presented here.
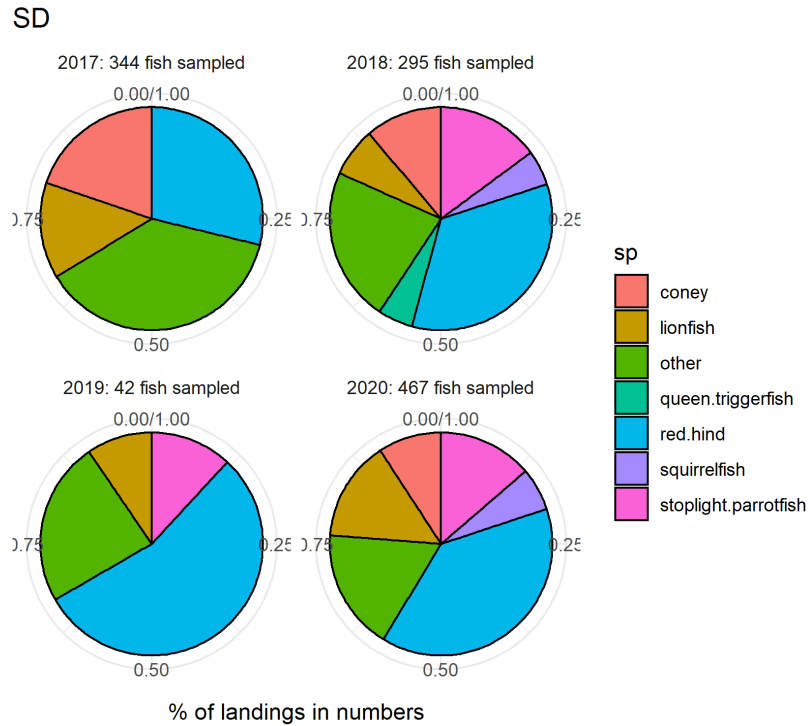

Figure 15. Landings species composition (based on fish numbers) for the scuba diving fishery.

Landings and effort data

From 2012 to 2020 the number of boats involved in diving fisheries varied from 2 to 8. However, no boat used this method during half or more of the time period. The lack of consistency in the boat coverage did not give the opportunity to differentiate year and boat effects. Consequently, it was impossible to get a standardization to track the annual trends in LPUE. For this reason the LPUE discussed in this paragraph are raw LPUE only.

For spiny lobster two trends in the raw data were observed; a general decrease in number of diving hours per trip over time and an increase in mean number of lobster caught per diving hour. Finally, the mean number of spiny lobsters caught per trip remains stable over time (Figure 16).
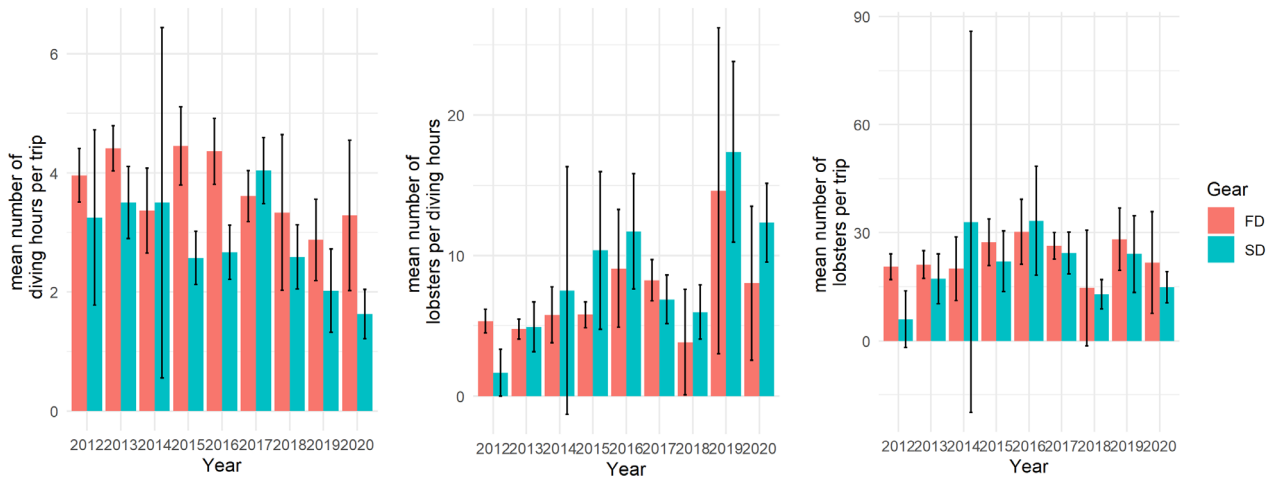
Figure 16. Raw numbers of diving hours per trip, in trips with spiny lobster catches (left), mean raw lobster numbers caught per diving hour (middle) and mean raw lobster numbers caught per trip (right) in Free and Scuba diving fisheries, 2012 to 2020.

Similar trends are observed in diving hours per trip for fish catches, however, the trend in fish weight caught per diving hour is slightly stronger for mixed reef fish (Figure 17). The mean weight of fish caught per trip also increased from 2012 to 2020 with particularly high values for free diving in 2020. During this year some fishermen went spearfishing and free diving during the same trips but reported their trip as free diving trips. It caused an overestimation of the number of fish caught while free diving during this year.
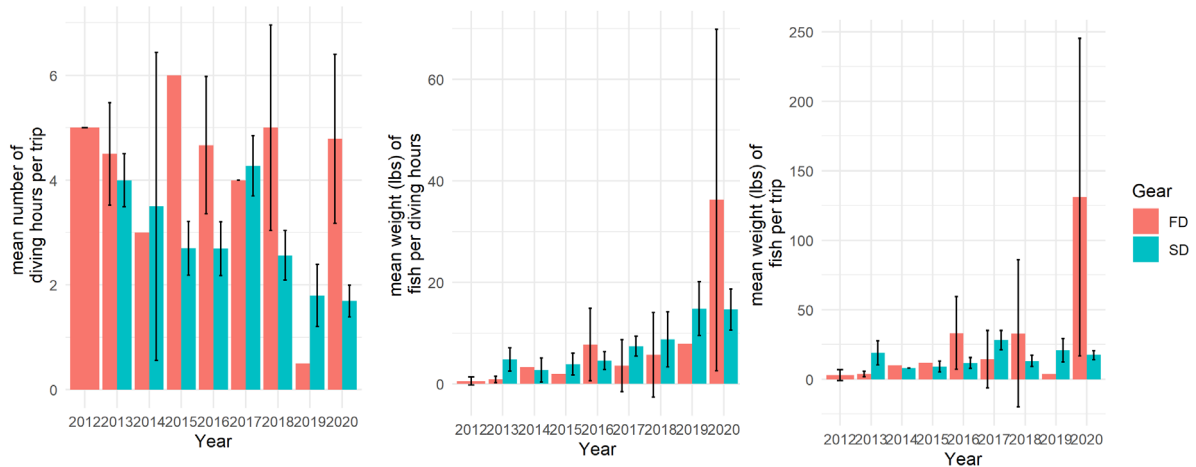

Figure 17. Raw number of diving hours per trip, in trips with mixed reef fish catches (left), mean raw mixed reef fish weight caught per diving hour (middle) and mean raw reef fish weight caught per trip (right) in Free and Scuba diving fisheries from 2012 to 2020.

3.3.3 Conch fishery

Landings and effort data

Landings and effort followed the same trends for conch catches as in the spiny lobster diving fishery. The number of conch caught per trip stayed relatively stable over time while the number of conch caught per diving hours increased over time (Figure 18).
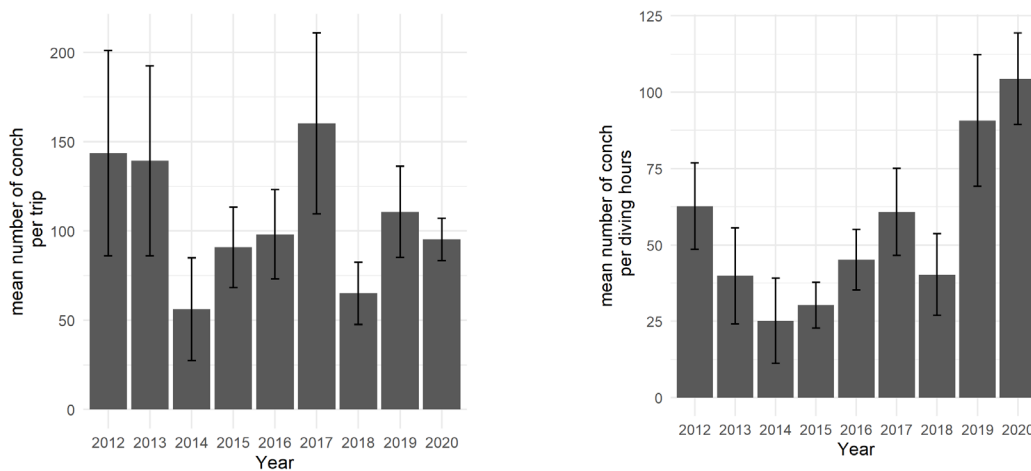

Figure 18. Mean raw conch number caught per diving trip (left) and mean raw conch number caught per diving hours (right) in Scuba diving fishery targetting conch from 2012 to 2020.

Landings lip thickness and shell-length composition

Lip thickness and shell-length measurements of the landings of conch were available for the period 2012-2020, with exception of 2014. There were some variations observed between years in the both mean lip thickness and mean shell-length, but with no clear trend (Figure 19). The mean lip thickness varies between 24.8 and 31.1mm (in 2017 and 2020, respectively). The range of variation for mean shell-length was small (half centimeter maximum difference in the annual mean for both sexes combined). The difference between sexes in mean lip thickness were not significant for most years, except in 2016, when lip thickness was markedly lower for females than for males. From 2012 to 2020 the females landed tended to be larger than males, by approximately 1 cm. In 2020 the reported lip thickness was higher by more than 3mm from the average thickness from 2012 to 2019. This particularly high value might be caused by the discovery of older conch 'hotspots' during this year.
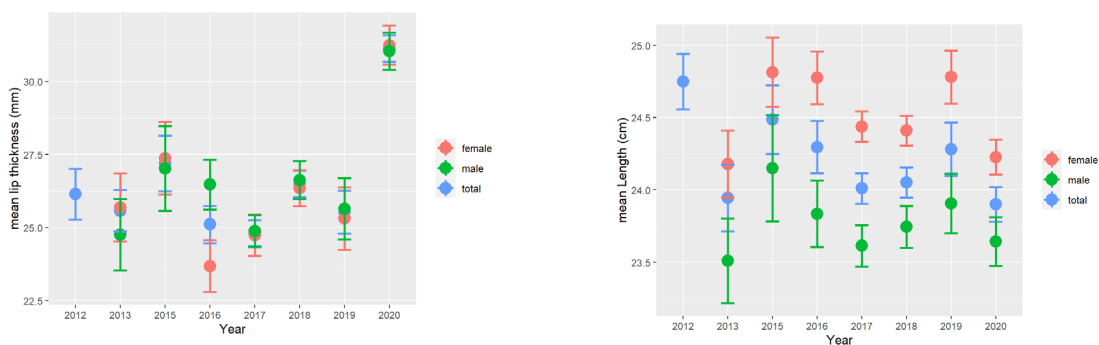

Figure 19. Mean lip thickness (left) and length (right) of conch from the landings (per sex and for sexes combined).

3.4 Handlines

3.4.1 Species composition of the landings

Species composition data for the handline fishery was scarce, and covers only 4 trips in 2017, 2 in 2019 and 8 in 2020 (Figure 20). In 2017 two of the three trips caught only Bigeye scad, *Selar crumenopthalmus*, while the other two mostly caught Serranidae. In 2019 and 2020, the sampled trips mainly caught Red hind and Coney. These two type of sampled trips actually correspond to two different types of fishery, one on smaller pelagic fish, the other on large benthic fish.
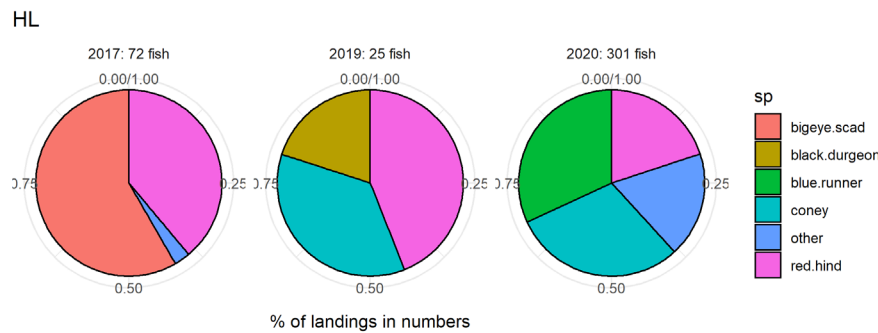

Figure 20. Species composition per year of the landings of the handline fishery (in weight) for the fish species only.

3.4.2 LPUE standardisation using a GLM model

Three boats that fished during more than half of the time period, >4 years out of 9 were selected. The effort in the handline fishery was represented in the LPUE-analysis by both the duration of the fishing operation and by the number of lines used. In the full model, including all effects (Table 8), only these 2 terms (length of time and number of lines), related to the effort, were significant (positive effect). Removing the non-significant effects one by one, the final model includes only effort-related variables (Table 8 and Figure 21). That the year effect was not significant indicates that in none of the years the abundance was different from the other years.

Table 7. Number of year selected boats were observed fishing from 2012 to 2020.

Boat	1	6	12
Coverage	9/9	7/9	7/9

Table 8. Significance of each model term tested by removing them stepwise and comparing to the full model (GLM model of reef fish landings per trip). See also Table 3.

Full model

Model term removed	Number of parameters	AIC	Log-Likelihood ratio	p.value	Signif. level
<none>		229.44			
factor(Year)	8	219.88	0.6288	0.7510538	
factor(Month)	11	224.99	1.3191	0.2311959	
factor(Boat_name)	2	227.24	0.6873	0.5061111	
logLines	1	232.94	4.2734	0.0422105	*
logDur	1	245.10	14.6145	0.0002728	***

Final model

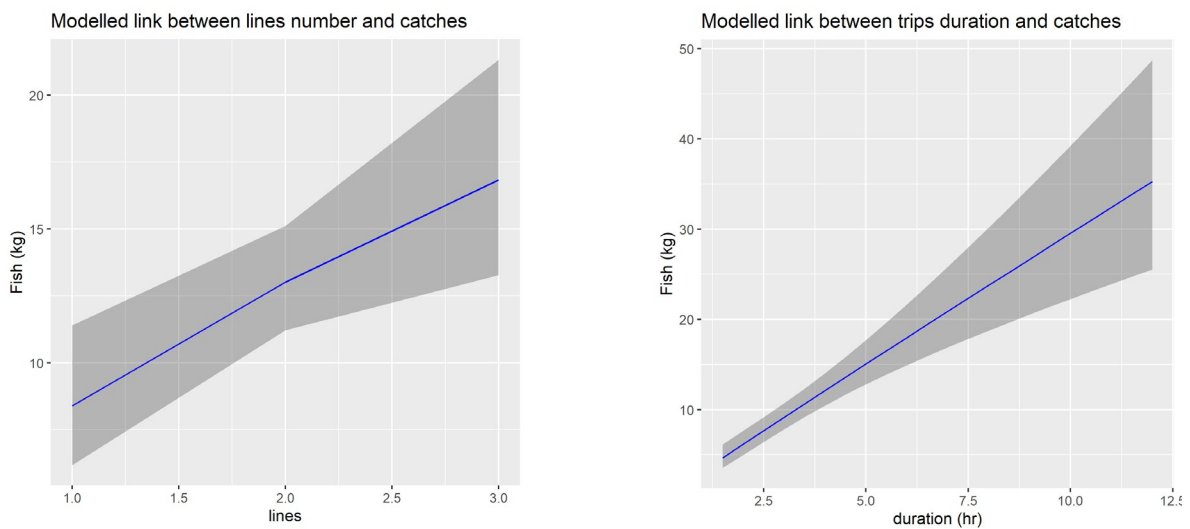| Model term removed	Number of parameters	AIC	Log-Likelihood ratio	p.value	Signif. level
<none>		214.89			
logLines	1	222.16	9.430	0.002784	**
logDur	1	261.46	60.942	7.681e-12	***

Figure 21. Modelled effect of fishing effort (number of lines and duration) on the landings of reef fish from the handline fisheries per trips (in numbers). Blue lines represent the modelled effect (and associated 95% uncertainty in grey).

3.5 Trolling

3.5.1 Species composition of the landings

The two most abundant species in the landings of the trolling fishery were Dolphin fish, *Coryphaena hippurus*, and Houndfish, *Tylosurus crocodilus*, depending on the year (Figure 22).
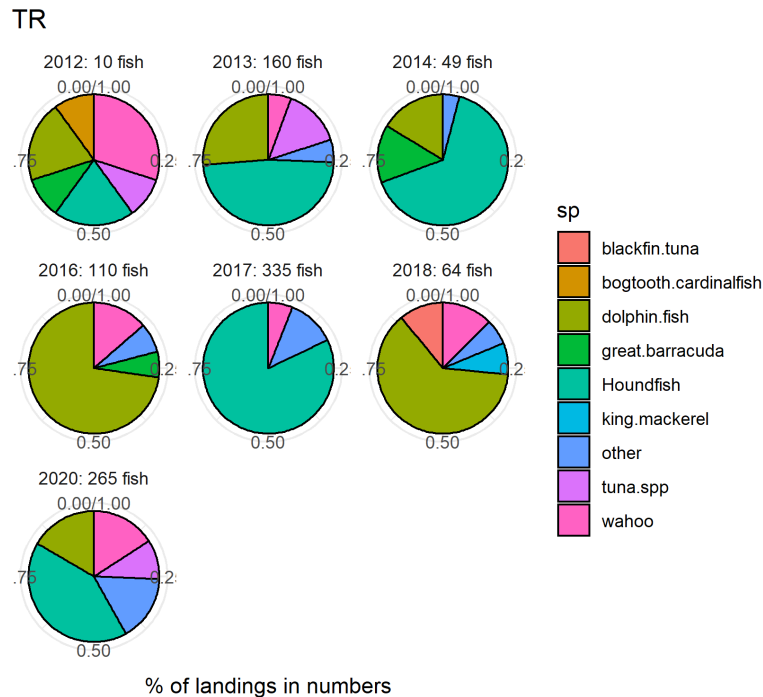

Figure 22. Yearly species composition of the landings of the trolling fishery (in weight) for the fish species.

3.5.2 LPUE standardisation using a GLM model

Four boats were selected fishing during more than half of the tie period (>4 years over 9) to conduct the analysis (Table 9). As for the handline fishery, the effort was represented by the number of lines and the duration of the fishing operation in the LPUE analysis. In the full model, including all effects (Table 10), the year and the boat effect were significant. Removing the none significant effects one by one, the final model includes a boat effect and a year effect (Table 10 and Figure 23). The year effect suggests a higher abundance of pelagic fish in recent years of the period covered by the study.

Table 9. Number of year selected boats were observed fishing from 2012 to 2020.

Boat	1	2	19	25
Coverage	7/9	6/9	6/9	5/9

Table 10. Significance of each model term tested by removing them one by one and comparing to the full model (GLM model of reef fish landings per trip). See also Table 3.

Full model

Model term removed	Number of parameters	AIC	Log-Likelihood ratio	p.value	Signif. level
<none		345.24			
factor(Year)	8	358.09	3.1032	0.0050610	**
factor(Month)	10	344.81	1.5935	0.1288592	
factor(Boat_name)	3	367.48	8.0719	0.0001219	***
logLines	1	344.26	0.7435	0.3917715	
logDur	1	345.28	1.4995	0.2252330	

Final model

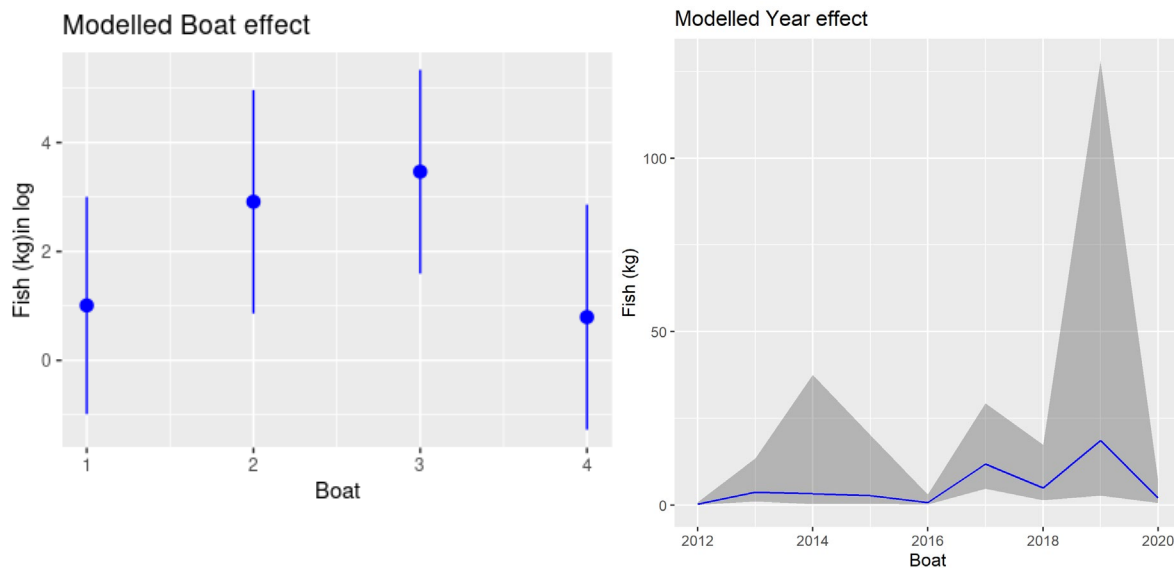| Model term removed	Number of parameters	AIC	Log-Likelihood ratio	p.value	Signif. level
<none>		347.49			
factor(Year)	8	373.16	5.7475	9.763e-06	***
factor(Boat_name)	3	372.65	10.7577	5.549e-06	***

Figure 23. Modelled boat and year effects on the landings of pelagic fish from the trolling fisheries per trips (in weight). Blue lines represent the modelled effect (and associated 95% uncertainty in grey).

4   Discussion

De Graaf et al. (2015) analyzed an earlier version of the St Eustatius fisheries database. They estimated an average annual landings over the years 2012-2015 of 11 tonnes for lobster, 3.8 tonnes for reef fish and 0.8 tonnes of pelagic fish. They report an average number of ~500 trips per year over this period. The analyses presented here estimate much higher annual landings over the same period, which is mostly explained by the much higher effort estimated (1,001 trips per year). This nearly two-fold difference in effort between the current study and de Graaf et al. (2015) is due to the raising done in this report to take account of the incomplete coverage of the boat activity survey (i.e. activity data covering about 50% the active days in the month in 2012-2017, Figure 1). This correction was not done in the 2015 study.

The later preliminary study of Kitson-Walters (2018) focusing on the year 2017 estimated annual landings of lobster and fish of 52 and 17 tonnes, respectively. Those were substantially higher than the estimates for the same year from the present study of 25 and 9 tonnes, respectively. Estimates of effort and landings from 2017 should be considered with caution, because of the lack of boat activity survey for the last 3 quarters of the year. Both studies attempted to correct for the lack in the effort data by inferring boat activity from the logbook data, but in both cases, the methods used come with high associated uncertainties.

More generally, given the many gaps in the data, any attempt to estimate catches has to make doubtful assumptions and such calculations should be considered as highly uncertain. The main assumptions that were made in the present study were:

- Boat activity for the days without activity survey is similar to the activity for the days with survey. This extrapolation represents more than 50% of the effort over the period 2012-2017.

- The proportion of the fishing time allocated to the different fishing methods and the resulting landings per trip are the same for the boats providing logbooks and for those not providing logbooks (this extrapolation represents between 30% and 50% of the boats, depending on the year).

- In previous reports the raising method use to estimate these missing proportions was based on an average proportion along the entire time period (2012-2018). Subsequently, adding data impacted the previous estimates average proportion. Henceforth, the raising method will considered the period before 2018 as an 'historic' time period with raising done based on average proportion from 2012 to 2018. Then, the new data will be treated by period of two years to calculate average proportion to achieve the raising without impacting previous data.

The coverage of the activity survey has markedly increased in the last years of the period studied here (Figure 1), and has been close to 100% in the last three years. This is a positive development, which, if sustained, will result in a much better quantification of the fishing effort. Even so, for several of the fishing metiers, the number of catches quantified (in terms of totals caught) and species-specific subsampling for species composition and size-structure data remain very low (Table 1) and variable, with as a consequence that the confidence/credibility of the results remain seriously compromised. For conch and lobster the attained samples sizes were the best.

The precision of the catch estimates could be improved by increasing the number of boats delivering logbooks. Given the diversity of fishing patterns observed in the available logbooks (i.e. distribution of fishing time between different fishing methods), and also given the significant boat effect in most GLM models, it seems necessary to increase the number of boats delivering logbooks.

The St Eustatius lobster fishery (with between 16 and 48 tonnes caught annually) contributes a very small fraction of the regional total annual landings of spiny lobster for the Caribbean (30,000 tonnes,

de Graaf et al. 2015). However, considering the small size of the fishing area, these landings correspond to a production of at least 500kg/km^2 annually, one the highest reported in the Caribbean (Table 4.1 from de Graaf et al. 2015). De Graaf et al. (2015) bring forward the hypothesis that high yields on St. Eustatius may be related to a naturally high productivity in this area, explained by a combination of large nursery areas and unlimited recruitment of pelagic larvae most likely from other Caribbean islands. However, they also note that such high yields per km^2 have only been observed in the Caribbean during the developing phase of the lobster fishery between 1970 and 1990, as based on overfishing, and consider it doubtful that these high landings were sustainable.

The individual size of the lobsters landed may be a reason for concern. Average size of landed lobster was 95 mm and 102.5 mm for females and males respectively and 41% of the landed lobsters were under the minimum legal size of 95 mm CL along the entire time period. This problem is even more severe for females, for which 56% are harvested below the legal size. Dilrosun (2004) measured 43 harvested lobsters in November 2003 and reported an average length of 110 mm CL, well above the minimum legal size. It is important to understand why undersized lobsters are now being landed by the fishermen. Do the fishermen intentionally target smaller individuals to answer a market demand for plate-sized undersized lobsters? The size data might be biased due to the sample design, which included mainly lobsters from the trap fishery and from one fisherman that consistently participate to the data collection. If these two reasons could be excluded, the small size of the lobsters currently landed on St Eustatius could indicate a truncation of the age/size structure of the population towards younger/smaller individuals, which would suggest that the stock is overexploited. The estimation of F/F_{MSY} ratio proxies support this hypothesis and indicated a fishing mortality above sustainable level along the entire time period, with fishing mortality being higher for females than males.

Abundance indices for lobster in the scuba fisheries were not presented. Indeed, boats are supposed to present variable catchabilities and the lack of consistency in the fleet involvement over time make it impossible to differentiate boat and year effects. The only time trend that might be used to identify trends in spiny lobster stock biomass will be the standardised number extracted from the trap fishery. This index shows strong variability over the time period considered with no unidirectional trend. Continuation of a robust fishery monitoring (catch, effort, length frequency) is needed to provide fishery-dependent indices (time series) that could be used to provide a more accurate assessment of the stock and, thereby, a better basis for management. With respect to the lobster fishery, we conclude that the apparent decline in lobster abundance in 2018 (based on the lobster trap LPUE index) and persistent high landing rate of sublegal-sized lobsters (discussed above) is worrisome. Furthermore, the high values for F/F_{MSY} proxies mainly for females supports the idea of an unsustainable fishing mortality. While, given all the uncertainties in the data, it might be premature to already conclude that there is actual overfishing taking place, it seems that the early signs of overfishing might already be present. Therefore, in any case the lobster catches should be monitored with more precision and caution and the landing of sublegal lobsters needs to be controlled with better awareness and enforcement.

Serranidae are typically among the most valuable commercial reef fish species. While the mixed fish catch from the lobster trap fishery are dominated by low-value Acanthuridae and Ostraciidae, two Serranid species (Coney, *Cephalopholis fulva* and Red Hind, *Epinephelus guttatus*) jointly still make up 12% of the total trap landings weight. The Coney is a small species, whereas the Red Hind is a medium-sized grouper and the most vulnerable of the two to overfishing. This latter species has even disappeared from catches in various areas of the Caribbean (Debrot and Criens 2005, Vermeij et al. 2019, Debrot et al. 2020). For St Eustatius, de Graaf et al. 2015 indicated both of these species to likely be seriously over-exploited. The amount of data on these species and other low-value bycatch species like parrotfish and surgeonfish from the current port sampling activity is limited and patchy. Hence, more intensive port sampling data and additional fishery-independent study of these two species of value and concern are certainly to be recommended. Likewise, the data on the handlining, trolling and the newly developing FAD fishery (next paragraph) are so few and scattered that a significant analysis of trends does not yet seem warranted. Although these gears are very sporadically used in St. Eustatius, it might be valuable to keep track of this gear's effort developments to adapt the port sampling design if they become more frequently used in the following years

Finally, of particular concern is the fact that an unofficial Fish Attracting Device (FAD) fishery has been developing around St. Eustatius, but data on this fishery are not yet available. This fishery targets

pelagic fish species that also caught in the small-scale trolling fishery. These species are totally different from those caught in the traditional and predominant reef-based fishery. FAD fisheries have proven very problematic elsewhere, leading to cases of severe overfishing and serious disputes between fishermen (Samples and Sproul 1985, CRFM 2015, Sadusky et al. 2018). There is an urgent need to obtain data from this new development and to guide the development of this fishery. While FAD fishing likely presents valuable opportunities for further fisheries development in St. Eustatius (see also recommendations for Bonaire: Debrot and van den Berg 2019), it should be managed and contained, not only to allow it to develop to its full potential but also to avoid the many pitfalls associated with this form of fishing (Figure 21).

Figure 21. Example of FAD catch from nearby Saba. (Photo: Ayumi Kuramae).

In conclusion, our main recommendations in terms of both management and research and monitoring are:

Management:
- Improve control of and compliance with lobster size-limit regulations.
- Develop a FAD fishery management plan as part of a St. Eustatius fisheries development plan.

Research and monitoring:
- Improve port sampling (for overall catch recording) and subsampling (for size-frequency data) intensity to cover at least one third of the trips dedicated to each key fishing metiér (these are the following fisheries metiérs: lobster pot, conch, spearfishing, trolling, FAD fishery). At present we are jointly developing better guidelines for these port sampling activities.
- Conduct a closer study on the Queen triggerfish, the Coney and the Red Hind. Do this by combining more intensive port sampling and fisheries-independent studies on the distribution and abundance of these species around St. Eustatius.

Quality Assurance

Wageningen Marine Research utilises an ISO 9001:2015 certified quality management system. This certificate is valid until 15 December 2021. The organisation has been certified since 27 February 2001. The certification was issued by DNV GL.

Furthermore, the chemical laboratory at IJmuiden has EN-ISO/IEC 17025:2017 accreditation for test laboratories with number L097. This accreditation is valid until 1th of April 2021 and was first issued on 27 March 1997. Accreditation was granted by the Council for Accreditation. The chemical laboratory at IJmuiden has thus demonstrated its ability to provide valid results according a technically competent manner and to work according to the ISO 17025 standard. The scope (L097) of de accredited analytical methods can be found at the website of the Council for Accreditation (www.rva.nl).

On the basis of this accreditation, the quality characteristic Q is awarded to the results of those components which are incorporated in the scope, provided they comply with all quality requirements. The quality characteristic Q is stated in the tables with the results. If, the quality characteristic Q is not mentioned, the reason why is explained.

The quality of the test methods is ensured in various ways. The accuracy of the analysis is regularly assessed by participation in inter-laboratory performance studies including those organized by QUASIMEME. If no inter-laboratory study is available, a second-level control is performed. In addition, a first-level control is performed for each series of measurements.
In addition to the line controls the following general quality controls are carried out:
- Blank research.
- Recovery.
- Internal standard
- Injection standard.
- Sensitivity.

The above controls are described in Wageningen Marine Research working instruction ISW 2.10.2.105. If desired, information regarding the performance characteristics of the analytical methods is available at the chemical laboratory at IJmuiden.

If the quality cannot be guaranteed, appropriate measures are taken.

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Justification

Report C059/21
Project Number: 4318100260

The scientific quality of this report has been peer reviewed by a colleague scientist and a member of the Management Team of Wageningen Marine Research

Approved: Esther Beukhof
                     Researcher

Signature:

Date: 09-07-2021

Approved: Drs. J. Asjes
                     Manager integration

Signature:

Date: 09-07-2021

Annex 1 : Logbook forms used in data collection

STATIA FISHERIES RESEARCH LOGBOOK

Date:		Boat:		No. Crew:

Fishing gear: **Pot**   **Diving**

Free Diving ☐   Scuba Diving ☐

	Pot	Diving	
No. of pots hauled:			:Speargun (Yes/No)
Soak time (days):			:Dive time (hours)
Depth max/min (ft):	/	/	:Depth max/min (ft)
			:Divers (No.)
Fishing zone:			:Fishing zone
Reef Fish (lbs):			:Reef Fish (lbs)
Red fish (lbs):			:Red fish (lbs)
Lobster (No.):			:Lobster (No.)
Short lobsters (No.):			:Short lobsters (No.)
Berried Females (No.):			:Berried Females (No.)
Other_____(No/lbs):			:Other_____(No/lbs)

STATIA FISHERIES RESEARCH LOGBOOK

Date:		Boat:		No. Crew:

Fishing gear:	Handline	Longline
No. of lines:		
Duration (hours):		
Depth (ft):		
Quadrat:		
Reef Fish (lbs):		
Red fish (lbs):		
Pelagics (lbs):		
Other_____(No/lbs):		

Finshing gear:Trolling	Species	No.	Lbs
No. Lines:			
Duration (hr):			
Quadrat:			

Annex 2 : Life history parameters

Table Life History parameters used as input for F/F_{MSY} proxy estimations

Species	L∞	References L_{50}	L_{95}	
Acanthurus bahianus	189	160	165*	(Mutz, 2006); (Hawkins et al., 2007); (Wolfe, 2003)
Acanthurus chirurgus	268	175	180*	(Nagelkerken & Faunce, 2007); (Mutz, 2006)
Acanthurus coeruleus	276	145	150	(Mutz, 2006); (Hawkins et al., 2007)
Balistes vetula	450	250	280	(de Albuquerque, Martins, de Oliveira, de Araújo, & Ribeiro, 2011; Munro, 1983b; Rivera-Hernández et al.,2019)
Cephalopholis fulva	377	185	215	(Burton, Potts, & Carr, 2015; Marques, 2011)
Epinephelus guttatus	393	215	270	(Sadovy et al., 1994); (Ault, Bohnsack, & Meester,1998); (Cushion, 2010); (Ault et al., 2002)
Holocentrus adscensionis	265	146	159	(Munro, 1983a);(Shinozaki-Mendes, Vieira Hazin, De Oliveira, & De Carvalho, 2007)
Mulloidichthys martinicus	300	175	185	(Munro, 1983a)
Ocyurus chrysurus	484	213	245	(Trejo-Martínez, Brulé, Mena-Loría, Colás-Marrufo, & Sánchez-Crespo, 2011); (Ault, Smith, Luo, Monaco, & Appeldoorn, 2008); (Ault et al., 2002)
Sparisoma viride	357	205	235	(Choat, Robertson, Ackerman, & Posada, 2003); (Figuerola, Matos-Caraballo, & Torres, 1998); (Ault et al., 2002)

*Values are estimated based on the steep growth curve of closely related species. It is assumed that there is little difference between L50 and L95, the assumption is based on the parameters found for *A. coeruleus*.

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Annex 3 : Monthly estimates of effort, catch rates and landings per métier per species category.

Lobster :

Monthly variations num. trips fishing lobster

Monthly variations of the lobster yields
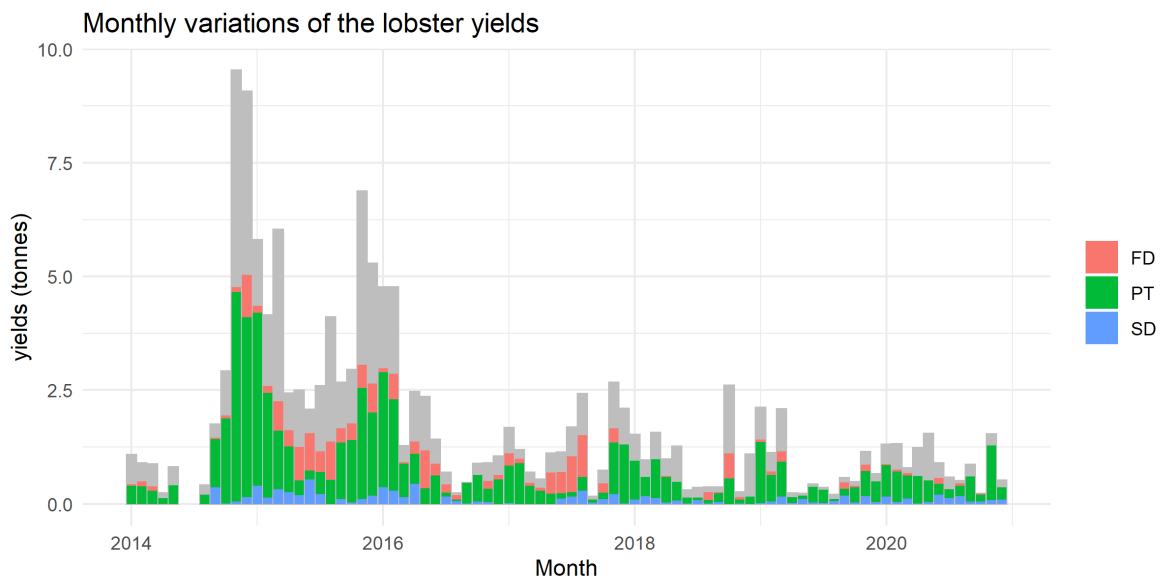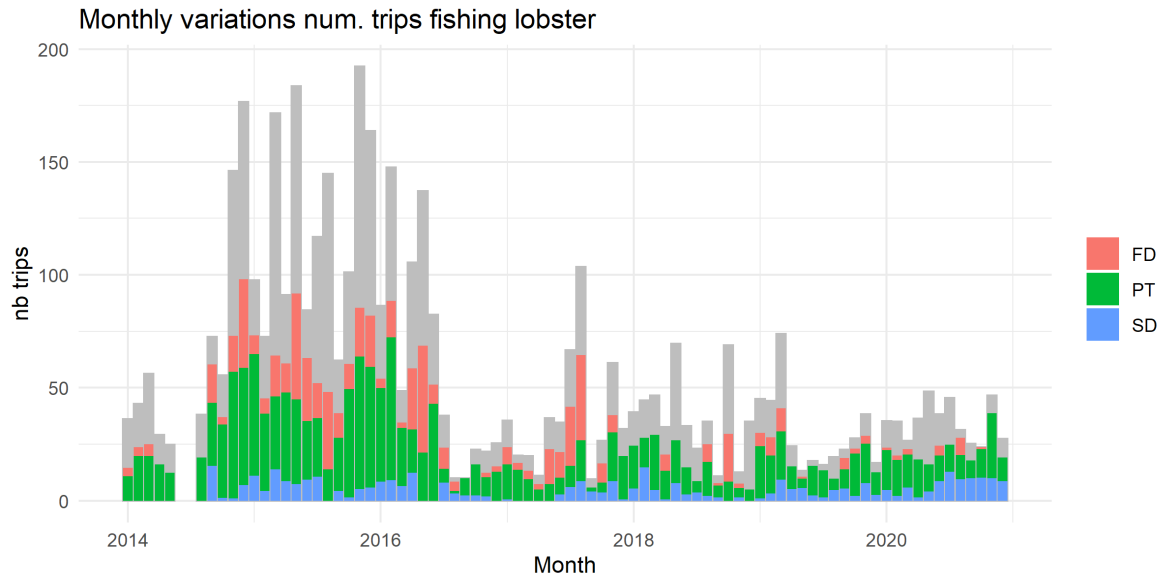
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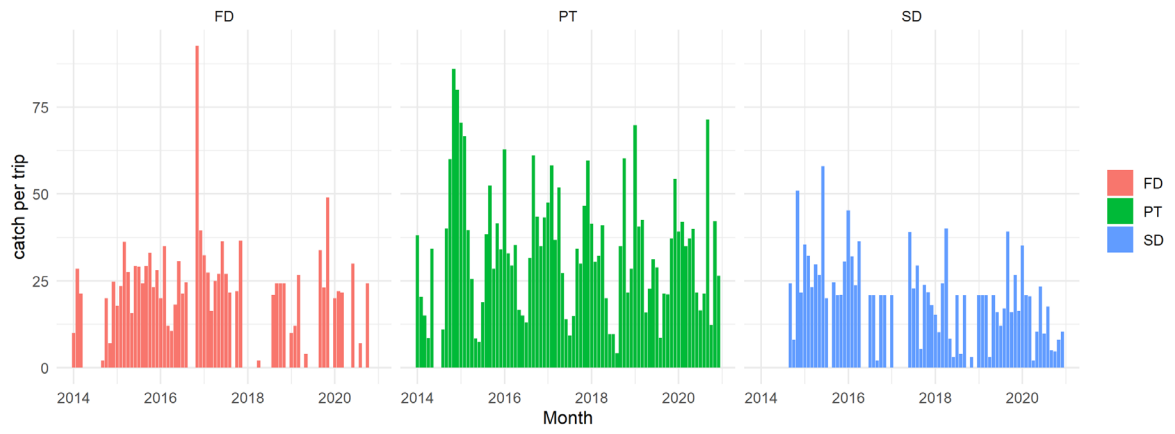
Reef fish :

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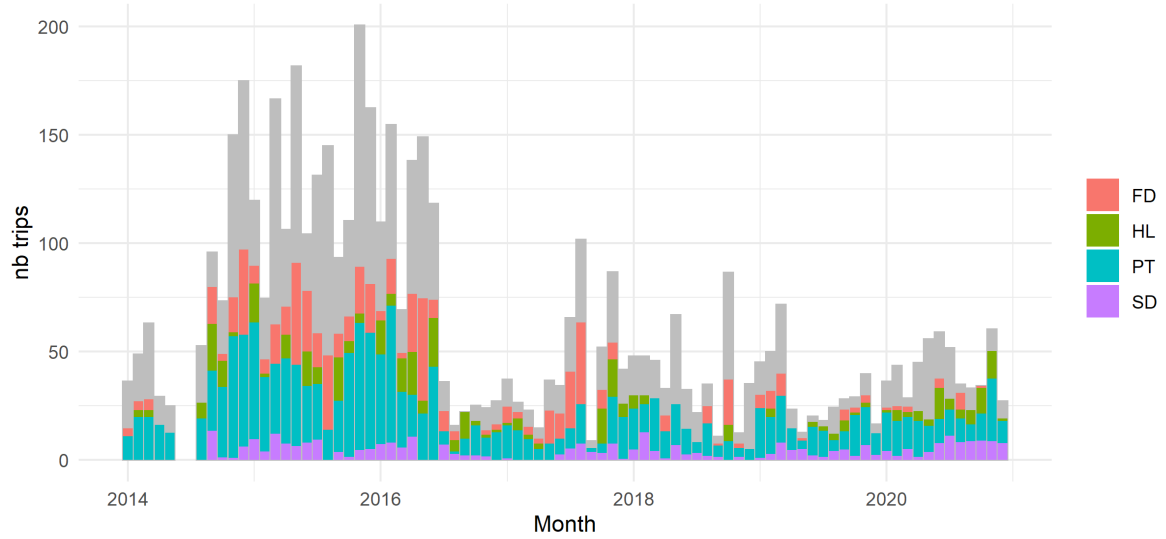
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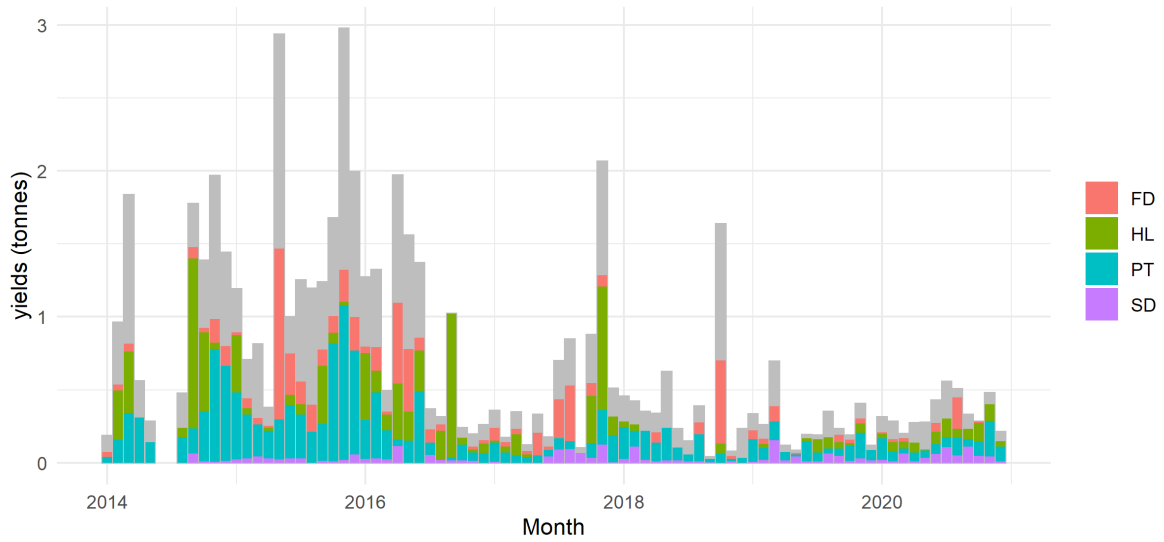
Monthly variations catch rates reef fish

Pelagic fish

Monthly variations num. trips fishing pelagic fish

Monthly variations of the pelagic fish yields
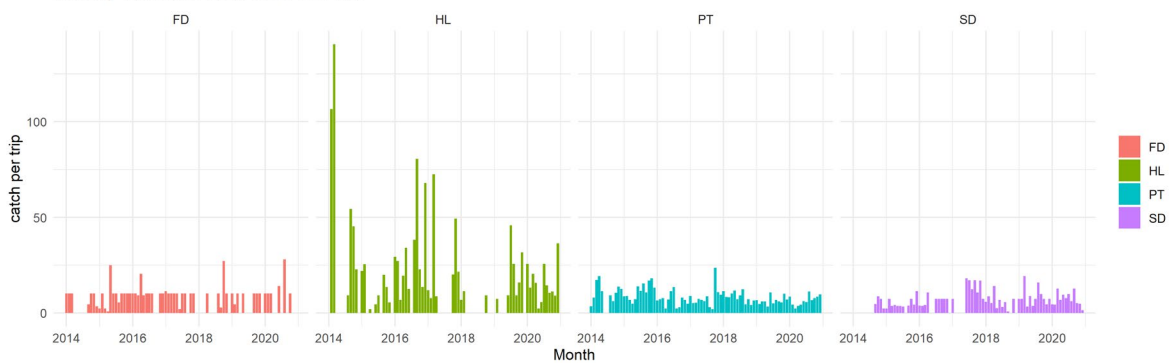
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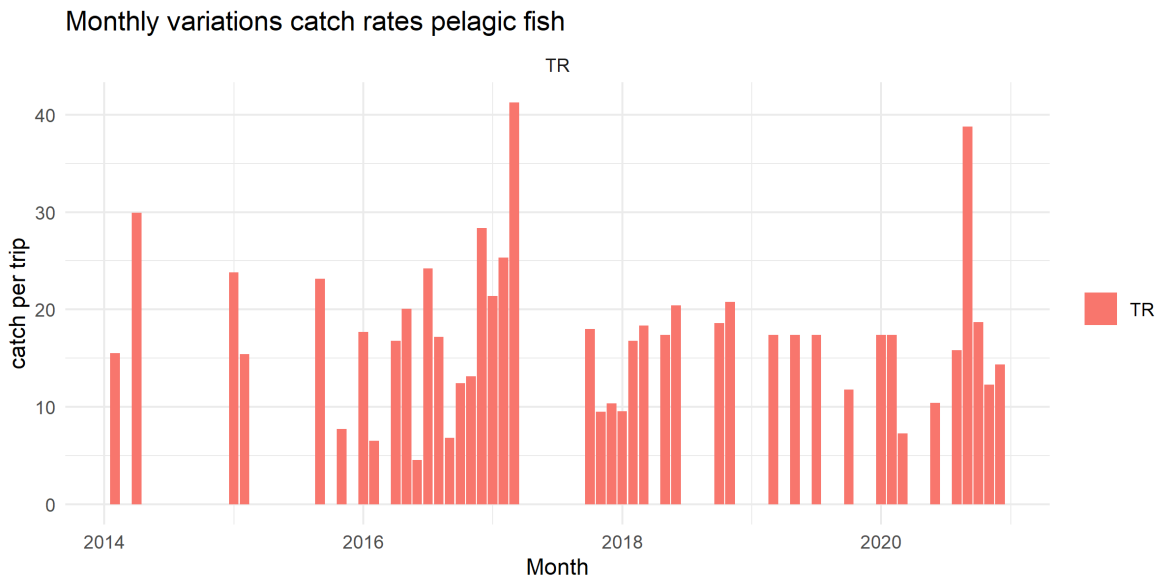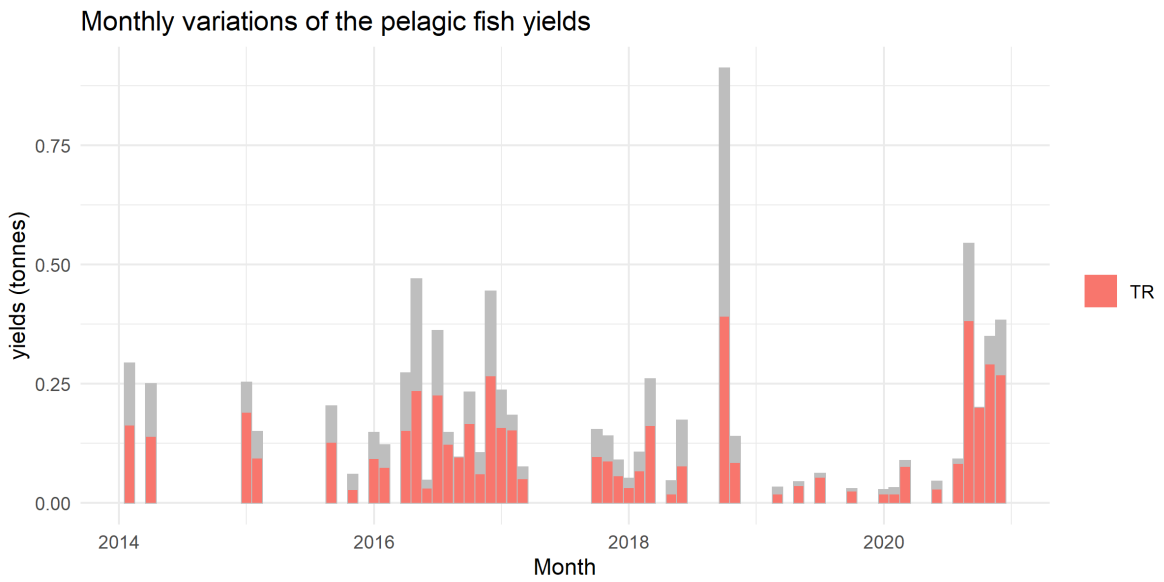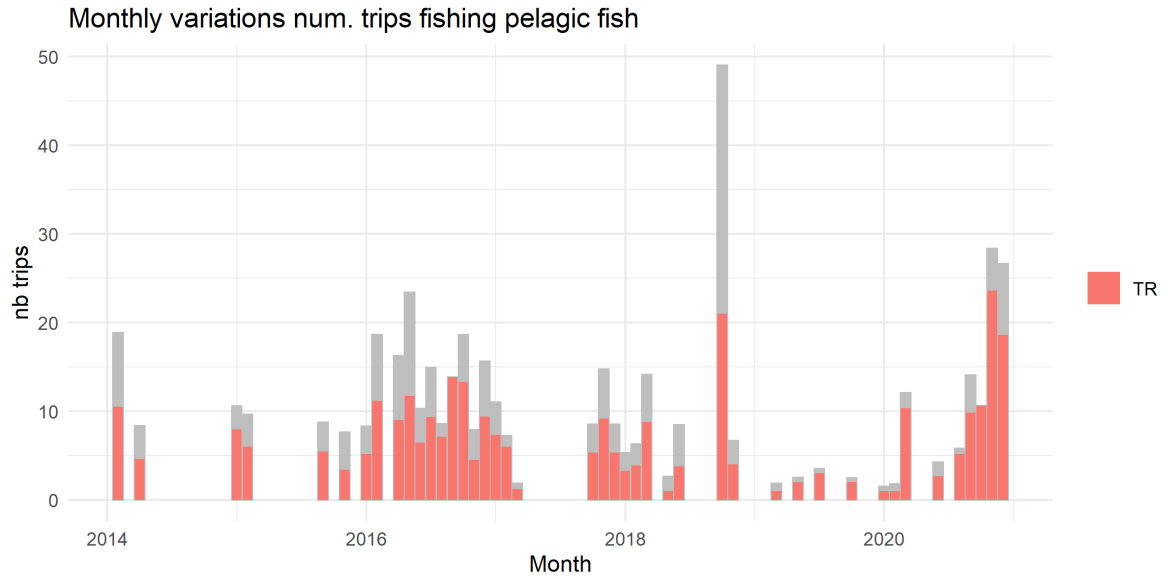
Conch

Monthly variations num. trips fishing
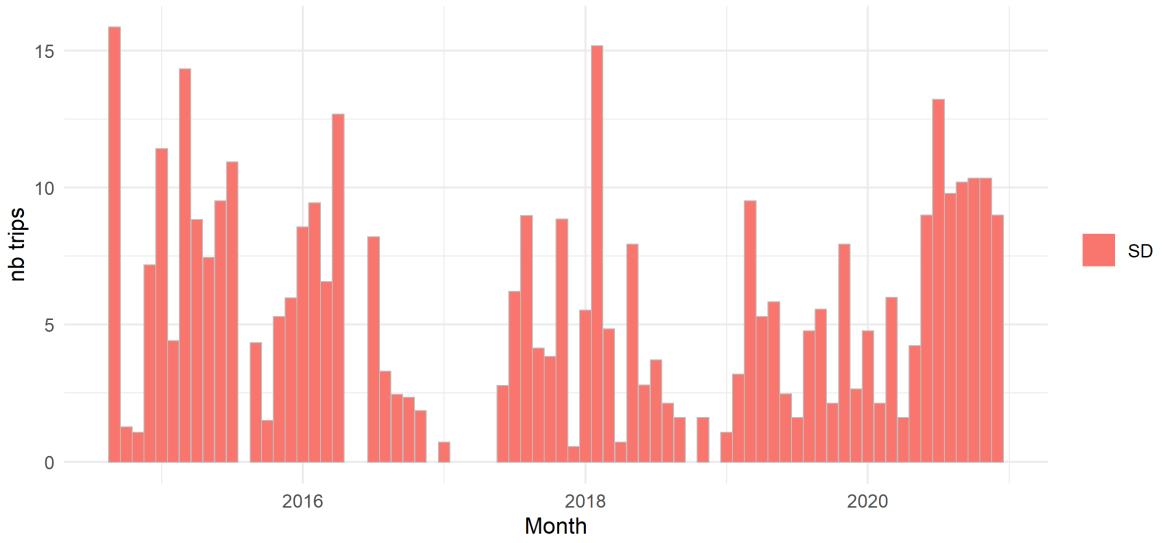
Monthly variations of the conch catches
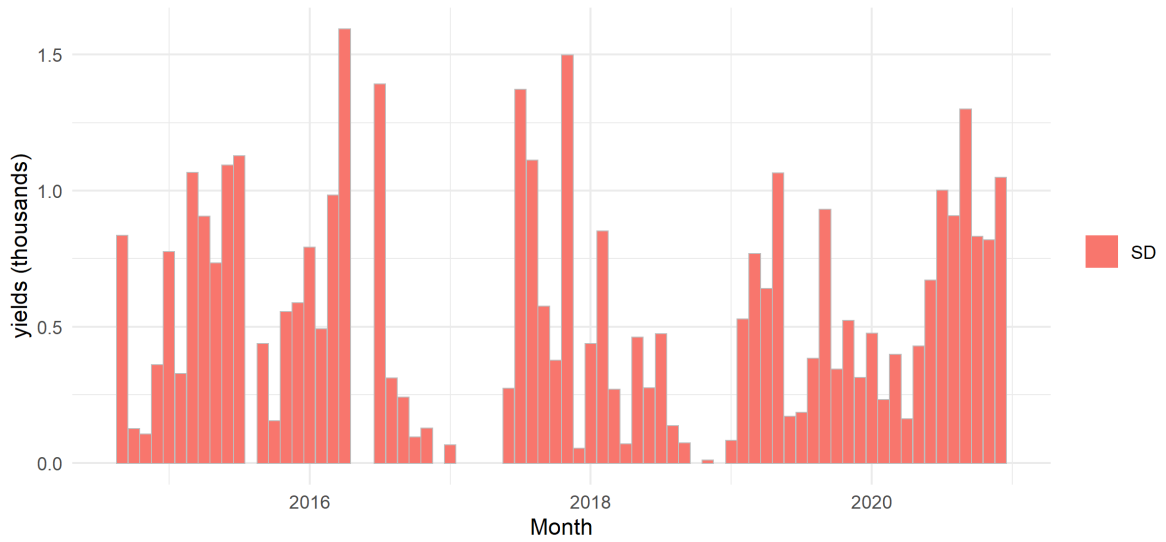
Monthly variations catch rates conchs
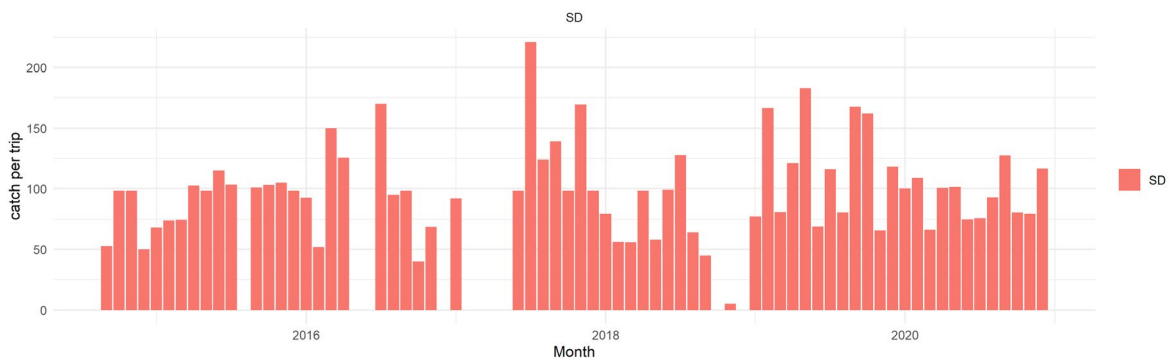
Annex 4 : Length frequency distribution of the reef fish landings in fish traps

The following plots show the length distribution of the landings per species and per year, with the mean length shown as a vertical line.

Length distribution blue tang
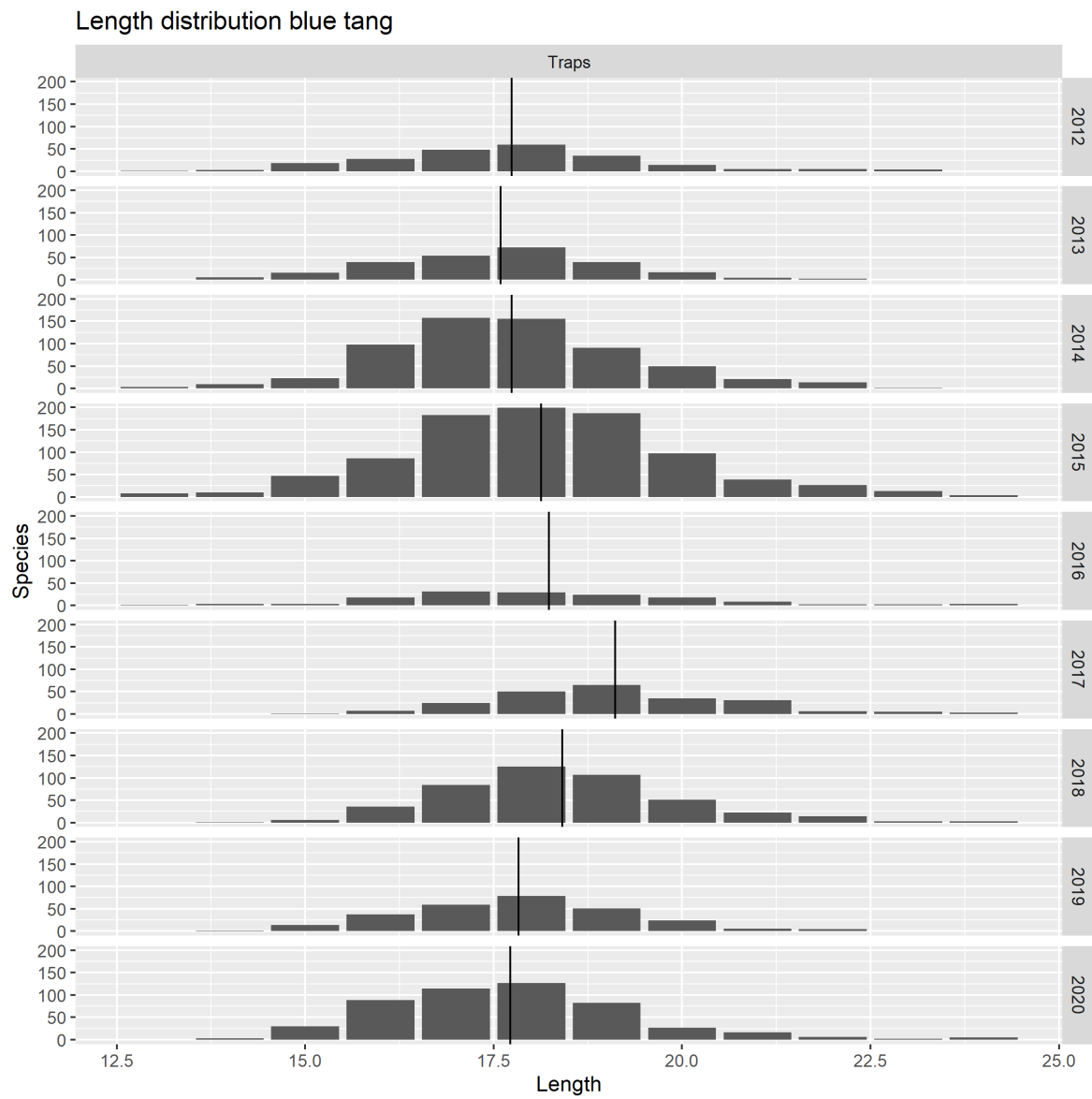
Length distribution coney
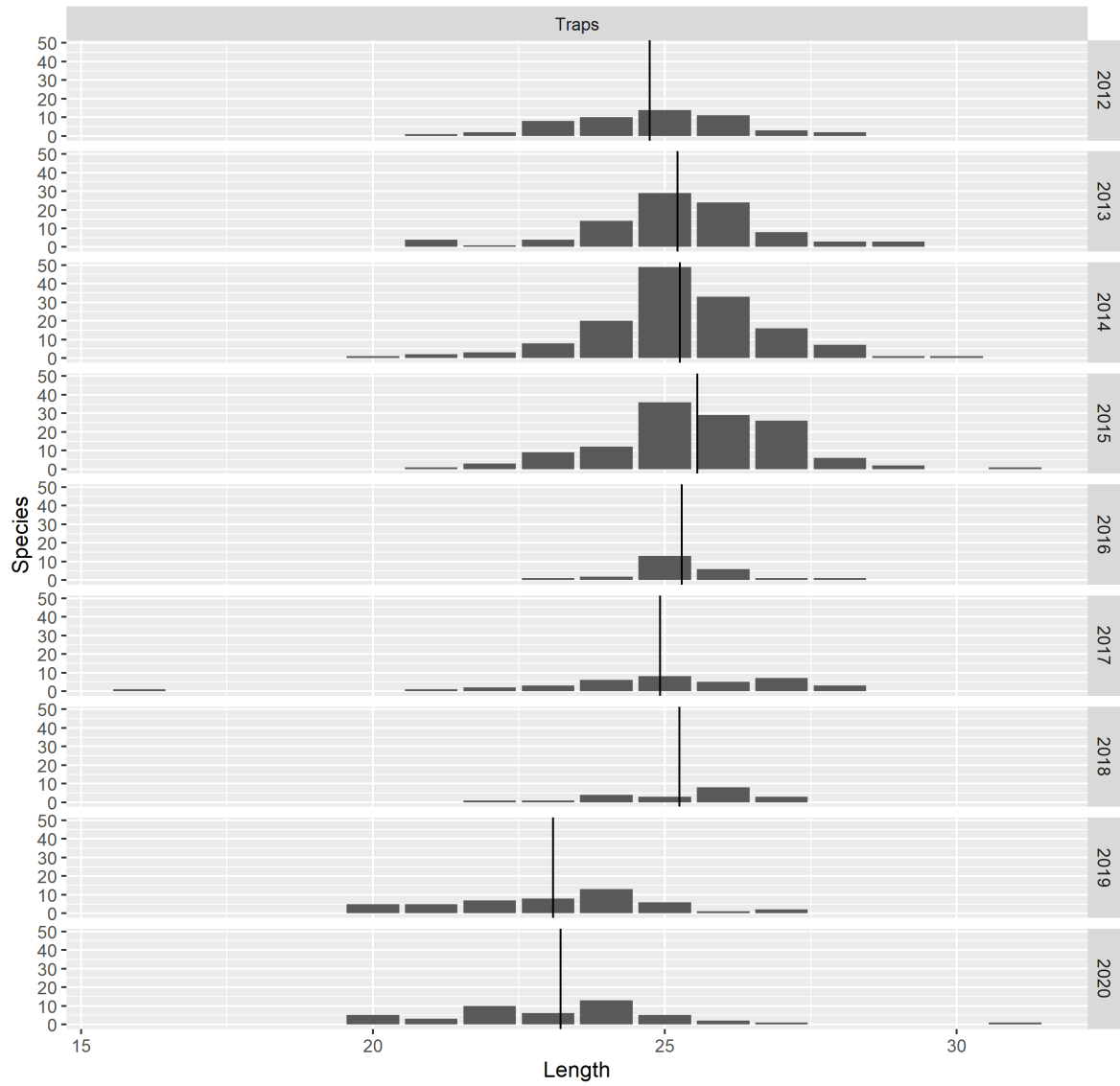
Length distribution doctorfish
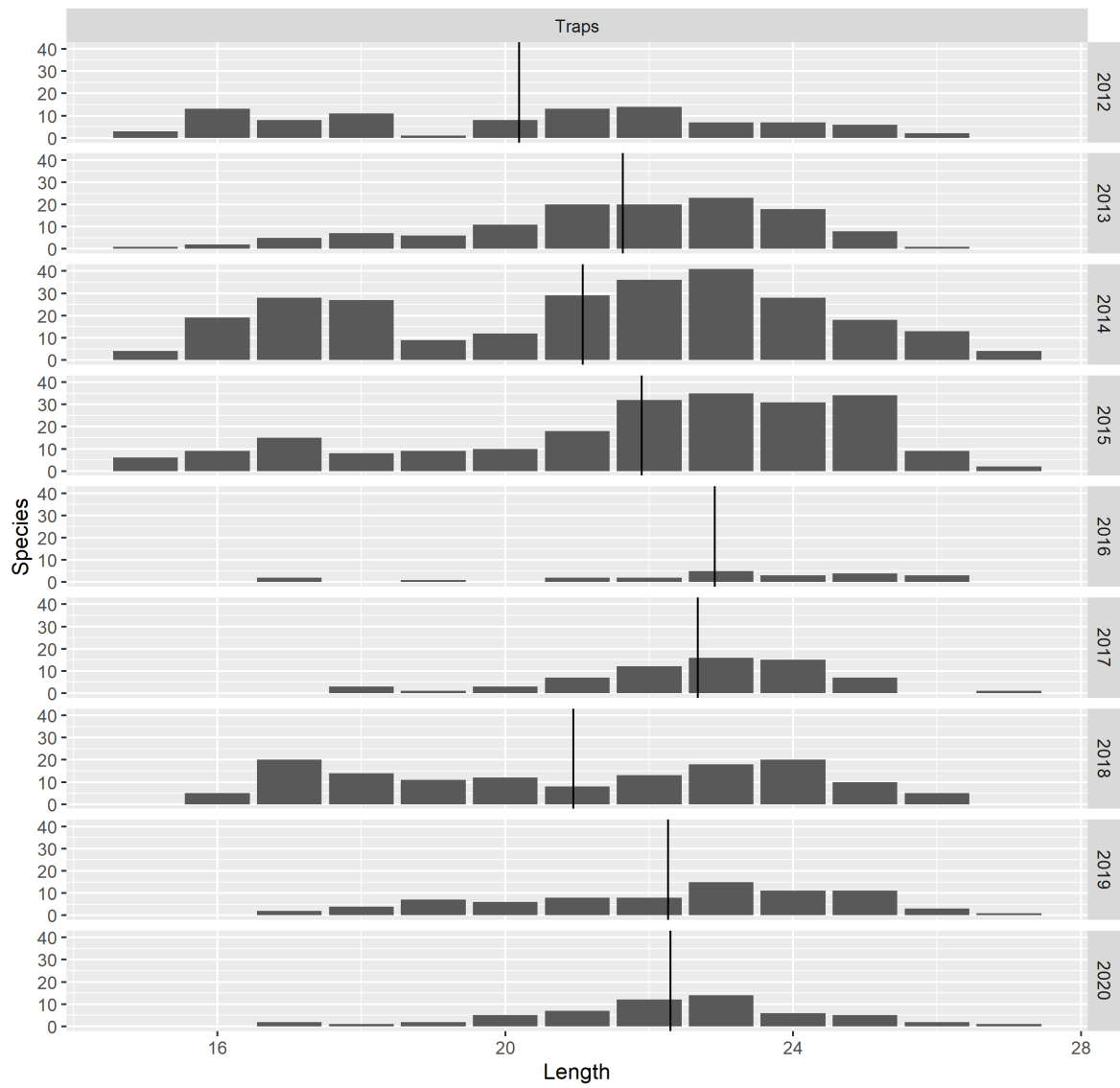
Length distribution honeycomb cowfish
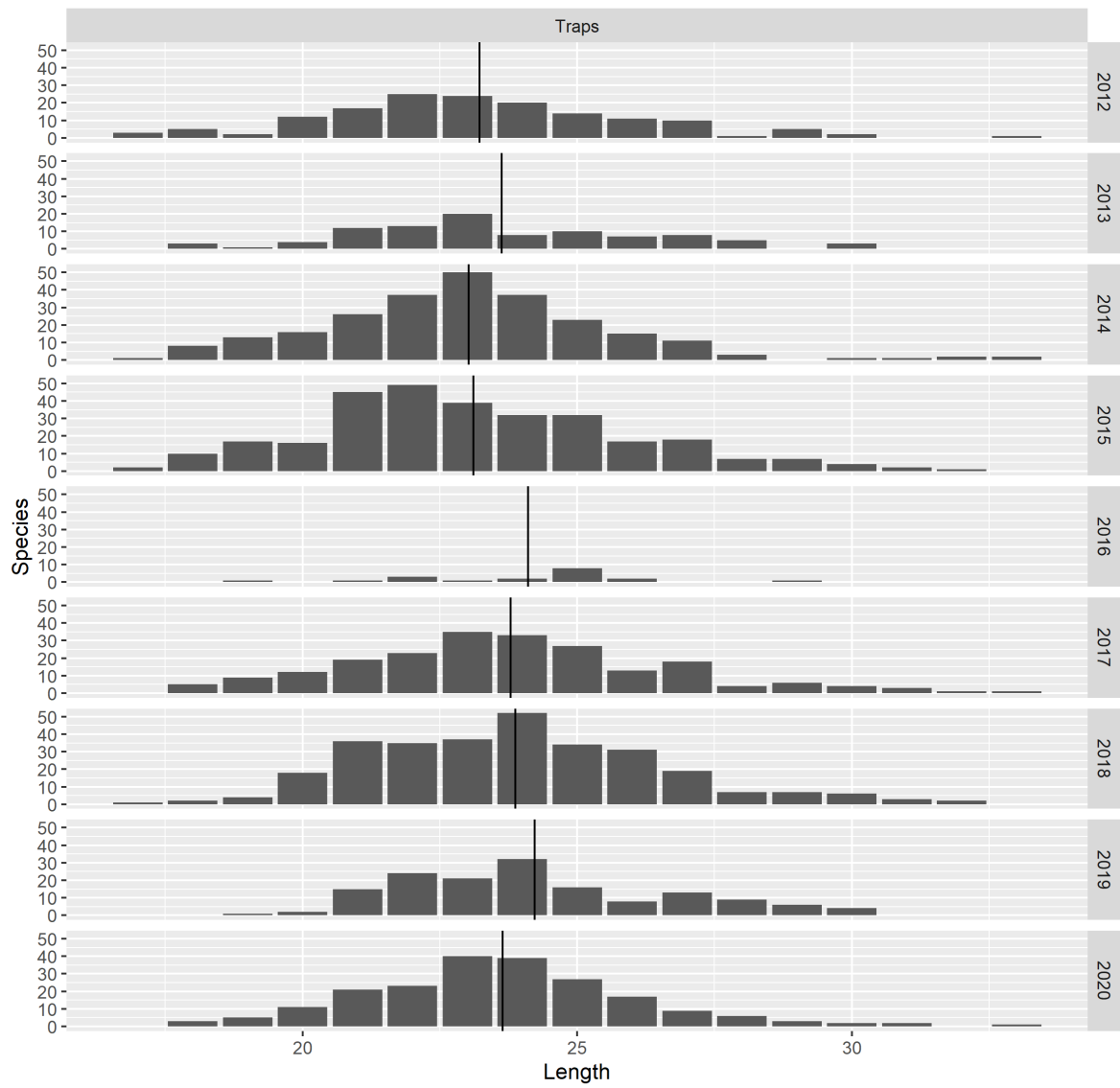
Length distribution ocean surgeon
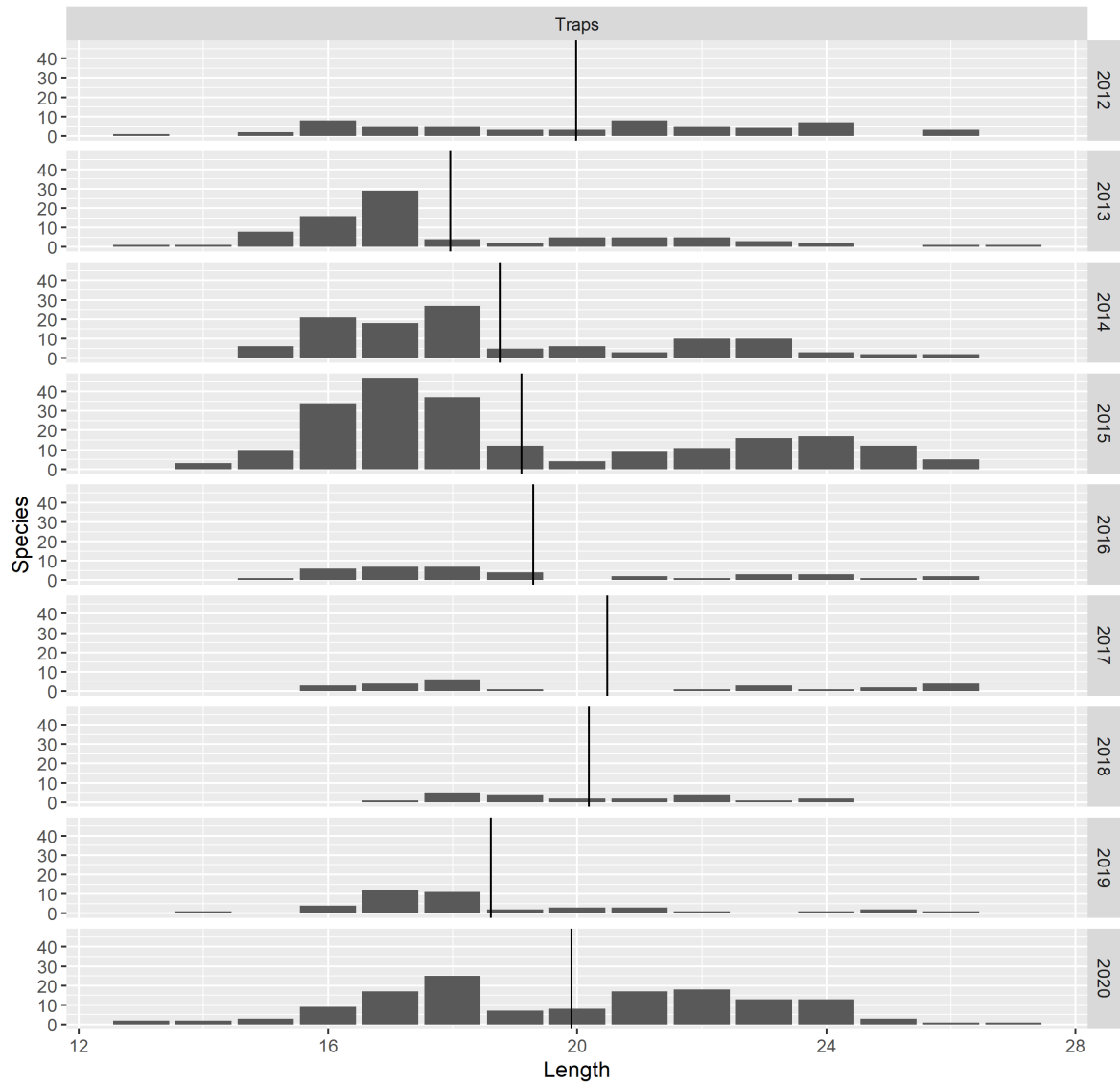
Length distribution queen triggerfish
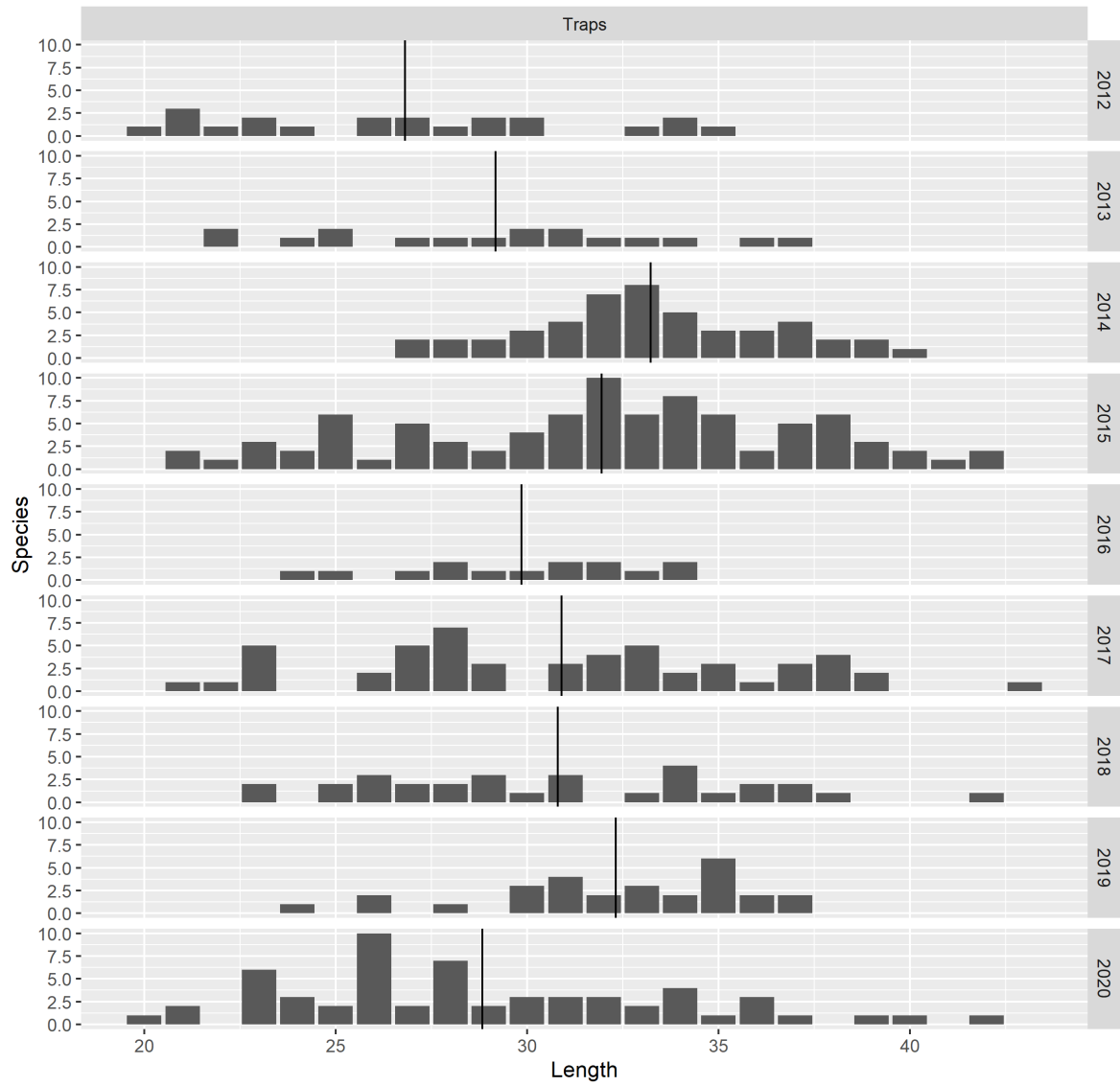
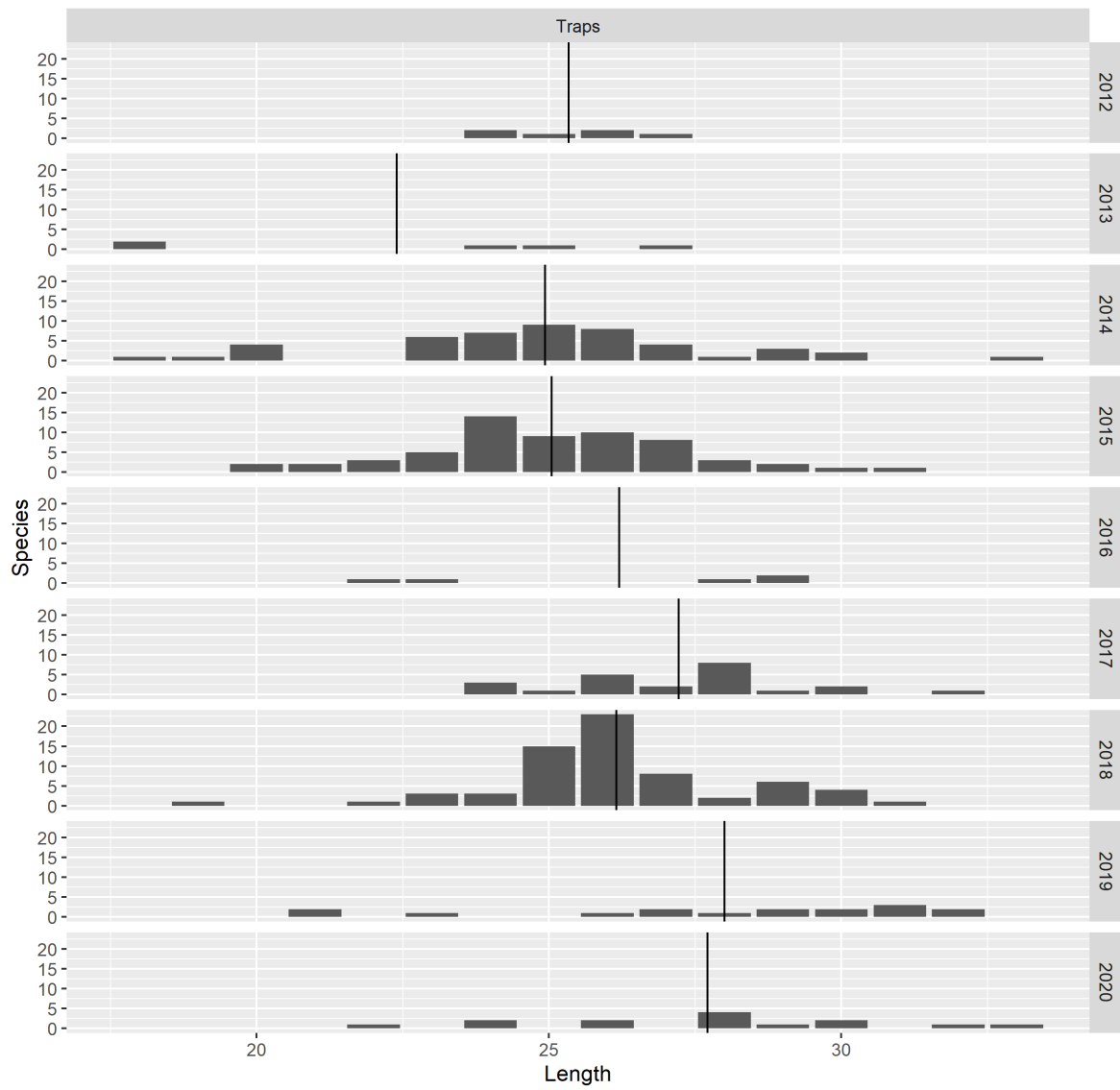Length distribution red hind
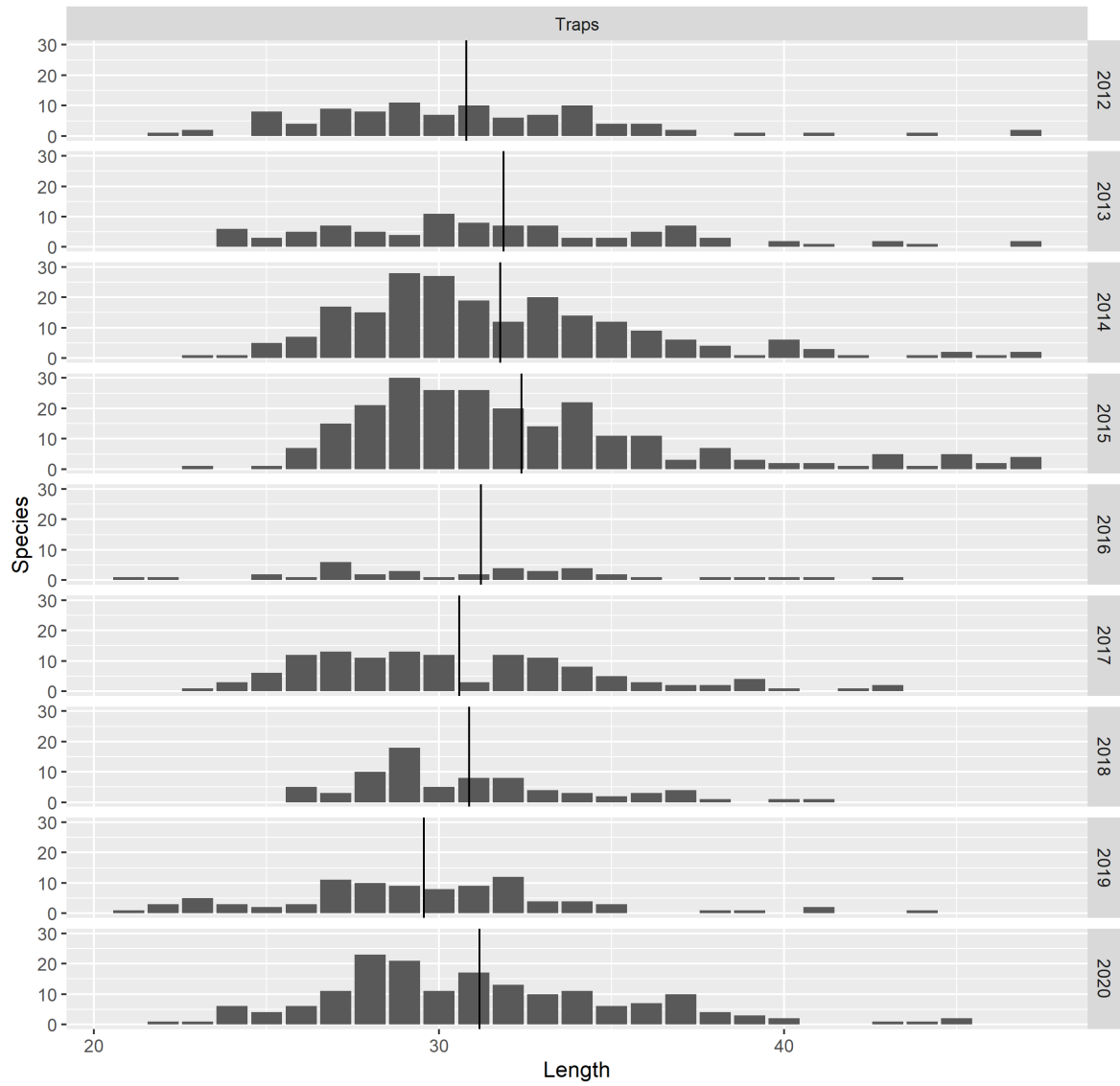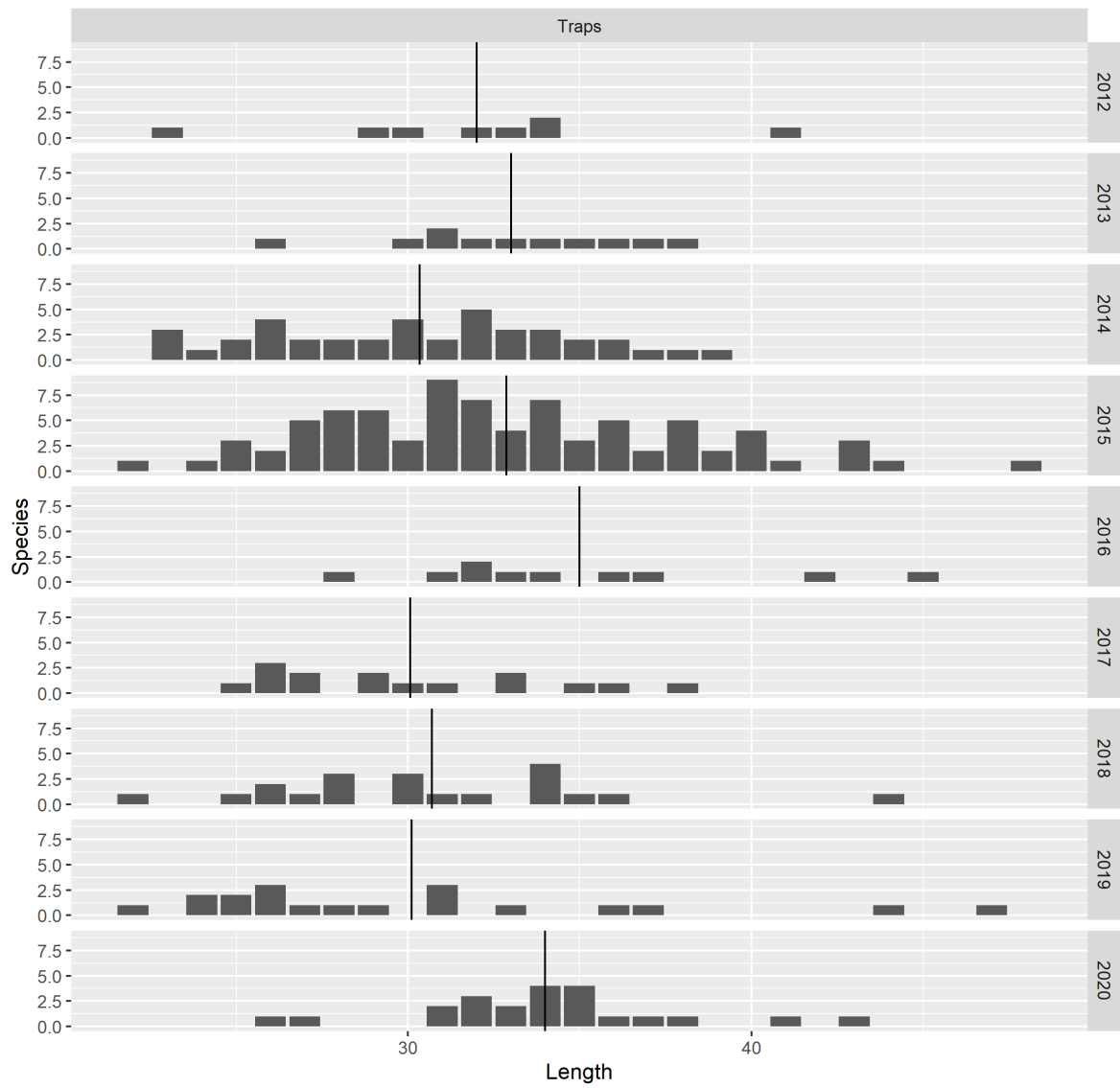
Length distribution squirrelfish
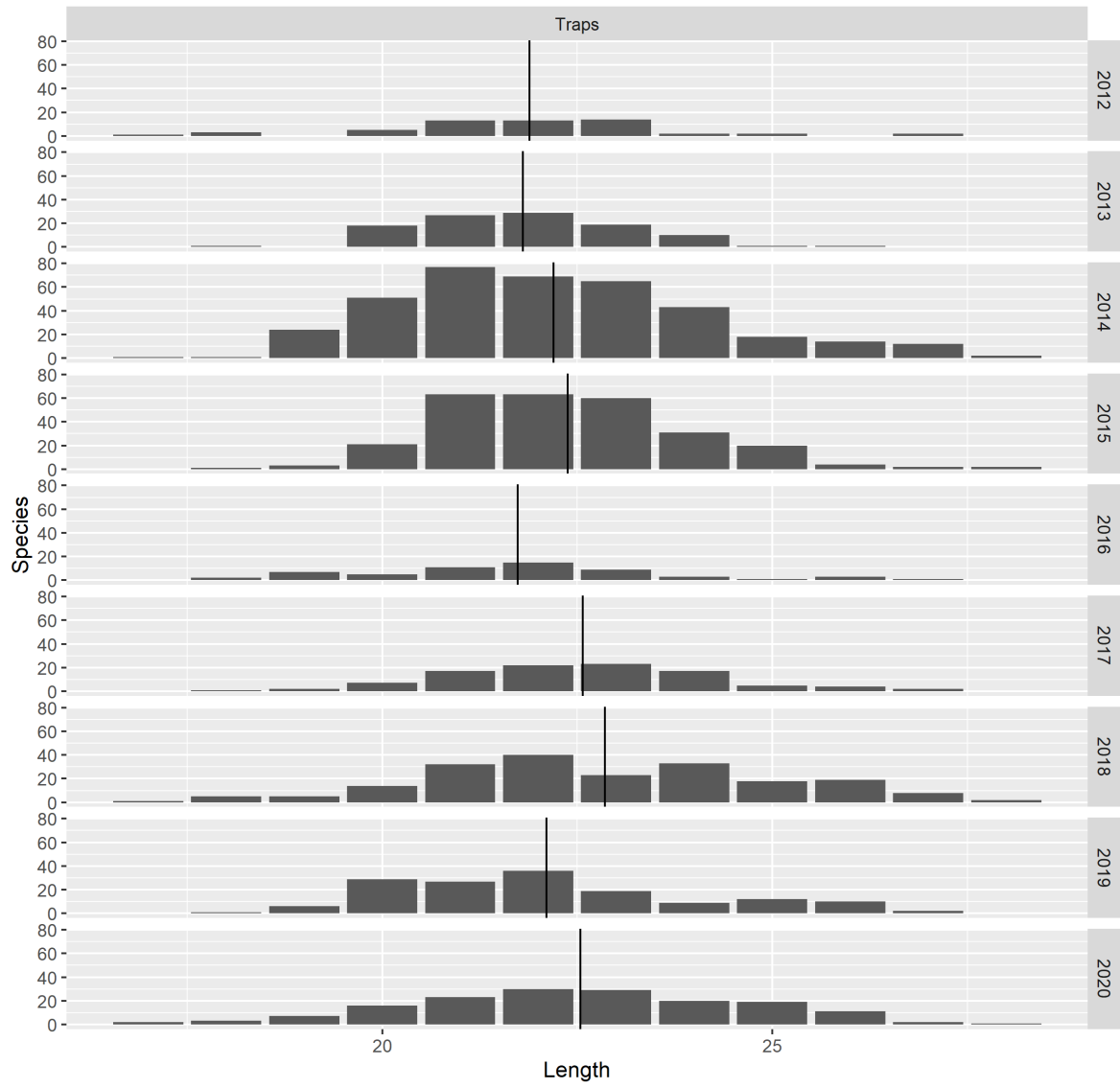
Length distribution stoplight parrotfish
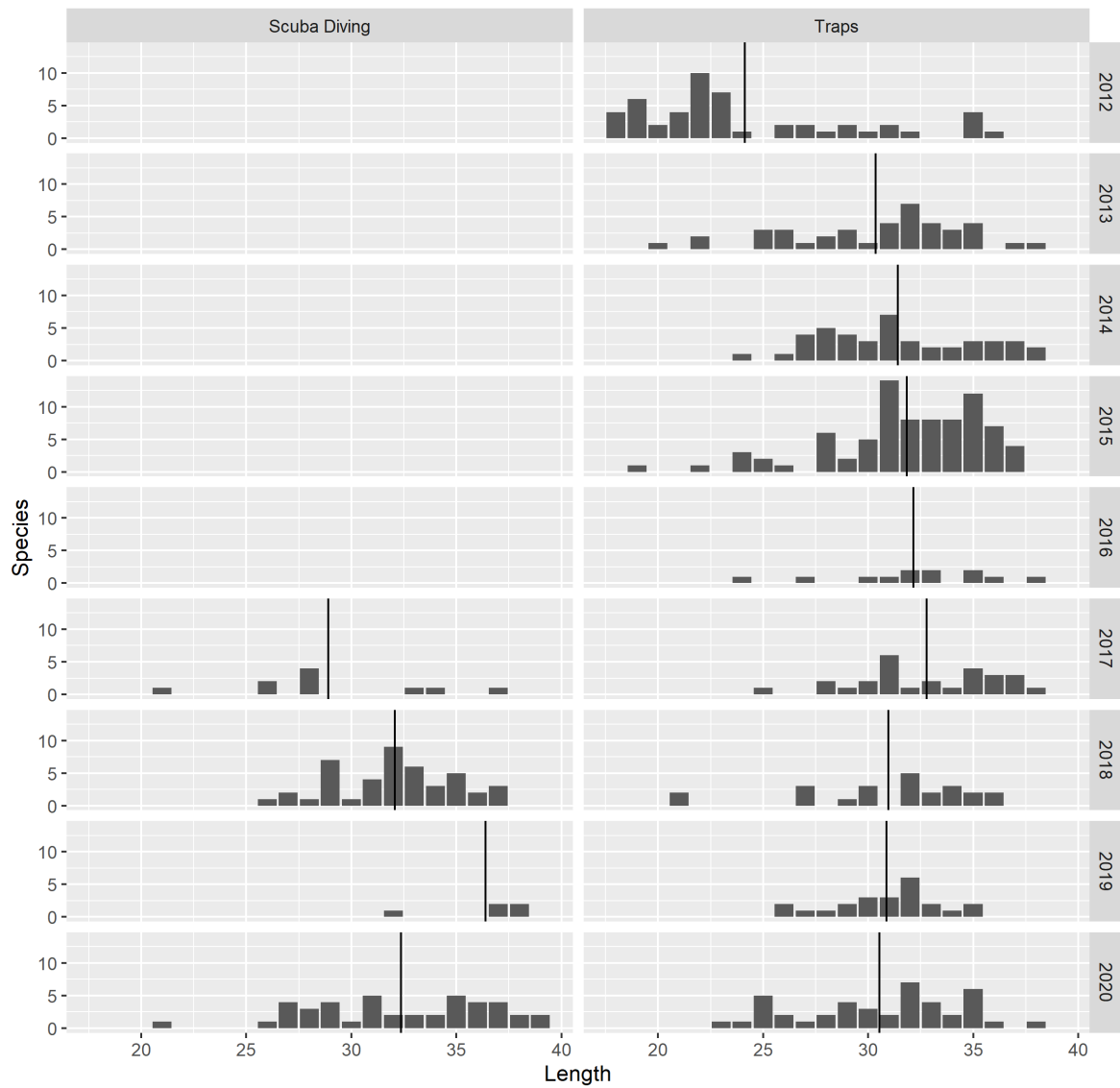
Length distribution white spotted filefish
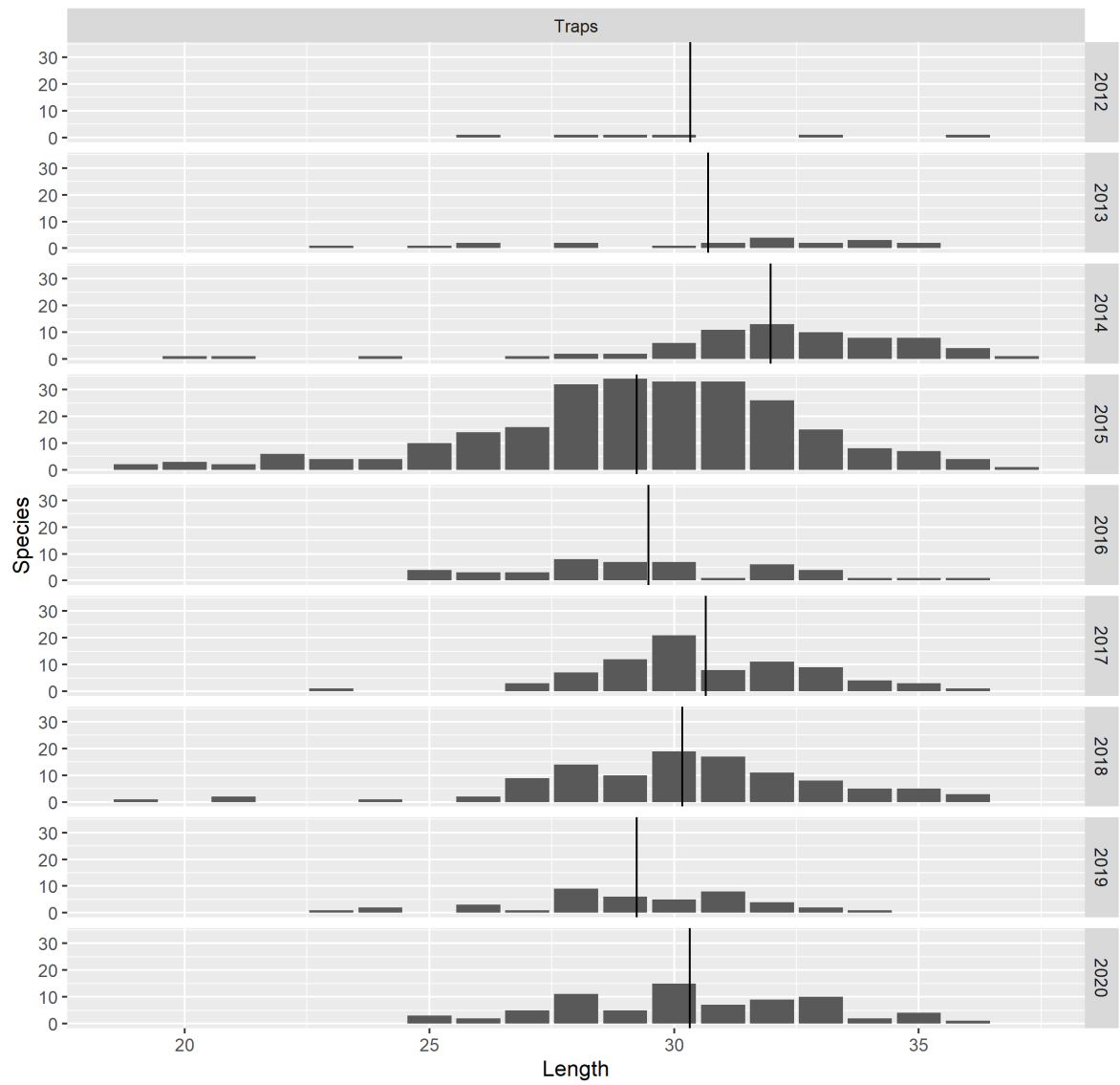
Length distribution yellow goat fish

Length distribution yellowtail snapper

With knowledge, independent scientific research and advice, **Wageningen Marine Research** substantially contributes to more sustainable and more careful management, use and protection of natural riches in marine, coastal and freshwater areas.

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