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This is a "Post-Print" accepted manuscript, which has been Published in "Science of the Total Environment"

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Please cite this publication as follows:

Zhang, S., Yang, X., Gertsen, H., Peters, P., Salánki, T., & Geissen, V. (2018). A simple method for the extraction and identification of light density microplastics from soil. *Science of the Total Environment*, 616-617, 1056-1065.
<https://doi.org/10.1016/j.scitotenv.2017.10.213>

You can download the published version at:

<https://doi.org/10.1016/j.scitotenv.2017.10.213>

A simple method for the extraction and identification of microplastic from soil

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1 **Abstract:** This article introduces a simple and cost-saving method developed to extract,
2 distinguish and quantify microplastics in soil. A floatation method using distilled water was
3 used to extract light density polyethylene microplastics from soil samples. Microplastics and
4 impurities were identified using a heating method (3-5 seconds at 130°C). The number and
5 size of particles were determined using a camera (Leica DFC 425) connected to a microscope
6 (Leica wild M3C, Type S, simple light, 6.4X). Quantification of the microplastics was
7 conducted using a developed model. Results showed that the floatation method was effective
8 in extracting microplastics from soils, with recovery rates of approximately 90%. After being
9 exposed to heat, the microplastics in the soil samples melted and were transformed into
10 circular transparent particles while other impurities, such as organic matter and silicates were
11 not changed by the heat. Regression analysis of microplastic weight and particle volume after
12 heating showed the best fit ($R^2 = 99\%$, $p < 0.001$). Recovery rates based on the empirical model
13 method were over 80%. Results from field samples collected from North-western China
14 prove that our method of repetitive floatation and heating can be used to extract, distinguish
15 and quantify light density polyethylene microplastics in soils. Microplastic mass can be
16 evaluated using the empirical model.

17 **Key words:** Light density polyethylene microplastics; soil; extraction; identification;

18

19 1. Introduction:

20 Disposable plastic is heavily used world-wide but unfortunately, not always recycled
21 (Hidalgo-Ruz et al., 2012; Steinmetz et al., 2016; Yan et al.). Generally, plastic, in its many
22 forms, ends up as debris in the environment (Yan et al., 2015). It is widely deposited and
23 dispersed in water, sediment and soil, which threatens the sustainable development of healthy
24 ecosystems for generations (Cooper and Corcoran, 2010; Hidalgo-Ruz et al., 2012). Recently,
25 many studies have focused on the adverse effects of plastics in the aquatic environment

26 (Hidalgo-Ruz et al., 2012; Wang et al., 2016b). Microplastics are small particles < 5 mm
27 derived from plastic from different origin and composition such as polyethylene,
28 polypropylene, polysterol etc. They are insoluble in water, nondegradable (according to
29 standardized tests) and possess different physicochemical properties which are key aspects
30 that determine its bioavailability to organisms (Leslie, 2014; Rocha-Santos and Duarte, 2015).

31 Many studies focus on the fate and effect of microplastics in the aquatic environment
32 whereas there is a gap of knowledge on microplastics in the terrestrial environment. Plastics
33 can enter into the terrestrial ecosystem by areal deposition, transport from landfills, sludge
34 application to agricultural land and the application of light density polyethylene (LDPE)
35 plastic film as mulch on agricultural land. Plastic mulch application to millions of ha
36 worldwide with increasing tendency has let to plastic pollution of the terrestrial environment,
37 however no monitoring on the occurrence of microplastics in the agricultural soils exists and
38 no analytical techniques are described for their extraction from soils and identification. Their
39 fate in the terrestrial environment is unknown and only few studies exist on the effect on soil
40 organisms (Huerta et al. 2016, Huerta et al. 2017) and the accumulation in the terrestrial food
41 chain (Huerta et al. 2017).

42 However, studies from the aquatic environment indicated that the external adsorption of
43 microplastics could cause harm to algal species in water mainly due to the microplastics
44 inhibiting the process of photosynthesis (Bhattacharya et al., 2010; Wright et al., 2013). Some
45 studies found that microplastics are ingested by aquatic fauna, leading to their death and/or
46 accumulate in the food chain possibly resulting in death (Hirai et al., 2011; (Hidalgo-Ruz et
47 al., 2012), Lwanga et al., 2016; Van Cauwenberghe et al., 2015b; Welden and Cowie, 2016).
48 Studies have also shown that organic pollutants can be attracted to microplastics (Brady and
49 Weil, 2000; Frias et al., 2010; Gall and Thompson, 2015; Hirai et al., 2011; Leslie, 2014;
50 Seltenrich; Yan et al.). In some cases, diethyl phthalate is added to plastic mulch during the
51 production process (Zhang et al., 2014b), and thus the use of this mulch introduces potentially
52 toxic substances directly into the environment.

53 It is a challenge to study the fate and effects of microplastics in the terrestrial environment,
54 especially in agricultural soils where plastic mulch is applied.

55 It is obvious how important it is to quantify microplastic content and distribution,
56 especially in agricultural soils where it could provide the crucial data needed for refining the

57 agricultural use of LDPE plastic mulch and cleaning up production processes (Leslie, 2014;
58 Steinmetz et al., 2016).

59 In recent years, several methods have been developed to detect microplastics in water,
60 focusing mainly on polyethylene (PE) and polypropylene Polymers (PP), and extract them
61 from sediments (Hidalgo-Ruz et al., 2012; Wang et al., 2016b). A floating method using
62 saturated solutions of NaCl, NaI, ZnCl₂, or sodium polytungstate (SPT) is usually used to
63 extract plastics from the sediment of water bodies due to the differences in density(Hidalgo-
64 Ruz et al., 2012; Imhof et al., 2012; Nuelle et al., 2014b). Microplastics are easily extracted
65 from sediment containing few impurities and few organic matter by aa floatation using a high
66 density solution (Hidalgo-Ruz et al., 2012; Wang et al., 2016b). However, microplastics could
67 be strongly adsorbed onto soil particles (Brady and Weil, 2000), thus making extraction via
68 flotation difficult. Low density PE and PP are widely used in agriculture to reduce
69 evaporation and increase soil temperature (Lwanga et al., 2016; Steinmetz et al., 2016; Yan et
70 al., 2015), and both, PE and PP densities are $<1 \text{ g cm}^{-3}$, thus saturated solutions of saline can
71 be replaced by distilled water (density is 1 g cm^{-3}) for agricultural soils. Generally, the particle
72 size and quantity of microplastics can be determined using a microscope (e.g. TM3030,
73 Hitachi, Japan)(Wang et al., 2016b) allowing microplastics to be classified as small
74 microplastics (SMP) measuring $<1 \text{ mm}$ or large microplastics (LMP) measuring 1 mm to 5
75 mm (Imhof et al., 2012; Van Cauwenberghe et al., 2015b). Microplastics with a diameter $>$
76 1mm are easily detected whereas microplastics with a diameter $< 1 \text{ mm}$ are difficult to
77 identify (Hidalgo-Ruz et al., 2012; Lenz et al., 2015). Mass determination is typically only
78 possible for larger quantities of bigger, more visible microplastics. Thus, most studies reveal
79 the total number of particles per volume of water or sediment (Hidalgo-Ruz et al., 2012).
80 Infrared spectroscopy and Raman micro-spectroscopy are considered to be the most reliable
81 yet most costly methods used to identify the properties of microplastics, but cannot identify
82 the mass content of microplastics in soils (Frias et al., 2010; Hirai et al., 2011;; Nor and
83 Obbard, 2014; Lenz et al., 2015; Wang et al., 2016b). Soils consist of many diverse
84 components which have a wide range of densities due to differing concentrations of soil
85 organic matter (SOM), sand, clay, silt and organic fibers. Some parts of SOM (density
86 ranging from 0.5 to 0.8 g cm^{-3}) (Barrios et al., 1996; Perie and Ouimet, 2008; Castanha et al.,
87 2012) are similar to microplastic density which leads to difficulties when trying to separate
88 microplastics from soil (Hidalgo-Ruz et al., 2012). Furthermore, the presence of some soil
89 components, such organic fibers and silicates, in the extracts makes the identification of

90 LDPE microplastics more difficult and a new detection/identification method should be
91 developed and verified (Hidalgo-Ruz et al., 2012; Steinmetz et al., 2016). Thus, the
92 identification of microplastics in soils needs to be developed and tested before assessing the
93 microplastic content of different soils The quantification of microplastics in soil is a challenge
94 and the analytical techniques required for this analysis still need to be innovated (Rillig, 2012;
95 Steinmetz et al., 2016).

96 In this study, we aimed to develop an accurate, simple, cost-saving method to assess small
97 microplastic particles of LDPE and PP measuring <1 mm in agricultural soils focusing on
98 four main goals: (1) extraction, (2) validation, and (3) identification of microplastic particles
99 in floatation separated from soil impurities (e.g. organic matter, organic fibers and silicates) as
100 well as (4) quantification of the number and size of particles. We used different soil types to
101 develop the method in hopes that the method could be widely used to determine the content
102 and the particle size distribution of microplastics in soils. This new method will be helpful in
103 evaluating the risks of microplastics in the soil, especially in farmland with long-term plastic
104 mulching use.

105 **2. Materials and Methods**

106 The characteristics of LDPE and PP were considered during the development of the
107 approach. Both LDPE and PP (Riblon, Ter Hell Plastic GmbH) were white and grounded into
108 irregularly-shaped particles by the company. The sizes of the LDPE particles were < 150 μm
109 and the PP particles were < 400 μm (Table 1, Figure 2). The densities for both kinds of plastic
110 particles were less than 1g cm^{-3} . In order to avoid contamination during the experiments, the
111 laboratory was thoroughly cleaned before the experiments were carried out and kept clean
112 throughout the duration of our testing. Clothes made from plastic fibres were not allowed in
113 the laboratory.

114 **2.1 Incubation of soil samples**

115 In this experiment, a floating method was used for extracting microplastic from different
116 soil samples. Soil types such as clay soil (15.8% clay, 64.6% silt, 19.6% sand, 3.23% OM
117 and pH(KCl) 7.84), loess soil (10.4% clay, 76.3% silt, 13.2% sand, 4.20% OM and pH(KCl)
118 7.16) and sandy soil (3.97% clay, 31.3% silt, 64.7% sand, 7.4% OM and pH(KCl) 6.63)
119 were used in this experiment. To homogenize, all soil samples were air-dried and sieved at 2
120 mm. LDPE and PP were added to soil samples at five concentration gradients (0.05%, 0.1%,

121 0.2%, 0.5% and 1.0%, w/w) with three replicates for each plastic. In each treatment, the
122 microplastic particles were mixed homogeneously with 10 g of soil which was adjusted to a
123 soil moisture content of 20%. The mixture was mixed in an aluminium cup (100 ml) and
124 sealed with plastic wrap. No debris dropping from the plastic wrap was verified before
125 experiment, and this was also proved in the control treatment table 5). In order to incorporate
126 the microplastics into the soil, the soil was incubated at 4°C for the duration of a week
127 (Hidalgo-Ruz et al., 2012; Paul and Clark, 1989; Reber, 1975).

128 2.2 Extraction of microplastic using a flotation method

129 The extraction procedures are illustrated in Figure 1A. Firstly, 50 ml of distilled water was
130 added to an aluminium cup containing the soil sample (10 g) and a glass stick was used to stir
131 the soil and water together manually in order to get a homogeneous suspension. Secondly, the
132 glass stick was rinsed off using distilled water and the water was collected in the original cup.
133 The cup was then refilled with water. Thirdly, the box was left over night in order to allow the
134 soil particles to deposit on the bottom of the cups. The floating SOM, microplastic and other
135 impurities (on the surface water) were poured off and filtered using filter paper (pore diameter
136 $< 3 \mu\text{m}$). The same procedure was carried out at least four times for each sample until no
137 materials were seen floating on the surface of the water or detected on the sides of the cups.
138 The soil solutions were subjected to two hours of ultrasonic vibrations and then filtered again
139 using the same filter paper. In order to reduce contamination during this process, all filter
140 papers were covered by a light aluminium specimen box during the process of filtration.

141

142 2.3 Validation of extraction

143 Filters with floatation were dried to constant weight at 60°C in an oven and then weighted
144 (W_1). After the floatation was brushed off, the filters were weighted again (W_2). The total
145 floatation (W_g) and the microplastics floatation (W_p) can be calculated by using the following
146 equations: $W_g = W_1 - W_2$, and $W_p = W_g - W_{\text{control}}$. W_{control} is the floatation collected from the
147 control treatment with no addition of microplastics. Recovery (R) was calculated by using the
148 following equation: $RR = W_p / W_{\text{added}}$. The W_{added} is the initial weight of the PE or PP added to
149 the soils.

150

151 2.4 Identification of microplastic under the microscope

152 The floatation included microplastic, organic matter, quartz, and fibres. The floatation was
153 dried on the filter at 60°C. The dry floatation was placed on a glass slide and evenly

154 distributed over the slide using a small brush in order to avoid any overlapping or clusters.
155 The bristles of the brush were tested before the sweeping procedure was performed and no
156 plastic debris from the bristles was dislodged from the brush. After putting the slide under the
157 microscope (Leica wild M3C, Type S, simple light) (6.4 X Zoom), a photo (Figure 2, before
158 heating) was taken using a high resolution camera (Leica DFC 425) linked to a computer with
159 image software (Leica Applicate Suite 4.8) in order to identify the number and size
160 distribution of the particles. The glass slide was then heated in an oven for 3-5 seconds at
161 130°C and a second photo (Figure 2, after heating) was taken. After heating, the microplastics
162 on the glass slide were transformed from irregular particles into circular, transparent, shiny
163 particles due to melting. Impurities such as organic matter, organic fibres and silicates did not
164 change by heating. Comparing the photos from pre- and post-heating, the LDPE/PP particles
165 can be clearly distinguished from the impurities and identified. Thus, the second photo was
166 used to distinguish the microplastic from organic matter, quartz and organic fibres while the
167 first photo was used to determine the number, sizes and shapes of the microplastic particles.
168 The identification procedures are shown in Figure 1 B.

169

170 2.5 Mass estimation of microplastics using an empirical model

171 Due to the great variance in the shapes of microplastics in natural soils, it is difficult to
172 quantify the volume of the microplastics. However, all of the microplastic particles in the soil
173 samples were transformed into circular bubble-like shapes after heating, so it was easy to
174 calculate the volume of the microplastics. In order to reduce any negative effects due to the
175 impurities (organic matter and organic fibres), pure LDPE and PP were used to develop the
176 models of microplastic weight and microplastic characteristics after heating. Six samples of
177 pure LDPE and PP were weighed randomly on glass slides and the microplastic were
178 scattered evenly across the slides. The weight of LDPE was 0.0010 g, 0.0017 g, 0.0031 g,
179 0.0015 g, 0.0011 g and 0.0020 g on the glass slides, respectively. The weight of PP was
180 0.0011 g, 0.0013 g, 0.0014 g, 0.0031 g, 0.0040 g and 0.0019 g on the glass slides, respectively.
181 The LDPE and PP on the glass slides were heated at 130 °C for 3-5 seconds and then photos
182 were taken under the microscope. The software of Image J (Schindelin et al., 2015) was used
183 to calculate the number of microplastic and the area they covered after heating under the
184 following conditions: (1) image type was 8bit; (2) colour and resolution of the image was
185 adjusted to make sure all particles were included; (3) the redundant small colour pixels
186 originating from glass slide and microscope were deleted. Heat actually changed the form of

187 microplastic so that microplastic volumes could be roughly calculated using hemisphere
188 measurements. The radius of each pellet was calculated using the vertical scanning area
189 (vertical angle of view) obtained by Image J (Figure S1). Regression analysis was carried out
190 on the following equations: the equations of weight (LDPE or PP) vs. counted number, weight
191 vs. area, and weight vs. volume, which were obtained after heating based on the area covered
192 by microplastics. Additionally, an empirical model was built to predict the mass of
193 microplastics in soils according to the best fitting model.

194

195 2.6 Validation of the identification and empirical model

196 The density of PP is similar to that of LDPE (Table 1) whereas the size range of PP is
197 wider than LDPE. Thus, PP was used for validation. Nine samples of pure PP were randomly
198 weighed and then each sample was added to 10 g of clay soil (3 replicates), sandy soil (3
199 replicates) and loess soil (3 replicates). 10 g and 3 replicates of clay soil, sandy soil and loess
200 soil were weighed. All of the samples were subjected to the steps of incubation, floatation,
201 identification and calculation as mentioned above. The weight of the PP samples were 0.0014
202 g, 0.0013 g and 0.0010 g for clay soil; 0.0010 g, 0.0016 g and 0.0015 g for sandy soil; and
203 0.0015 g, 0.0013 g and 0.0014 g for loess soil. Measurements of the size distribution of the PP,
204 the recovery of the microplastic ($W_{\text{model}}/W_{\text{added}}$) and the relative error (RE, $|W_{\text{model}} - W_{\text{added}}|$
205 $/ W_{\text{added}}$) of mass were all used to validate the methods of identification and the empirical
206 model. W_{model} was the mass calculated after heating using the experimental model.

207

208 2.6 Statistical analysis

209 Quantitative data of floatation collection, microplastic recovery, microplastic size,
210 microplastic number, and microplastic weight were described as mean \pm STD. Significant
211 differences ($p < 0.05$) were tested via ANOVA and pothoc test Dunnet T3, since the data
212 followed a normal distribution tested by the KS test. The single linear regression analysis of
213 weight (LDPE/PP) vs. floatation and weight (LDPE/PP) vs. particle properties after heating
214 were all carried out using SPSS 17 (SPSS Institute Inc., 2007). Vector line figures were
215 drawn using Sigmaplot 10.0.

216 3.5 A case study to validate the technique – microplastic in soils of northwest China

217 This new developed method was used to identify the microplastic distribution in soils
218 collected from the Loess plateau in Northwest China. Plastic mulch, containing mostly LDPE
219 with a density of $< 0.93 \text{ g cm}^{-3}$, has been used to reduce evaporation from agricultural fields
220 for the past 20 years. From this area, we collected a total of sixty soil samples from the
221 surface layer (0-10 cm, N=30) and a deeper layer (10-30 cm, N=30) in an agricultural field
222 (N=20), an orchard (N=20) and a greenhouse (N=20) in August 2016.

223 3. Results

224 3.1 Extraction

225 A process of three time floatation or 2 h ultrasonic was necessary to extract the
226 microplastics from the soil. A fourth flotation did not extract further microplastics, the water
227 surface remained clean. The weight of floatation (microplastic with impurities) followed the
228 following orden ($p < 0.05$) for all LDPE and PP concentrations (Tab. 2): sandy soil > loess >
229 clay > pure sand (Table 2). No microplastics were found in the control treatment.

230 3.2 Validation of extraction

231 The recovery rate based on the weight of LDPE ranged from $86.0\% \pm 0.8$ to $102.7\% \pm 4.2$ in
232 loess, , $103.0\% \pm 4.8$ to $128.0\% \pm 34.0$ in sandy soil, $89.9\% \pm 0.3$ to $104.0\% \pm 8.4$ in clay soil
233 and 87.9 ± 31.1 to 112.7 ± 22.0 in pure sand, respectively (Table 3).

234 For the PP, the recovery rate ranged from $97.5\% \pm 4.0$ to $126.7\% \pm 15.4$ in loess soil, from
235 $80.0\% \pm 38.0$ to $138.0\% \pm 39.0$ in sandy soil, from $97.0\% \pm 0.8$ to $121.3\% \pm 8.8$ in clay soil and
236 from $91.2\% \pm 10.6$ to $121.3\% \pm 27.7$ in pure sand, respectively. It was obvious that the standard
237 deviations were higher for the 0.05% concentration measurements than for the other
238 concentrations for both LDPE and PP, especially in sandy soil and pure sand.

239 Concerning the weight of microplastic, the regression curves were estimated between the
240 added weight of LDPE/PP and the recovered weight (Figure 3). The results showed that the
241 regression coefficients (R^2) were all over 98.5% and most of them were even close to 99.9%.

242 3.3 Empirical model developing for microplastic quantification

243 The relationship between the weight of added microplastic and particle numbers and the
244 area and volume after the heating treatment were regressed (Table 4). The results showed that
245 the relationship between particle number and weight, especially for PP, was relatively low

246 ($R^2=0.74$). However, there was a good relationship between the added LDPE/PP weight vs.
 247 area/volumes (after heating) when these were regressed ($R^2>0.99$, $p<0.01$). The real volumes
 248 of LDPE/PP can be calculated based on the weight of the LDPE/PP added and the real density
 249 of PE/PP (PP 0.92 g cm^{-3} , PE 0.91 g cm^{-3}). The volumes roughly calculated using hemisphere
 250 measurements were close to 4.5 times that of the actual volumes of PE/PP and were 1/9 of the
 251 volume of the spheres calculated using the radius of the vertical view area after heating by
 252 Image J. The equation of 1/9 volume of the sphere is shown below:

$$253 \quad V = \frac{4}{27} \sum_{i=1}^n \sqrt{\frac{S_i^3}{\pi}} \quad (1)$$

254 In the equation (1), V is the volume of the plastics after heating; n is the number counted,
 255 S is the vertical angle of the view area of the plastic after melting at $130 \text{ }^\circ\text{C}$ for 3-5s.

256 Many reports have indicated that most of the microplastic found in agricultural fields are
 257 PE and PP and their densities are close to 0.90 g cm^{-3} (Hidalgo-Ruz et al., 2012; Liu et al.,
 258 2014a; Steinmetz et al., 2016). Thus, the mass of microplastic in the field could be roughly
 259 calculated based on the model shown below:

$$260 \quad m = \frac{4}{27} \rho \sum_{i=1}^n \sqrt{\frac{S_i^3}{\pi}} \quad (2)$$

261 In the equation (2), m is the weight of plastics, ρ is the density of plastics (0.90 g cm^{-3}),
 262 and n is the number counted.

263 3.4 Validation of empirical model

264 The size distribution of PP detected after floatation and before heating using ImageJ were
 265 close to the original proportions, especially for the particles measuring $100\text{-}250 \text{ }\mu\text{m}$ and >250
 266 μm in all three soil types (Table 5). The mass of PP calculated using the empirical equation
 267 was also close to the original (recovery was 81.9-98.6%; RE, 9.8-18.1%) and even more
 268 accurate when the density of 9.2 g cm^{-3} was used in the calculations (recovery was 83.7-
 269 100.8%; RE, 9.3-16.3%) in the soil types.

270 3.5 Application of the technique to the case study

271 The results indicate that both, the number and the weight of microplastics were higher in
272 the deeper soil layers (10-30 cm) of the agricultural field, while they were lower in the deeper
273 soil layers (10-30 cm) of the orchard (Table 6). Most of microplastic sizes were $> 100 \mu\text{m}$ and
274 the weights were $< 0.54 \text{ mg kg}^{-1}$ in agricultural field, orchard field and green house.
275 Microplastics with a size $< 50 \mu\text{m}$ were found only in the deeper soil layers of the greenhouse
276 soil.

277 4. Discussion

278 4.1 Microplastic extraction from soil

279 Given the rapid increase in plastic production, contamination of the environment by small
280 plastic fragments, referred to as microplastics, has become a hot topic especially with relation
281 to aquatic organisms (Lusher et al., 2015; Van Cauwenberghe et al., 2015a) and water bodies
282 (Cózar et al., 2015; Seltnerich, 2015). Admittedly, microplastics found in water bodies and
283 solid medium, such as sediment from oceans or river beds, have been elutriated or floated,
284 either by pure water or saturated NaCl and NaI, attributing to the density differences between
285 microplastics (PE, PP and polyvinylchloride) and extraction solvents (Hidalgo-Ruz et al.,
286 2012; Nuelle et al., 2014b). “White pollution”, derived from plastic mulching use in
287 agriculture systems, has become the biggest issue for the surroundings and soil quality (Liu et
288 al., 2014b).

289 Most of the plastic mulching used for agriculture is made of LDPE and PP. Since the
290 density of this kind of plastic is typically $< 1 \text{ g cm}^{-3}$, distilled water can be used to float the
291 microplastic in the soil and can thus reduce clean-up costs as well as environment pollution
292 (Lwanga et al., 2016; Steinmetz et al., 2016; Yan et al., 2015). However, other high-density
293 solutions can also be used for microplastic extraction and flotation if the density of the
294 microplastics in the soils is unknown. The process would be same except for the fact that the
295 distilled water would be replaced by saturated solutions. Some studies have shown that the
296 microplastic in soils can be efficiently extracted using a methanol and dichloromethane
297 solution and can be weighted after evaporation to dryness under high temperatures (Fuller and
298 Gautam, 2016). However, these extraction methods could overestimate microplastic
299 concentration and contribute to the dissolution of soil organic matter at 180°C (Schnitzer and
300 Khan, 1998). In the present study, we developed a simple, low-cost method to extract
301 microplastic from soils using distilled water. Although impurities influenced microplastic

302 collection, high regression coefficients were observed between the purified floatation and the
303 microplastic that were initially added (Figure 3). It was obvious that most of the plastics and
304 impurities were floated before the fourth floatation (Figure S2). One previous report indicated
305 that ultrasonic waves could be beneficial when trying to separate particles from soils (Cooper
306 and Corcoran, 2010; Hidalgo-Ruz et al., 2012). In this study, soil samples were shaken and
307 exposed to ultrasonic waves for 2 hours after the fifth floatation but nothing was found on the
308 surface of water (Figure S2). However, in order to make sure all of the microplastic could be
309 floated on surface of the water, an ultrasonic step was also suggested after the last floating
310 procedure. At the time of the extraction, the highest deviation errors in the collection and
311 recoveries were observed for sandy soil (Table 3). This implies that the higher organic matter
312 and the fine particles in the soil samples strongly influenced the microplastic recovery (Brady
313 and Weil, 2000; Dekker and Ritsema, 1994; Zhang et al., 2014a). In addition, since the
314 extraction and floating procedures took a long time for each sample, centrifugal and vacuum
315 methods could be used for shorting the deposition and filtration times (Fossi et al., 2014;
316 Steinmetz et al., 2016; Thompson et al., 2004).

317 4.2 Identification of microplastic

318 Microplastic in the environment, especially in water bodies (salt or fresh water) and
319 sediment, can be detected, counted and weighed which can be attributed to either the large
320 volume of samples and less impurities or the larger particles (1mm-5mm) being examined
321 (Hidalgo-Ruz et al., 2012; Nuelle et al., 2014b; Wang et al., 2016b). However, soil
322 heterogeneity is complex, especially when considering all of the components such as mineral
323 soil, organic matter, chemical elements and plant/animal residues (Brady and Weil, 2000;
324 Zhang et al., 2014a; Zhang et al., 2016). In our study, not only microplastic were extracted
325 and floated, but organic matter and other residues were also floated in water which was
326 similar to other studies (Brady and Weil, 2000; Dekker and Ritsema, 1994; Hidalgo-Ruz et al.,
327 2012). All of the floatation needs to be identified, especially when distinguishing microplastic
328 from other impurities under a microscope. There are two ways to separate microplastic from
329 floatation. One way is to remove other impurities from the sample. In order to remove
330 impurities from the microplastic, H₂O₂ was used to remove organic matter and plant residues
331 in water sediments samples (Majewsky et al., 2016; Nuelle et al., 2014a; Wang et al., 2016a).
332 However, H₂O₂ not only removed the impurities from the sample, but it also digested some of
333 the microplastic (Claessens et al., 2013). Another way to remove impurities is to find some

334 advanced method to distinguish microplastic from impurities. In the present study, a heating
335 method was used to identify microplastic in all of the floatation samples by comparing the
336 shape change between pre- and post-heating. It was easy to distinguish between the impurities
337 and the microplastic, which showed transparent, circular, and shiny properties, or had a big
338 change of shapes, e.g. plastic fibre rolled up after heating(Figure2 and Figure S2). However,
339 two points need to be taken into account: heating temperature and time. The melting point of
340 LDPE and PP is 115-135 °C and 130-171 °C, respectively. If the temperature is too high or the
341 heating time is too long, the properties of the melted microplastic such as transparency,
342 circular form and shine would not be observed. In this study, we found that the microplastic
343 was transformed into transparent, circular and shiny areas after the floatation on the glass
344 slide was heated for 3 to 5 seconds at a temperature of 130 °C. The photo was then taken with
345 a camera connected to the microscope. Previous research indicated that microplastic could be
346 weighted after extraction but the limitation comes after trying to measure the size of the
347 microplastic in the samples(Schnitzer and Khan, 1998). With the help of the microscope and
348 Image J, particles size, shape and number could be determined using the photos taken before
349 and after heating. Therefore, the heating method is strongly recommended when trying to
350 identify microplastic collected from soil samples. Also, we found that the size distribution of
351 microplastic after extraction from three the soil types proved that this identification method
352 for microplastic is valid (Table 4). However, in this study, only the size of the
353 microplastic >20 µm could be clearly detected based on the resolution of the microscope and
354 the camera combination (16.4X). Concerning further testing, the limit of particle size
355 detection could be smaller if a digital camera with a higher resolution was connected to the
356 microscope. Furthermore, this method can't be used to distinguish the chemical components
357 of microplastic, which can be detected using the method of thermal analysis, infrared
358 spectroscopy or Raman micro-spectroscopy (Frias et al., 2010; Hirai et al., 2011; Lenz et al.,
359 2015; Majewsky et al., 2016; Nor and Obbard, 2014; Wang et al., 2016b).

360 4.3 Quantification of microplastic with models

361 For microplastic, only the total number of particles per given volume of water or sediment
362 were measured due to the small size and limited numbers of the microplastic and weighing
363 difficulties(Hidalgo-Ruz et al., 2012; Wang et al., 2016b). In this study, we developed some
364 models to predict the weight of microplastic in soils. After heating, the shape of the plastic
365 changed to the form of a circle when using the vertical view in 2 dimensions and to nearly a

366 hemisphere in the 3 -dimensional view. The regression between weight and volume was the
367 most accurate ($R^2 >99\%$) (Table 4), and the 1/9 volume of sphere calculated by the radius of
368 the vertical area after heating was close to the real volume of PE and PP. The empirical model
369 for mass prediction was build according to the volume and the accuracy of the model was
370 acceptable (recovery 82-99%) according to the validation using the steps of floatation by
371 distilled water, identification under a microscope after heating method, and quantification
372 based on an empirical model for three kinds of soil.

373 4.4 Case study: Microplastics in the soils of Northwest China

374 The plastic mulch used in fields in China are typically made of LDPE which has a density
375 of less than 0.93 g cm^{-3} (Yan et al., 2015). In this case study, distilled water was used to
376 extract microplastic from soils. The results indicated that the amount and weight of the
377 microplastic found were different between the soil from the agricultural field and the soil of
378 the orchard. This can be attributed to the fact that the microplastics deposited in the surface
379 layers were easy washed away by runoff in the agricultural field. There was less runoff and
380 soil loss in the orchard thus reducing the microplastic movement in the surface layer (Zhang
381 et al., 2015). Light, small-sized microplastic can be easy carried away by runoff and soil
382 erosion in agricultural fields and thus most of the larger microplastic found $>100 \mu\text{m}$
383 remained in the deep soil layers. The weight of the microplastic found in each of the three
384 land-use areas was less than 0.54 mg kg^{-1} . These results could be referenced in future research.
385 Furthermore, the case study also proved that this method, used in conjunction with the
386 floatation and heating techniques, can be used to investigate microplastic in soils.

387 5. Conclusions

388 The method of repetitive floatation can be used in extracting LDPE and PP microplastic
389 from soils with high accuracy. The number of microplastic can be easily distinguished from
390 impurities, and microplastic particles can be counted by comparing photos taken before and
391 after heating under a microscope. Microplastic mass evaluated by empirical model method
392 were acceptable.

393 Acknowledgements

394 Thanks to Robin Palmer for carefully editing the grammar and suggestions of this
395 manuscript. The paper was sponsored by the project of Interactive Soil Quality Assessment in

396 Europe and China for Agricultural Productivity and Environmental Resilience (ISQAPER:
397 635750), National Natural Science Foundation of China (41471228, 41771313), and China
398 Scholarship Council(201606615020).

399 **Uncategorized References**

- 400 Barrios, E., Buresh, R.J. and Sprent, J.I., 1996. Organic matter in soil particle size and density fractions from
401 maize and legume cropping systems. *Soil Biology & Biochemistry*, 28(2): 185-193.
- 402 Bhattacharya, P., Lin, S.J., Turner, J.P. and Ke, P.C., 2010. Physical Adsorption of Charged Plastic Nanoparticles
403 Affects Algal Photosynthesis. *Journal of Physical Chemistry C*, 114(39): 16556-16561.
- 404 Brady, N.C. and Weil, R.R., 2000. Nature and properties of soils. Macmillan publishing company, New York.
- 405 Castanha, C., Trumbore, S.E. and Amundson, R., 2012. Mineral and Organic Matter Characterization of Density
406 Fractions of Basalt- and Granite-Derived Soils in Montane California. An Introduction to the Study of
407 Mineralogy.
- 408 Claessens, M., Van Cauwenberghe, L., Vandegehuchte, M.B. and Janssen, C.R., 2013. New techniques for the
409 detection of microplastics in sediments and field collected organisms. *Marine Pollution Bulletin*, 70(1-2):
410 227-233.
- 411 Cooper, D.A. and Corcoran, P.L., 2010. Effects of mechanical and chemical processes on the degradation of
412 plastic beach debris on the island of Kauai, Hawaii. *Marine Pollution Bulletin*, 60(5): 650-654.
- 413 Cózar, A., Sanz-Martín, M., Martí, E., González-Gordillo, J.I., Ubeda, B., Ágálvez, J., Irigoien, X. and Duarte,
414 C.M., 2015. Plastic accumulation in the mediterranean sea. *PLoS ONE*, 10(4).
- 415 Dekker, L.W. and Ritsema, C.J., 1994. HOW WATER MOVES IN A WATER REPELLENT SANDY SOIL .1.
416 POTENTIAL AND ACTUAL WATER REPELLENCY. *Water Resources Research*, 30(9): 2507-2517.
- 417 Fossi, M.C., Coppola, D., Bainsi, M., Giannetti, M., Guerranti, C., Marsili, L., Panti, C., de Sabata, E. and Clo, S.,
418 2014. Large filter feeding marine organisms as indicators of microplastic in the pelagic environment:
419 The case studies of the Mediterranean basking shark (*Cetorhinus maximus*) and fin whale
420 (*Balaenoptera physalus*). *Marine Environmental Research*, 100: 17-24.
- 421 Frias, J., Sobral, P. and Ferreira, A.M., 2010. Organic pollutants in microplastics from two beaches of the
422 Portuguese coast. *Marine Pollution Bulletin*, 60(11): 1988-1992.
- 423 Fuller, S. and Gautam, A., 2016. A Procedure for Measuring Microplastics using Pressurized Fluid Extraction.
424 *Environmental Science & Technology*, 50(11): 5774-5780.
- 425 Gall, S.C. and Thompson, R.C., 2015. The impact of debris on marine life. *Marine Pollution Bulletin*, 92(1-2):
426 170-179.
- 427 Hidalgo-Ruz, V., Gutow, L., Thompson, R.C. and Thiel, M., 2012. Microplastics in the Marine Environment: A
428 Review of the Methods Used for Identification and Quantification. *Environmental Science & Technology*,
429 46(6): 3060-3075.
- 430 Hirai, H., Takada, H., Ogata, Y., Yamashita, R., Mizukawa, K., Saha, M., Kwan, C., Moore, C., Gray, H., Laursen,
431 D., Zettler, E.R., Farrington, J.W., Reddy, C.M., Peacock, E.E. and Ward, M.W., 2011. Organic
432 micropollutants in marine plastics debris from the open ocean and remote and urban beaches. *Marine
433 Pollution Bulletin*, 62(8): 1683-1692.
- 434 Imhof, H.K., Schmid, J., Niessner, R., Ivleva, N.P. and Laforsch, C., 2012. A novel, highly efficient method for
435 the separation and quantification of plastic particles in sediments of aquatic environments. *Limnology
436 and Oceanography-Methods*, 10: 524-537.
- 437 Lenz, R., Enders, K., Stedmon, C.A., Mackenzie, D.M.A. and Nielsen, T.G., 2015. A critical assessment of visual
438 identification of marine microplastic using Raman spectroscopy for analysis improvement. *Marine
439 Pollution Bulletin*, 100(1): 82-91.
- 440 Leslie, H., 2014. Review of Microplastics in Cosmetics. 14/29, IVM Institute for Environmental Studies,
441 Amsterdam.
- 442 Liu, E.K., He, W.Q. and Yan, C.R., 2014a. 'White revolution' to 'white pollution'-agricultural plastic film mulch in
443 China. *Environmental Research Letters*, 9(9).
- 444 Liu, E.K., He, W.Q. and Yan, C.R., 2014b. 'White revolution' to 'white pollution'—agricultural plastic film mulch
445 in China. *Environmental Research Letters*, 9(9): 091001.
- 446 Lusher, A.L., Hernandez-Milian, G., O'Brien, J., Berrow, S., O'Connor, I. and Officer, R., 2015. Microplastic and
447 macroplastic ingestion by a deep diving, oceanic cetacean: The True's beaked whale *Mesoplodon mirus*.
448 *Environmental Pollution*, 199: 185-191.
- 449 Lwanga, E.H., Gertsen, H., Gooren, H., Peters, P., Salanki, T., van der Ploeg, M., Besseling, E., Koelmans, A.A.
450 and Geissen, V., 2016. Microplastics in the Terrestrial Ecosystem: Implications for *Lumbricus terrestris*
451 (*Oligochaeta*, *Lumbricidae*). *Environmental Science & Technology*, 50(5): 2685-2691.
- 452 Majewsky, M., Bitter, H., Eiche, E. and Horn, H., 2016. Determination of microplastic polyethylene (PE) and
453 polypropylene (PP) in environmental samples using thermal analysis (TGA-DSC). *Science of the Total
454 Environment*, 568: 507-511.
- 455 Nor, N.H.M. and Obbard, J.P., 2014. Microplastics in Singapore's coastal mangrove ecosystems. *Marine
456 Pollution Bulletin*, 79(1-2): 278-283.
- 457 Nuelle, M.-T., Dekiff, J.H., Remy, D. and Fries, E., 2014a. A new analytical approach for monitoring
458 microplastics in marine sediments. *Environmental Pollution*, 184: 161-169.
- 459 Nuelle, M.T., Dekiff, J.H., Remy, D. and Fries, E., 2014b. A new analytical approach for monitoring microplastics
460 in marine sediments. *Environmental Pollution*, 184: 161-169.

- 461 Paul, E.A. and Clark, F.E., 1989. Soil microbiology and biochemistry. Academic Press, San Diego.
- 462 Perie, C. and Ouimet, R., 2008. Organic carbon, organic matter and bulk density relationships in boreal forest
463 soils. Canadian Journal of Soil Science, 88(3): 315-325.
- 464 Reber, H., 1975. Soil microbiology: a critical review. Butterworth & Co. (Publishers) Ltd, London, New York.
- 465 Rillig, M.C., 2012. Microplastic in Terrestrial Ecosystems and the Soil? Environmental Science & Technology,
466 46(12): 6453-6454.
- 467 Rocha-Santos, T. and Duarte, A.C., 2015. A critical overview of the analytical approaches to the occurrence, the
468 fate and the behavior of microplastics in the environment. Trac-Trends in Analytical Chemistry, 65: 47-
469 53.
- 470 Schindelin, J., Rueden, C.T., Hiner, M.C. and Eliceiri, K.W., 2015. The ImageJ ecosystem: An open platform for
471 biomedical image analysis. Molecular Reproduction and Development, 82(7-8): 518-529.
- 472 Schnitzer, M. and Khan, S.U., 1998. Soil organic matter. Elsevier science publishing company INC, New York.
- 473 Seltenrich, N., New link in the food chain? Marine plastic pollution and seafood safety. Environ. Health Perspect,
474 123(2): A34-A41.
- 475 Seltenrich, N., 2015. New link in the food chain? Marine plastic pollution and seafood safety. Environmental
476 Health Perspectives, 123(2): A34-A41.
- 477 Steinmetz, Z., Wollmann, C., Schaefer, M., Buchmann, C., David, J., Troeger, J., Munoz, K., Fror, O. and
478 Schaumann, G.E., 2016. Plastic mulching in agriculture. Trading short-term agronomic benefits for
479 long-term soil degradation? Science of the Total Environment, 550: 690-705.
- 480 Thompson, R.C., Olsen, Y., Mitchell, R.P., Davis, A., Rowland, S.J., John, A.W.G., McGonigle, D. and Russell,
481 A.E., 2004. Lost at sea: Where is all the plastic? Science, 304(5672): 838-838.
- 482 Van Cauwenberghe, L., Claessens, M., Vandegehuchte, M.B. and Janssen, C.R., 2015a. Microplastics are taken
483 up by mussels (*Mytilus edulis*) and lugworms (*Arenicola marina*) living in natural habitats.
484 Environmental Pollution, 199: 10-17.
- 485 Van Cauwenberghe, L., Devriese, L., Galgani, F., Robbens, J. and Janssen, C.R., 2015b. Microplastics in
486 sediments: A review of techniques, occurrence and effects. Marine Environmental Research, 111: 5-17.
- 487 Wang, W., Ndungu, A.W., Li, Z. and Wang, J., 2016a. Microplastics pollution in inland freshwaters of China: A
488 case study in urban surface waters of Wuhan, China. Science of The Total Environment.
- 489 Wang, W.F., Ndungu, A.W., Li, Z. and Wang, J., 2016b. Microplastics pollution in inland freshwaters of China: A
490 case study in urban surface waters of Wuhan, China. Science of the Total Environment.
- 491 Welden, N.A.C. and Cowie, P.R., 2016. Long-term microplastic retention causes reduced body condition in the
492 langoustine, *Nephrops norvegicus*. Environmental Pollution, 218: 895-900.
- 493 Wright, S.L., Thompson, R.C. and Galloway, T.S., 2013. The physical impacts of microplastics on marine
494 organisms: A review. Environmental Pollution, 178: 483-492.
- 495 Yan, C.Y., He, W.Q., Liu, S. and Cao, S.L., 2015. Application of mulch films and prevention of its residual
496 pollution in China. Science press, Beijing.
- 497 Zhang, S.L., Huffman, T., Zhang, X.Y., Liu, W. and Liu, Z.H., 2014a. Spatial distribution of soil nutrient at depth
498 in black soil of Northeast China: a case study of soil available phosphorus and total phosphorus.
499 Journal of Soils and Sediments, 14(11): 1775-1789.
- 500 Zhang, S.L., Yan, L.L., Huang, J., Mu, L.L., Huang, Y.Q., Zhang, X.Y. and Sun, Y.K., 2016. Spatial
501 Heterogeneity of Soil C:N Ratio in a Mollisol Watershed of Northeast China. Land Degradation &
502 Development, 27(2): 295-304.
- 503 Zhang, Y., Wang, L., Du, N., Ma, G.P., Yang, A.M., Zhang, H., Wang, Z.G. and Song, Q.X., 2014b. Effects of
504 diethylphthalate and di-(2-ethyl)hexylphthalate on the physiology and ultrastructure of cucumber
505 seedlings. Environmental Science and Pollution Research, 21(2): 1020-1028.
- 506 Zhang, Z.Y., Sheng, L.T., Yang, J., Chen, X.A., Kong, L.L. and Wagan, B., 2015. Effects of Land Use and Slope
507 Gradient on Soil Erosion in a Red Soil Hilly Watershed of Southern China. Sustainability, 7(10): 14309-
508 14325.

509

Table 1 Size distribution of the original microplastics of PE and PP

	Density (g cm ⁻³)	Components of Particle sizes			
Plastic Diameter		<50µm	50-100µm	100-150 µm	
PE	0.92	50%	22%	23%	
Plastic Diameter		<50µm	50-100µm	100-250µm	>250µm
PP	0.91	0.2%	1.5%	38.1%	58.4%

Table 2 The floatation collection by distilled water after LDPE and PP added, and the floatation collection by distilled water after PP added for a week.

	PE/PP added(mg)	Concentration (% w/w)	Floatation collection (mg) (Mean±STD)			
			Loess	Sandy	Clay	Pure sand
LDPE	0	0.00	39.6±1.0b	121.7±14.2a	9.3±0.9c	0d
	5.0	0.05	44.7±1.8b	128.1±18.9a	14.6±1.2c	4.4±1.6c
	10.0	0.10	48.2±0.5b	133.7±30.4a	18.4±1.3c	11.3±2.2c
	20.0	0.20	62.2±3.1b	146.8±3.3a	29.0±2.0c	20.8±0.9d
	50.0	0.50	86.0±8.5b	173.2±11.8a	54.3±0.2c	50.5± 0.8c
	100.0	1.00	131.9±5.2b	244.7±3.0a	112.4±11.3c	89.6±9.4d
PP	0	0.00	39.6±1.0b	121.7±14.2a	9.3±0.9c	0d
	5.0	0.05	45.2 ±8.3b	125.7±16.1a	15.1±2.9c	6.1±1.4d
	10.0	0.10	52.2±6.5b	135.5±4.3a	19.1±0.2c	11.0±1.2d
	20.0	0.20	59.1±2.4b	139.8±14.7a	33.6±2.4c	19.7±1.5d
	50.0	0.50	89.5±1.3b	173.8±5.0a	67.0± 5.7c	45.6±5.3d
	100.0	1.00	141.8±3.7b	238.4± 22.7a	110.5± 5.8c	106.7± 9.0c

Values followed by the same letter within the same rows are not significantly different by LSD's multiple range test ($P \leq 0.05$).

Table 3 The recovery rate of LDPE and PP in different treatments

	Concentration (% w/w)	Recovery (%) (Mean±STD)			
		Loess	Sandy	Clay	Pure sand
LDPE	0.05	102.7±4.2a	128.0±34.0a	104.0±8.4a	87.9±31.1a
	0.10	86.0±0.8a	120.0±16.0a	90.0±6.5a	112.7±22.0a
	0.20	113.3±5.7a	125.5±54.5a	98.3±6.9a	103.8±4.7a
	0.50	93.9±9.2a	103.0±4.8a	89.9±0.3a	101.0± 1.6a
	1.00	92.3±3.6a	123.0±11.2a	103.0±10.4a	89.6±9.4a
PP	0.05	112.0 ±20.5a	80.0±38.0a	114.7±22.1a	121.3±27.7a
	0.10	126.7±15.4a	138.0± 39.0a	97.0±0.8a	110.0±11.7a
	0.20	97.5±4.0a	90.5±2.5a	121.3±8.8a	98.5±7.4a
	0.50	99.9±1.5a	104.2± 18.4a	115.3± 9.7a	91.2±10.6a
	1.00	102.3± 2.6a	116.7± 8.4a	101.2± 5.3a	106.7± 9.0a

Values followed by the same letter within the same rows are not significantly different by LSD's multiple range test ($P \leq 0.05$).

Table 4 The regression of PE and PP between weight and number, weight and area, and weight and volume.

	Regression parameter	Equations*	R²	P
PE	Number vs. weight (mg)	$y = 0.001x - 0.35$	0.89	<0.01
	Area (mm ²) vs. weight (mg)	$y = 0.05x + 0.89$	0.97	<0.01
	Volume (mm ³) vs. weight (mg)	$y = 1.14x + 0.46$	0.99	<0.01
PP	Number vs. weight (mg)	$y = 0.01x - 0.74$	0.74	<0.01
	Area (mm ²) vs. weight (mg)	$y = 0.08x - 0.53$	0.98	<0.01
	Volume (mm ³) vs. weight (mg)	$y = 0.67x + 0.17$	0.99	<0.01

*.Note: “y” is the weight of PE/PP. Area of PE/PP is the area from vertical angle of view under microscope after melt. Volume of PE/PP was calculated by equation (1).

Table 5 Validation of identification and experimental model for MPs

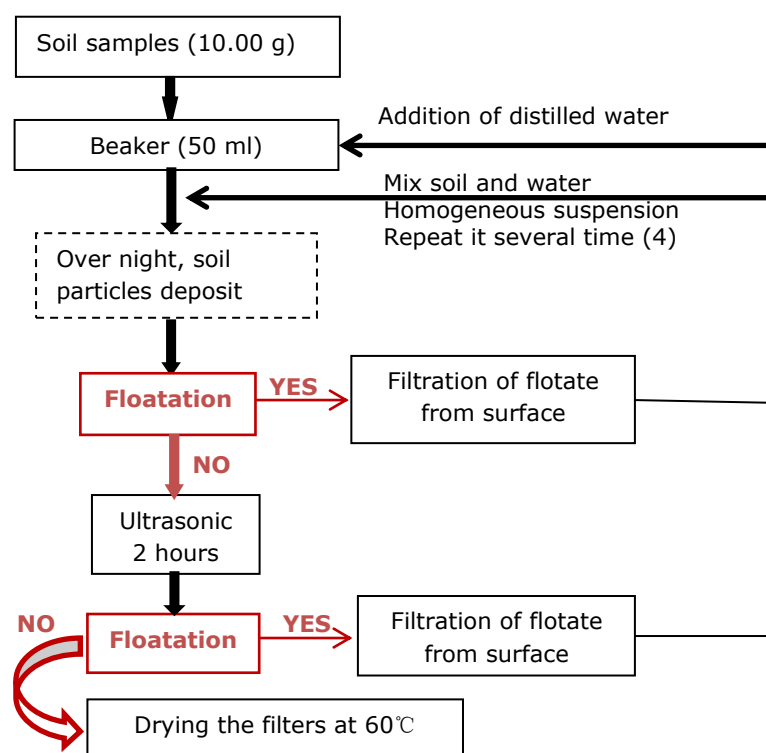
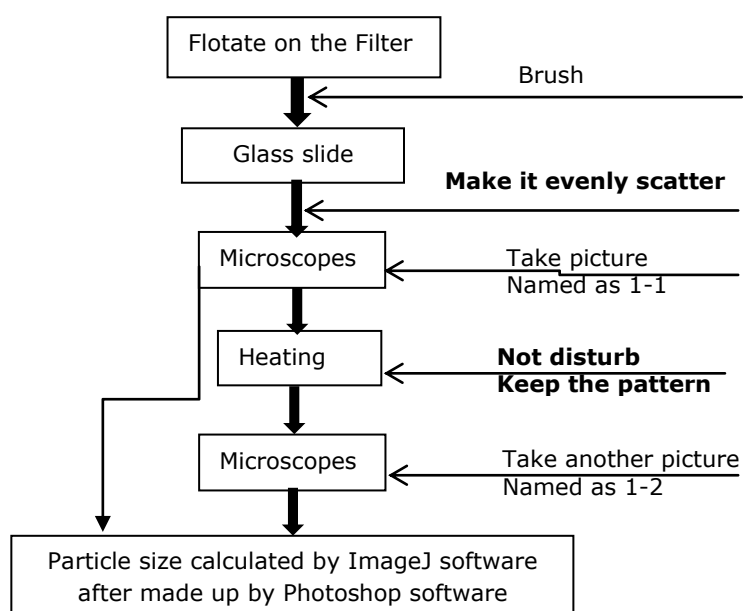
Soil types	N	PP size (%) (Mean±STD)			Recovery (%) (Mean±STD)		Mean RE of mass (Mean±STD)	
		<100 (nm)	100- 250(nm)	>250(nm)	Model $\rho=0.90 \text{ g cm}^{-3}$	Model $\rho=0.92 \text{ g cm}^{-3}$	$\rho=0.90 \text{ g cm}^{-3}$	$\rho=0.92 \text{ g cm}^{-3}$
Clay	3	0	0	0	0	0	0	
Loess	3	0	0	0	0	0	0	
Sandy	3	0	0	0	0	0	0	
Clay	3	5.9±2.4	34.8±7.1	59.3±6.2	87.5±6.4	89.4±8.2	12.8±7.2%	10.7±7.7%
Loess	3	7.1±7.0	34.8±3.3	58.2±9.9	81.9±8.9	83.7±9.1	18.1±8.9%	16.3±9.0%
Sandy	3	9.0±3.3	35.9±4.4	55.1±6.7	98.6±12.8	100.8±13.0	9.8±4.6%	9.3±6.5%

Table 6 Microplastics in soils of northwest China

Land use	N ₁	Soil depth (cm)	Numbers (n kg ⁻¹)			Mean±STD	CV	Content (mg kg ⁻¹)	
			<50(μm)	50-100(μm)	>100(μm)			Mean±STD	CV ₂
Agricultural field	10	0-10 cm	0	0	40	40±126	3.16	0.008±0.025	3.16
	10	10-30 cm	0	0	100	100±141	1.41	0.368±0.740	2.01
Fruit field	10	0-10 cm	0	40	280	320±329	1.03	0.540±0.603	1.12
	10	10-30 cm	0	20	100	120±169	1.41	0.460±0.735	1.60
Green house field	10	0-10 cm	0	20	80	100±254	2.54	0.130±0.307	2.84
	10	10-30 cm	40	0	40	80±193	2.42	0.024±0.051	2.11

1. “N” is the numbers of soil samples

2. “CV” is the variable coefficient

A: Extraction of MPs by flotation method**B: Identification of MP under microscope****Figure 1** Framework of extraction and identification

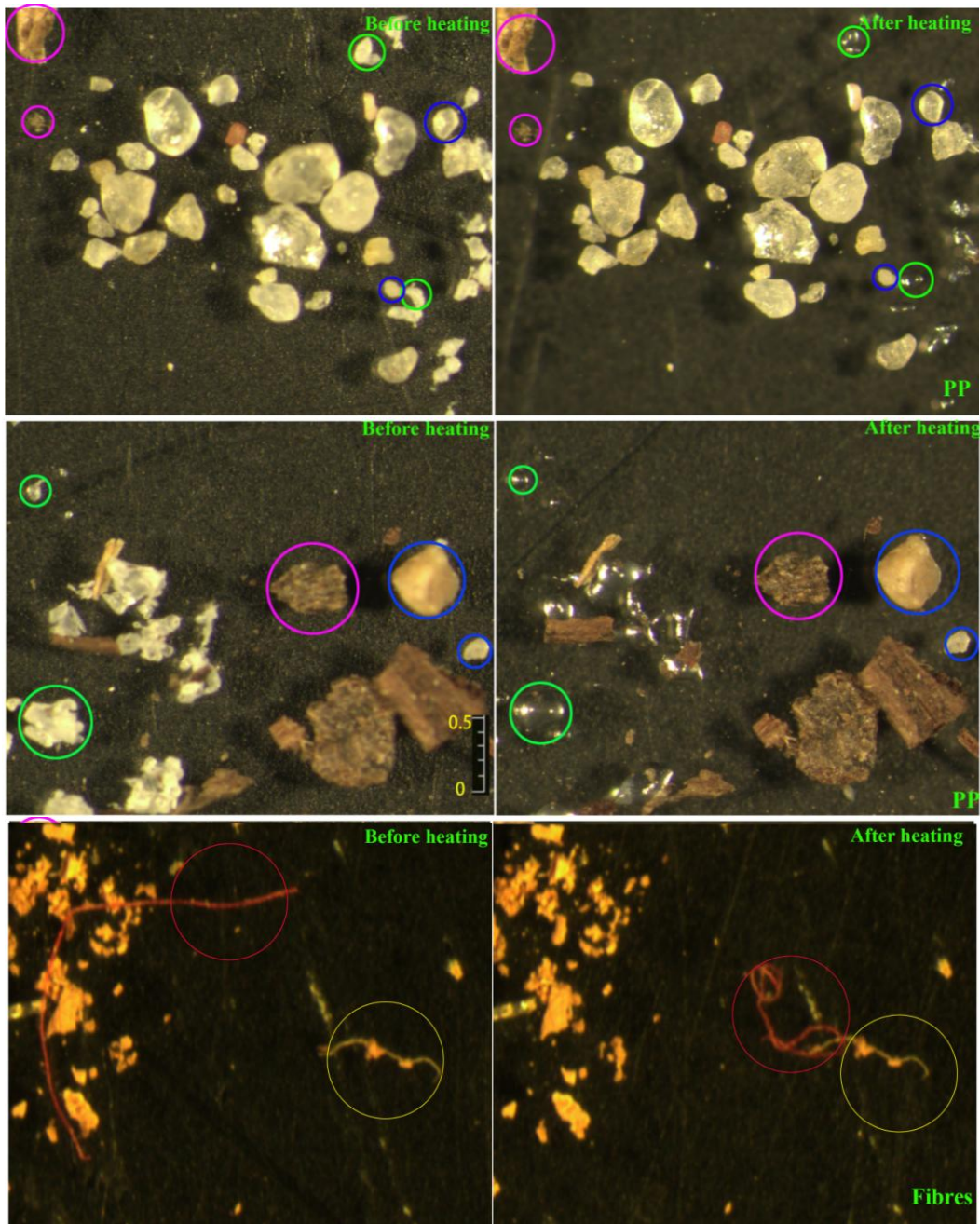


Figure 2 Definition of MPs from photos taken by camera (DFC 425) under microscope (6.4×).

A selection of particles are marked to show the shapes of MP changed to circle and transparent after heating, while quartzs and organic matters were not changed. Plastics, sands and organic matters were circled by green colour, blue colour and purple colour, respectively. Wood fibre was not changed, while plastics fibre was rolled up. Wood fibre and plastics fibre were circled by yellow colour, red colour. MPs are easy connected after melt when the distance between particles are less than 10 μm .

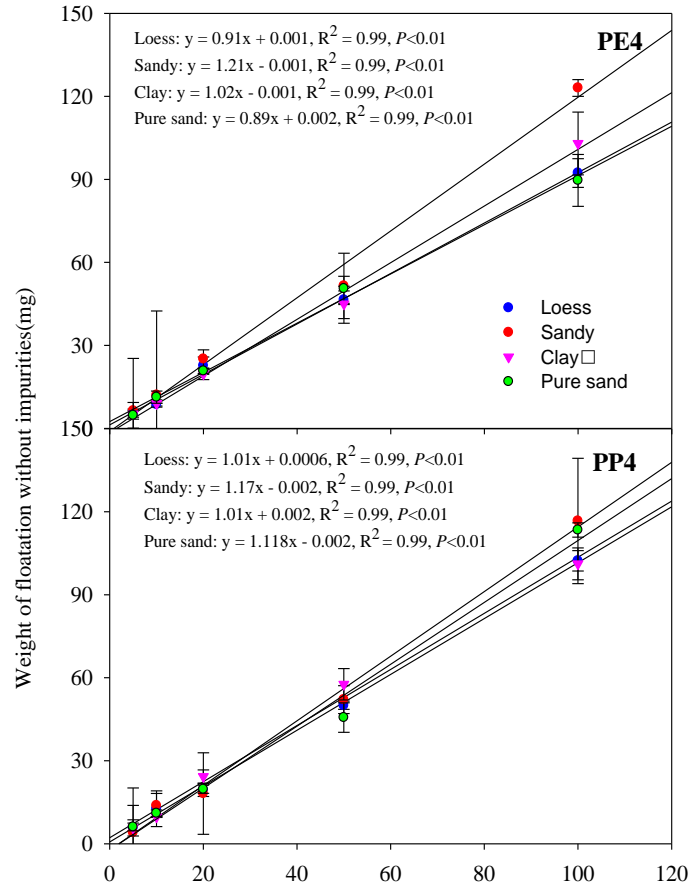


Figure 3 Recovery rate of incubation at 4°C (PE4 and PP4)

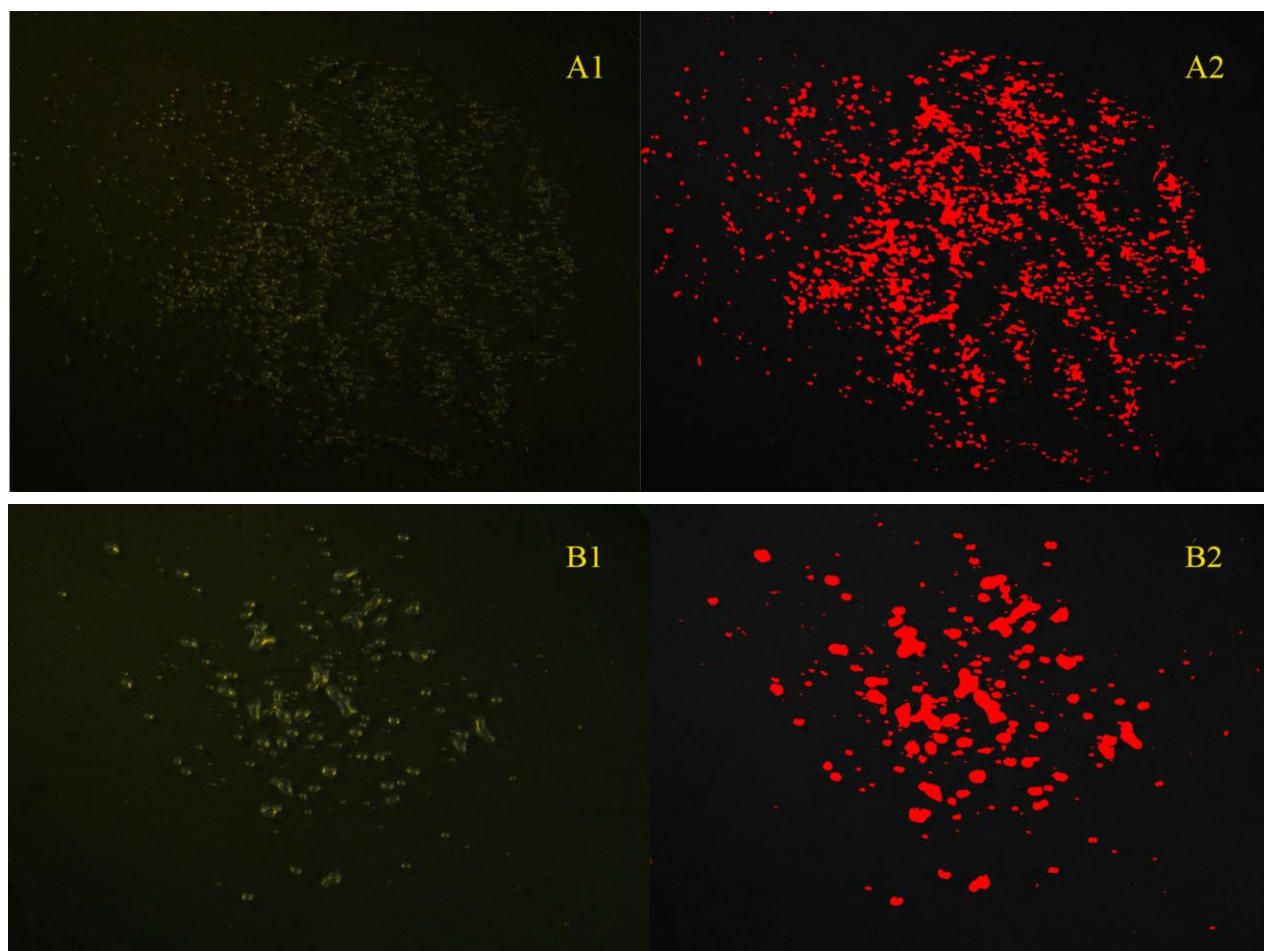


Figure S1 Photos of melted PE (A) and PP (B) under microscope (100×), and the numbers and shapes (A2 and B2) were calculated by Image J software.



Figure S2 No MPs or impurities floated on after four times' extraction by distilled water, then moved into the ultrasonic equipment, but MP was also not found on water surface under 2 hours' ultrasonic .

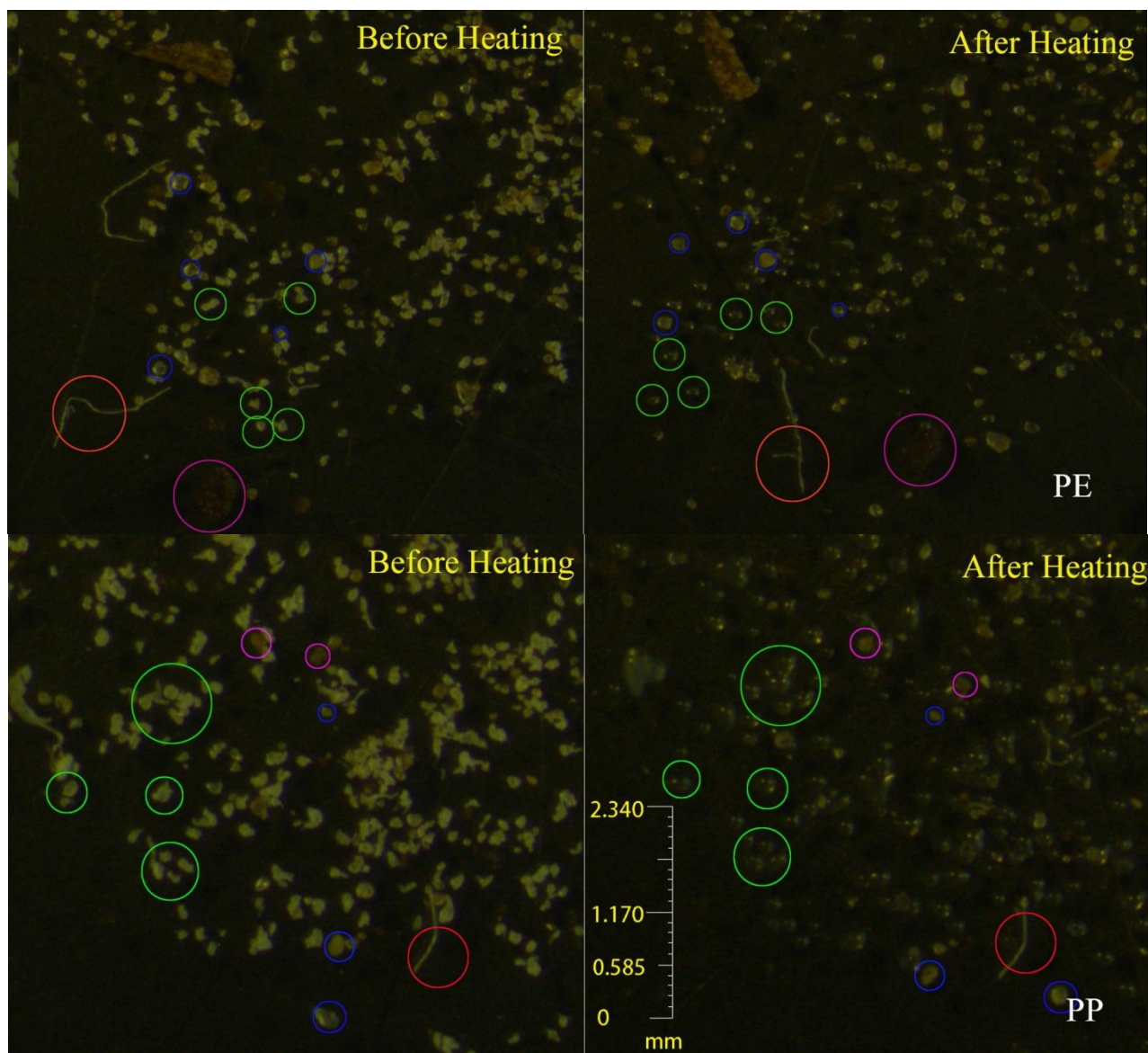
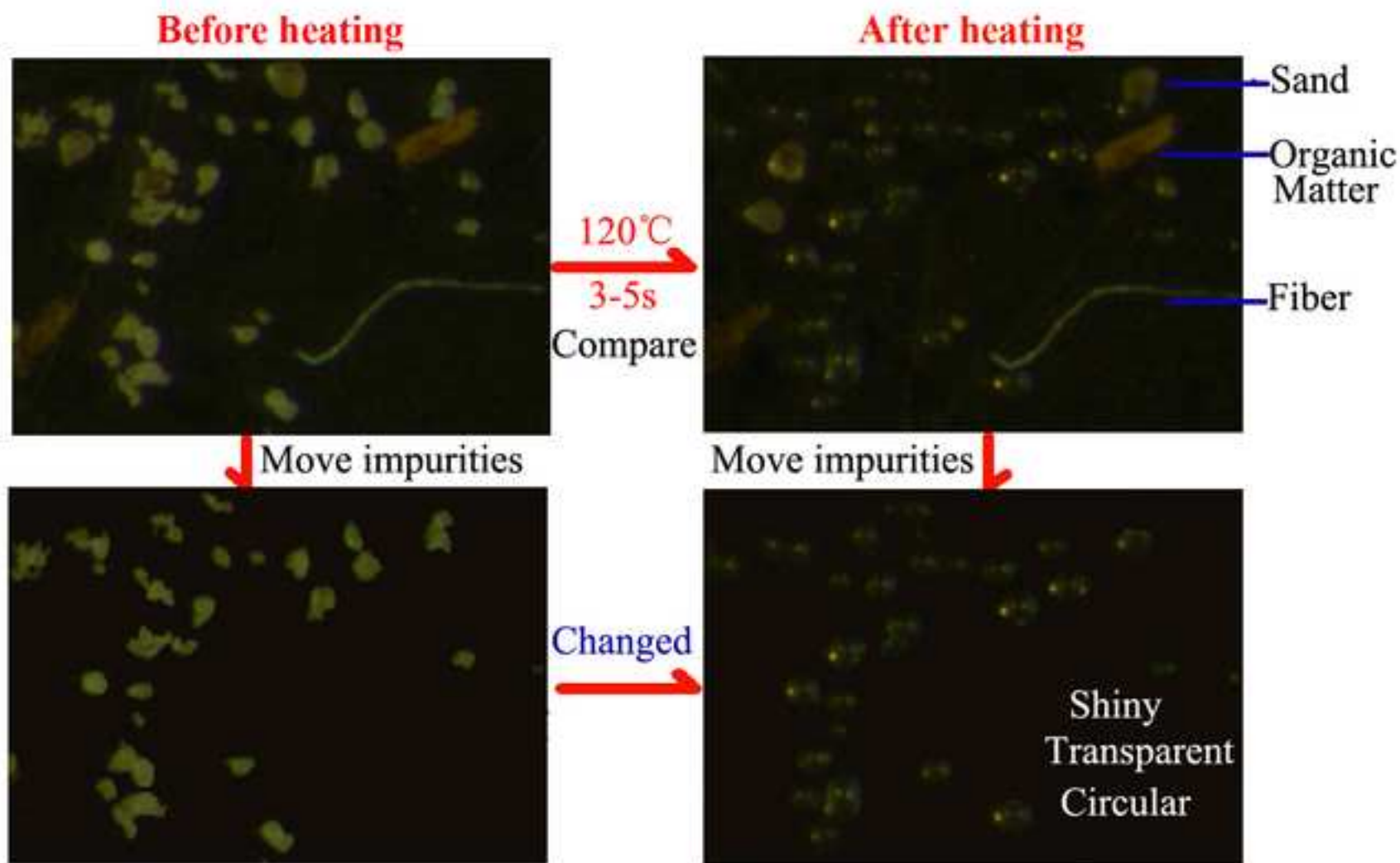


Figure S3 Identification of MPs under microscope (100 \times). A selection of particles were marked to show how the shapes of MPs changed to transparent circular forms after heating, while quartz, organic matter and wood fibres were not changed. Plastics, sands, organic matter and wood fibres were circled using green, blue, purple and red colours, respectively. MPs were easy connected together after melting when the distance between particles was less than 10 μm . The numbers were 186 and 155 in the pictures of PE and PP, respectively.



Highlights

- A low cost-saving method were developed to extract, distinguish and quantify MP from soils.
- Flootation method was efficient in extracting MP from soils.
- After heating, MPs were transformed into circular, transparent and shiny particles.
- Regression analysis of MP weight and particle volume after heating showed the best fit.