

Conversion of organic resources by black soldier fly larvae: Legislation, efficiency and environmental impact

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22 Abstract

23 To meet the projected substantial growth in the global demand for meat, we are challenged to 24 develop additional protein-rich feed ingredients while minimizing the use of natural 25 resources. The larvae of the black soldier fly (BSF) have the capacity to convert low-value 26 organic resources into a high quality protein source for pigs, chickens and fish and as such 27 may increase both the productivity and the efficiency of the food chain. The aim of this study 28 was to assess the environmental opportunities of BSF larvae reared on different sources using 29 up to date literature data on the efficiency of BSF larvae in converting such resources into 30 biomass. The current EU legislative framework was used to classify the various resources for 31 rearing insects. Data of forty articles published until 1 September 2017 were used, reporting 32 on in total 78 (mixtures of) resources used for growing BSF larvae. Data on the resource 33 conversion efficiency on dry matter (DM) and N basis was presented in 11 and 5 studies, 34 evaluating 21 and 13 resources, respectively. Resources studied included food and feed 35 materials (A, n=8 resources), foods not intended (anymore) for human consumption (B1, 36 n=4), and residual streams such as food waste (D, n=2), and animal manure (E, n=7). 37 Conversion efficiency varied from 1.3 to 32.8% for DM and from 7.4 to 74.8% for N. Using 38 life cycle assessment, our environmental results showed that resources within the legal groups 39 (i.e. A and B1) that are, at the moment, not allowed in EU as animal feed have in general a 40 lower impact in terms of global warming potential, energy use, and land use. On a per kg 41 protein basis, BSF larvae reared on a resource that contains food (e.g. sorghum) and feed (e.g. 42 dried distillers grains with solubles) products generally have higher environmental impacts 43 than conventional feed protein sources (fishmeal and soybean meal). Using insects as feed, 44 therefore, has potential to lower the environmental impact of food production but a careful 45 examination of the resource is needed in terms of environmental impact, safety and 46 economics.

47 **Key words:** *Hermetia illucens*; insects; life cycle assessment; resource use efficiency

- 49 Abbreviations: BSF, black soldier fly; DM, dry matter; EFSA, European Food Safety
- 50 Authority; EU, European Union; GHG, greenhouse gas; GWP, global warming potential;
- 51 LCA, life cycle assessment; N, nitrogen; PAPs, processed animal proteins.

52 **1 Introduction**

53 For assuring food security within the planet's carrying capacity, new ways are required to 54 increase protein production while minimizing the use of natural resources (Godfray et al., 55 2010; Foley et al., 2011). As the demand for meat is projected to grow with 76% (2005/2007-56 2050, Alexandratos and Bruinsma, 2012), there is in particular a need to find additional 57 protein-rich feed ingredients as well as alternatives for those associated with a high 58 environmental impact such as soybean meal (Veldkamp et al., 2012; van Huis et al., 2013). 59 Insects have been proposed to increase both the productivity and the efficiency of the food 60 chain (van Huis et al., 2013). Research on using insects as feed is rapidly evolving and several 61 reviews have recently been published on their nutritional value, potential organic resources, 62 and food safety (Veldkamp et al., 2012; van Huis, 2013; Barroso et al., 2014; Makkar et al., 2014; Pastor et al., 2015; Barragan-Fonseca et al., 2017; Testa et al., 2017; Varelas and 63 64 Langton, 2017; van der Fels-Klerx et al., 2018). In particular the larvae of the black soldier fly (*Hermetia illucens*, BSF¹) receive considerable interest as these have the ability to upcycle 65 various residual organic resources (Pastor et al., 2015) into protein-rich biomass fit as feed 66 ingredients for pigs, chickens and fish (e.g. Newton et al., 1977; Bondari and Sheppard, 1981; 67 68 De Marco et al., 2015). 69 From an environmental viewpoint, only a few studies quantified the impact of BSF larvae

(Smetana et al., 2016; Salomone et al., 2017). Current literature showed that the resource used to rear BSF larvae affects the environmental impact: BSF larvae fed with cattle manure and municipal waste seem to have a relatively lower environmental impact than those fed with e.g. beet pulp (Smetana et al., 2016). It is, however, unclear which resources or groups of resources have potential to reduce the environmental impact and how this relates to the legal status of using those biomass streams as a resource to feed larvae. Although in the European

¹ Abbreviations: BSF, black soldier fly; DM, dry matter; EC, European Parliament and Council; EFSA, European Food Safety Authority; EU, European Union; GHG, greenhouse gas; GWP, global warming potential.

76 Union (EU) it is currently not allowed to use insects as feed that are fed on resources 77 containing manure or waste, it is important to understand the potential of BSF larvae for 78 improving the productivity and resource use efficiency of our food system. Furthermore, 79 resources under study may already have applications as livestock feed ingredients (e.g. beet 80 pulp), which underlines the need for a clear differentiation and categorisation of resources for 81 applications within (i.e. as food or feed) and outside the food chain. As the choice of organic 82 resources for BSF larvae production is crucial for the economics, environmental footprint and 83 safety of the products, the research on the suitability of organic resources for BSF larvae 84 production has been rapidly growing over the years. Resources may differ greatly in their 85 impact on larval development time, biomass yield and quality, associated emissions and 86 residual matter (frass and exuvia). More data are becoming available on how efficient BSF 87 larvae actually convert the nitrogen (N) from resources into nitrogenous biomass. These data 88 allow more extensive assessments of the environmental impact of BSF larvae as an alternative 89 protein-rich feed ingredient. The aim of this study was to assess the environmental 90 opportunities of insects reared on different organic biomass resources and relate their 91 potential to the current EU legislation framework. We used a cross-disciplinary approach to 92 cover areas of law, animal sciences and environmental sciences simultaneously to critically 93 assess our current understanding of the concept of using BSF larvae to make our food system 94 more productive and sustainable.

95

96 **2 Method**

97 To assess the environmental opportunities of different organic resources and relate their
98 potential to the current legislation framework, we i) classified organic resources according to
99 EU legislation (2.1); ii) analysed literature data on organic resources used as feed for BSF

100 larvae and, where possible, calculated the feed conversion efficiency (2.2); iii) assessed the

101 environmental impact of BSF larvae reared on these organic resources (2.3).

102

103 2.1 Legal classification of organic resources and safety

104 In the EU, insects reared for food or feed fall under the definition of 'farmed animal' (Article 105 3.6 of Regulation (EC) No 1069/2009), which has certain consequences for the permission to 106 use a feed (organic resource or substrate) for a farmed animal. General rules for all feed in the 107 EU, including that for insects, are that it has to be (a) safe, and (b) it does not have a direct 108 adverse effect on the environment or animal welfare (Article 4 Regulation (EC) No 767/2009 109 and Article 15 of Regulation (EC) No 178/2002). In addition, there are requirements for feed 110 hygiene (Regulation (EC) No 183/2005) and the maximum contents of certain undesirable 111 substances in animal feed (Directive 2002/32/EC). We build on the demarcation between 112 insect feeding source options, as previously defined by the European Food Safety Authority 113 (EFSA, 2015), and used it for the classification of resources evaluated for insect rearing. The 114 legal status (allowed or not allowed) and justification for this status are presented in Table 1. 115 A more extensive description of the background of group of insect feeding source options can 116 be found in the supplementary material.

117

118 2.2 Literature review of bioconversion studies

There is a growing number of studies that focus on the use of BSF larvae to convert organic resources with purposes that relate to feed and biofuel production as well as waste management (Table S1). We performed a literature review to create an overview of organic resources used as feed and feed conversion efficiency of BSF larvae. Articles published in peer-reviewed scientific journals before September 1 2017 were retrieved from online databases (Scopus, Google Scholar) using initial search terms '*Hermetia illucens*', 'waste',

125 and 'conversion'. We extended our search for relevant articles via checking the reference list 126 and citations in each article. Though various studies reported conversion efficiencies on fresh 127 matter basis (insect biomass collected divided by the amount diet provided in %), obtained 128 efficiencies cannot be directly compared as considerable variation was present in the moisture 129 levels of the diets (12.3% in Lardé (1990) to 31.7% in Oonincx et al. (2015a)) and the larvae 130 (17.9% in Tschirner and Simon (2015) to 38.8% in Finke (2013)). With two resources both 131 being converted for 20% on fresh basis, on dry matter (DM) basis, one may be converted with 132 an efficiency of only 11% whereas for the other this would be 63%. We therefore focussed on 133 the conversion efficiencies on DM and N basis as is usual in insect feed conversion studies 134 (van Loon, 1991) and subsequently used these to calculate the environmental impact. 135 Forty articles evaluated in total 78 (mixtures of) resources (Table S1 in supplementary 136 material). BSF larvae were in particular fed with animal and human manure (Group E and 137 Group G, respectively), but also different types of food waste and various animal feed 138 materials have been tested (D and A, respectively). Few studies, however, evaluated the 139 suitability of gardening and forest material (E). Conversion efficiency on DM and N basis 140 was reported or could be calculated from data presented in 11 and 5 studies (Table 2), which 141 collectively evaluated 21 and 13 organic resources, respectively (Figure 1). As several 142 resources were fed as mixtures with different ratios (Liland et al., 2017; Rehman et al., 2017a; 143 Rehman et al., 2017b; Tinder et al., 2017), resources were tested twice (Tinder et al., 2017), 144 or fed at different feeding levels (Parra Paz et al., 2015), the total number of data points 145 exceeds the number of resources tested. In total, our dataset contained 62 values for DM 146 conversion and 34 values for N conversion. The studies differed in amount of resource 147 provided per larva and the number and age of the larvae at the start of the trial (Table 2). The 148 rearing temperature (~28°C) and relative humidity (~70%) were relatively similar among

149 studies. Timing of harvest differed among studies, varying from 5-6 instar and 16 day-old 150 BSF larvae to harvesting when one larva, 50% or all larvae were in the prepupal phase. 151 The DM conversion efficiency varied considerably among the 21 resources from 1.3% for 152 vegetal refuse and fruits (Parra Paz et al., 2015) to 32.8% for processed Chinese restaurant 153 waste (Zheng et al., 2012) (Figure 1). The N conversion efficiency in the 13 resources varied 154 from 7.4% for chicken and dairy manure (Oonincx et al., 2015b) to 74.8% for sorghum 155 (Tinder et al., 2017). Next to the variation in experimental set-up and rearing conditions 156 (Table 2), it should be noted that testing of multiple resources was suboptimal (see 157 Discussion) and, therefore, results into an underestimate of the conversion potential of the 158 BSF larvae.

159

160 2.3 Environmental assessment

Life cycle assessment (LCA) was applied according to ISO standards (2006b; a) to assess the environmental impact of larvae meal production. LCA is an internationally accepted and standardized holistic method to evaluate the environmental impact during the entire production chain (Guinée et al., 2002; Baumann and Tillman, 2004). LCA includes four phases, being goal and scope definition, inventory analysis (data collection), impact assessment (encompasses classification and characterization of the emissions and resources used), and interpretation of results.

168 *Goal and scope definition.* The goal of this study was to assess the environmental impact of

169 the production of fresh BSF larvae reared on different organic biomass resources. The

170 functional unit was a kg of fresh larvae and to compare it with other feed ingredients we also

171 expressed the impact per kg of larval protein.

172 *Inventory analysis.* Data related to the required inputs and outputs to produce one kg of fresh

173 larvae were obtained from literature (see section 2.2). In this study we only accounted for the

174 processes that are related to the environmental potential of the different resources: production 175 of the resource, processing of the resource, larvae rearing, and larvae/resource separation. We 176 assumed that the rearing plant is situated in The Netherlands. Not all studies identified in 177 section 2.2 (see Table 2) contained the data needed to perform the LCA and were therefore 178 excluded from the assessment of the environmental impact. Diener et al. (2011) was excluded 179 because no data were provided on the feeding level. Data of BSF larvae production based on 180 feeding seaweed (Liland et al., 2017) were not used as seaweed production technology is 181 currently under development and accurate estimates of the associated environmental impact 182 are unavailable. The control diet, i.e. processed wheat, however, was used in the analyses. 183 Data from Tschirner and Simon (2015) and Oonincx et al. (2015b) were not used as larvae did 184 show unacceptable growth due to feeding regime and the feeding substrate was not well 185 enough defined. Tinder et al. (2017) evaluated (mixtures) of feeding substrates twice and we 186 used the results of trail A. For the study of Parra Paz et al. (2015), the larvae to feeding 187 substrate ratio resulting in the highest conversion was used in the calculations. 188 Impact assessment. During the life cycle of a product, two types of environmental impacts are 189 considered: emissions of pollutants and use of resources, such as land or fossil-fuels (Guinée 190 et al., 2002). We assessed greenhouse gas (GHG) emissions, energy use, and land use. These 191 impacts were chosen because the livestock sector contributes significantly to both land use 192 and climate change worldwide (Steinfeld et al., 2006). Furthermore, energy use was used as it 193 influences global warming potential (GWP) considerably and plays an important role in the 194 rearing of insects (van Zanten et al., 2015). Land use was recalculated to square meters and expressed in m^2 kg of fresh larvae, whereas energy use was expressed in mega joules of 195 196 primary energy (MJ). The major GHGs related to livestock production (Steinfeld et al., 2006) 197 were included in this study: carbon dioxide (CO₂), methane (CH₄), and nitrous oxide (N₂O). 198 These GHGs were summed up based on their equivalence factors in terms of CO₂ (100 years'

time horizon) kg of fresh larvae: i.e. carbon dioxide (CO₂), biogenic methane (CH₄, bio): 28 kg CO₂-eq/kg, fossil methane (CH₄, fossil): 30 kg CO₂-eq/kg; and nitrous oxide (N₂O): 265 kg CO₂-eq/kg. Data related to emissions and resources were mainly obtained from databases and literature and are described in more detail in the next paragraphs. In case of a multifunctional process (e.g. production of soybean oil and meal), economic allocation was used, which is the partitioning of environmental impacts between co-products based on the relative economic value of the outputs (Guinée et al., 2002).

206

207 <u>2.3.1 Production of the resource</u>

208 As illustrated in Figure 1 feeding substrates consisted out of different organic products in 209 different proportions. The environmental impacts of most biomass resources within each 210 resource were derived from the ecoinvent database v3.3 (Table 3). Besides those biomass 211 resources, laying hen manure and dairy cattle manure were used as a feeding source. As there 212 was no specific data available about the use of manure for insect rearing, it was assumed that 213 emissions for using manure were equal to emissions from a laying hen and dairy farm. We 214 therefore accounted for CH4 and direct and indirect N2O emissions during the handling and 215 storage of chicken (laying hen) and dairy cattle manure (used as a feeding source). 216 To estimate emission of CH₄ and direct and indirect emission of N₂O from manure, a tier 2 217 approach was used based on country-specific data (Coenen et al., 2018; van Bruggen, 2018) 218 and IPCC default values (IPCC, 2006) (an emission factor of 0.03 CH₄ kg per laying hen per 219 year and 37.69 CH₄ kg per dairy cow per year, for direct N₂O 0.76 kg N excretion per laying 220 hen per year and 144 kg N excretion per dairy cow per year, 17.5 kg manure per laying hen 221 per year and 28,000 kg per dairy cow per year, and a default emission factor of 0.1, for 222 indirect N₂O: volatilisation 40% and an emission factor of 0.01).

224 <u>2.3.2 Resource processing</u>

225 Before the organic resources can be used as feeding substrate, processing is required. The 226 resource is mixed to create a homogeneous distribution of the different resources, and 227 grinding is done to create a texture that leads to an efficient digestion by the larvae (Parra Paz 228 et al., 2015b). Furthermore, drying and hydration processing were needed to obtain the 229 optimal moisture content (normally around 70%) of the feeding substrates. The impact of 230 grinding the material was assumed to be similar to the grinding of 1 kg of grains (ecoinvent). 231 For drying, we accounted for the removal of water per kg based on ecoinvent and adapted this 232 to each case.

233

234 <u>2.3.3 Larvae rearing</u>

235 Larvae were kept at a temperature of 28°C, a relative humidity of approximately 70% and 236 were fully grown after 16 days. A constant ventilation is needed to provide oxygen and 237 remove CO₂ and to avoid heat accumulation, which can occur due to high larval densities. 238 Light is not needed during larval development. The density of the larvae per crate was based 239 on Liland et al. (2017), amounting up to a density of about 830,000 larvae/m³. Energy needed for heating approximately one m³ of air to 28°C was about 0.57 kWh per day, based on data 240 241 obtained during experiments in climate chambers at the Laboratory of Entomology 242 (Wageningen University & Research, Wageningen, The Netherlands). 243 244 2.3.4 Larvae/resource separation 245 To harvest the larvae, the (remaining) resource will be sieved and we assumed that the energy

246 use of a sieving machine for nuts was similar as no other data was available (Brand: Yong

247 Qing, Model: XZS). The energy use to sieve one kg of larvae was about 0.025 kWh.

248

249 <u>2.3.5 Conversion to per kg protein basis</u>

To express environmental impact per kg larval protein, the DM and N values presented in the studies were used. For Tinder et al. (2017) the DM content and for Parra Paz et al. (2015) the DM and N contents of the larvae were not presented and the average DM and N presented in Oonincx et al. (2015a) were used (33.7 and 7.2%, respectively). For the studies of Rehman et al. (2017a, 2017b) a N content on DM basis of 6.7% (Newton et al., 1977) was used. To convert from N to protein, a conversion factor of 4.7 was used (Janssen et al., 2017).

256

257 **3 Results**

258 Table 4 presents the environmental impact per kg of fresh larvae per feeding substrate group. 259 Our results show that the environmental impact indeed largely depends on the type of 260 resource used. Of the different processes, the main environmental impact related to the 261 production of the resource, followed by processing of the resource, heating and lastly the 262 energy needed for the separation of the larvae from the (remaining) resource (Supplementary 263 material, Figure S1). Although similar conclusions were found in other studies (Oonincx and 264 de Boer, 2012; Smetana et al., 2016; Halloran et al., 2017), the relative contribution of each 265 process can easily shift depending on the type of resource used. Most data were available for 266 resources of Group A and Group E (10 and 12 values, respectively). Group A is the group that 267 is legally allowed and therefore represents the current situation. We do see, however, large 268 differences in the environmental impact within Group A (Table 4). In general, we can 269 conclude that resources that contain products that can also be used for human consumption 270 (food), like sorghum and cowpeas (Tinder et al., 2017), result in the highest environmental 271 impact (Figure 2 and Supplementary material Table S2). Resources that include co-products 272 or former foodstuffs generally used as feed, e.g. cookie remains tested in Oonincx et al. 273 (2015a), have a lower environmental impact in terms of GWP and energy use. Resources that

contain organic residual materials, i.e. products that are not used as food or feed, such as food
waste (Zheng et al., 2012) or manure (Rehman et al., 2017a), result general in in the lowest
environmental impact when expressed on per kg fresh larvae (Table 4) but not always when
expressed on per kg protein basis (see Figure 2 and Table S2). This relates to the conversion
factor used to express the impact on a per kg protein basis, which was higher for these
resources as mainly DM content was relatively low (i.e. on average 37.7% in Oonincx et al.
(2015a) and 21.9% in Rehman et al. (2017a, 2017b)).

281 BSF larvae have a high crude protein content and can replace fishmeal and soybean meal in 282 conventional livestock feeds. In Table 5 we compared the environmental impact of fishmeal 283 and soybean meal with the average environmental impact of larvae reared on resources 284 containing food ingredients, feed ingredients and residual resources. Our results show that 285 BSF larvae reared on resources containing residual resources offer potential to reduce the 286 environmental impact in terms of energy use and land use but not necessarily for GWP. While 287 BSF larvae reared on resources containing food or feed ingredients will most likely increase 288 the environmental impact. The control feeding substrates, i.e. processed wheat (Liland et al., 289 2017) in the food class and starter chicken feed (Oonincx et al., 2015a) and Gainesville diet 290 (Tinder et al., 2017) in the feed class, impacted the averages of the environmental impact 291 categories for these classes (see Figure 2). Excluding these feeding substrates would result in 292 a larger differences between the averages for these two categories being, respectively, 22 and 293 3 kg CO₂-eq, 192 and 43 MJ, and 79 and 0 m² per kg protein.

294

295 4 Discussion

296 4.1 Environmental impact

The BSF larvae have been suggested to play a role in promoting a circular economy viaupcycling of resources currently lost or not efficiently used in the food chain and acting as a

299 protein-rich feed ingredient for the livestock and aquaculture sectors (Makkar et al., 2014; 300 Henry et al., 2015). We assessed the environmental opportunities of insects reared on 301 different organic biomass resources described in the scientific literature and related their 302 potential to the current EU legislation framework. For a long time, the use of insects as food 303 and feed was not allowed at all but this situation has changed and since half a decade insects 304 are gaining more and more interest at the European level. In the summer of 2017, EU has 305 authorized the inclusion of insects in fish feed and it is expected that approval of insect 306 processed animal proteins (PAPs) to be fed to pigs and poultry is expected for 2019 307 (Andriukaitis, 2017). Next to a wider application of BSF larvae as a feed ingredient, 308 regulations for the resources to produce the insects will determine the degree to which BSF 309 larvae can be incorporated in the food system to make it more efficient and productive. At the 310 EU level it is currently not allowed to use insects as food or feed that are fed on resources 311 containing manure or waste due to safety regulations. At present, resources allowed for BSF 312 larvae production are those that are also fit for feeding pigs and poultry. From an 313 environmental perspective, it is crucial to consider BSF larvae production from residual 314 organic resources that are not considered as food or feed materials and currently left unused in 315 the food system. Our findings clearly show that only if we use residual streams as a feeding 316 substrate, BSF larvae production can result into environmental benefits (lower GHG 317 emissions and especially lower land use) compared to conventional protein-rich feed 318 ingredients with a high environmental impact. The studies of Smetana et al. (2016) and 319 Salomone et al. (2016) found similar results with high variations. Smetana et al. (2016) found values between 2.8 and 31.2 kg CO_2 -e and between 0.06 and 14.5 m² per kg of protein larvae 320 321 fed on municipal waste and beet-pulp, respectively. Salomone et al. (2016) found a value of 322 2.10 kg CO₂-e and of 0.05 m² per kg of protein, from larvae fed on municipal weight 323 (Salomone et al. 2016). Although GHG emissions can be reduced if residual streams are used,

it should be noted that there is limited information available related to potential emissions 324 325 from the resource or larvae which might have a substantial impact on the total GHG 326 emissions. Mertenat et al. (2019) measured CH₄ and N₂O emissions during BSF rearing on 327 food waste and concluded that CH₄ emissions were low along the rearing period while N₂O 328 did not differ significantly from the ambient, but tends to increase temporally after feeding 329 events. More research is needed on BSF N₂O emissions. Besides the direct environmental 330 impact that we assessed (as our aim was to compare the different resources) one could also 331 consider indirect consequences and a broader range of environmental impacts. The use of 332 organic resources can, for example, result in a competition with food, feed, fuel, and fertiliser 333 production for natural resources. The study of van Zanten et al. (2015) showed, for example, 334 that using food waste as feeding substrate for housefly larvae results in a direct competition 335 with bioenergy production, increasing the use of fossil fuels and subsequently resulted in a 336 higher environmental impact. Using residual streams with a limited application (e.g. manure 337 in The Netherlands due to the surplus) is therefore recommended to avoid this competition. 338 This competition can also be reduced by using residual streams as efficiently as possible, for 339 example, using the remaining material as fertiliser or to produce bioenergy. Before BSF 340 larvae production is implemented in practice more environmental assessment studies are 341 needed to get a better understanding about the role of BSF larvae within a sustainable food 342 system.

343

344 **4.2** Food safety

Although the use of residual streams as a feeding substrate offer the potential to reduce the environmental impact, they might result into food safety risks. It is therefore required to assess the potential associated food safety issues resulting from the use of residual streams as feed and, if food safety hazards are present, to investigate ways to mitigate them. For some

compounds that might pose a safety risk, incorporating BSF larvae in the food chain might
result into reduction of the compound (e.g. aflatoxin B1 in Bosch et al., 2017; Purschke et al.,
2017; Camenzuli et al., 2018) whereas for others, it might result into accumulation by BSF
larvae and pose risks (e.g. cadmium in Diener et al., 2015; van der Fels-Klerx et al., 2016;
Purschke et al., 2017).

354

355 4.3 Resource conversion efficiency

356 The characteristics of the organic resources play a pivotal role for this concept as well as how 357 efficiently these can be converted into insect biomass. Next to the amount of larval biomass 358 produced per unit of resource, the larval composition can be greatly influenced by the 359 resource. Crude protein contents of BSF larvae can range considerably with values from 360 34.9% of DM (Diener et al., 2009) to 57.0% (Dierenfeld and King, 2008), which would 361 impact the nutritional value, the required processing and, ultimately, the economics of 362 production. Our results show that BSF larvae can thrive on a wide range of organic resources, 363 but the DM conversion efficiency is known for less than 25% of the resources studied and N 364 conversion for 17% of the resources. Furthermore, it was noted that test procedures varied 365 considerably (Table 2) and that procedures in some studies were suboptimal to obtain 366 efficient conversion. For example, excessive fungal growth on the beet pulp was suggested to 367 have inhibited larval development and to have caused the observed low DM conversion factor 368 (Tschirner and Simon, 2015). Oonincx et al. (2015b) commented that the drying procedure 369 applied on the three manure types could have been detrimental to their nutritional value 370 and/or the microbiota in the manure. In Tinder et al. (2017) destructive sampling of larvae 371 was performed during the study, which reduced larval development and survival. 372 Furthermore, the latter study reported considerable variation in outcomes between two trials 373 with identical resources, which was potentially due to the use of different incubators and the

374 season in which the study was performed. Considering these issues, one should be cautious in 375 considering the presented efficiencies as representative for the resources tested. 376 Though the scientific literature describing studies on resource use by BSF larvae is rapidly 377 growing, the studies vary considerable in design. Standardised chemical characterization of 378 the organic resource used, basic rearing methodology, and post-harvest analyses of larvae and 379 residue are crucial to assess the potential of BSF larvae to convert such resources and to 380 improve our understanding of factors important for efficient conversion. Such standardised 381 operating procedures are in place for evaluating ingredients for livestock species. This has 382 resulted in publicly available feeding tables (e.g. Sauvant et al., 2004; CVB, 2011) describing 383 species-specific nutritional values of ingredients, which are instrumental for formulating diets 384 supporting optimal animal performance and use of resources. Researchers are preparing 385 standardised procedures for BSF larvae conversion studies and for reporting of findings 386 (Bosch et al., submitted), which will facilitate comparisons among studies and use of data for 387 future assessments of associated environmental impact for various resources used to produce 388 the larvae.

389 Both fishmeal and soybean meal are products present in the market since a long time, and 390 their production efficiency has increased in the previous decades, lowering their impacts on 391 the environment. We expect a similar increase in efficiency to evolve in the insect industry. 392 Feed optimization and genetic strain selection could lead to a general improvement of the 393 production efficiency, lowering the resulting environmental impact. At present, we are just 394 starting with understanding the factors that underlie the capacity to efficiently convert residual 395 feeding substrates. It is expected that efficiencies can be increased with the advancement in 396 understanding of how to optimise the interplay between the larvae and residing microbiota in 397 the feeding substrate during rearing and by tailoring BSF larvae strains to specific resources 398 by genetic selection. With many of the biological and technological concepts being already in

399 place and the short lifecycle of insects, it is possible that improvements can be achieved on 400 the short term with low costs relative to the livestock sector that ultimately lead to a more 401 economic larval production with a lower environmental footprint.

402

403 **5 Conclusions**

404 The number of studies evaluating the conversion of organic resources by BSF larvae is 405 growing, but vary considerably in design and few actually quantified conversion efficiency. 406 Our results on environmental impact show that resources within the legal groups that are, at 407 the moment, not allowed in EU as animal feed, have in general a lower environmental impact 408 than the ones that are currently allowed. BSF larvae reared on a resource containing residual 409 streams therefore offer potential to replace conventional feed protein sources and, thereby, to 410 lower the environmental impact of food production. More studies evaluating specifically these 411 residual resources as well as the assessments of potential food safety risks are required to 412 relax EU legislation and to bring promising residual streams into the food chain via BSF 413 larvae. BSF larvae reared on a resource that contains food and feed products generally have 414 relatively high environmental impacts. Further developments BSF production technology will 415 lower the environmental impact for these resources as well as making the production more 416 economic and competitive and contributing to reduction of the need for fishmeal and soybean 417 meal as animal feed.

418

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for helping with the processing of data and calculations of the environmental impact.

423

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Tables and figures

| Group | Description | Legal status | Legal justification |
|-------|--|-----------------|--|
| A | Animal feed materials according to the EU catalogue of feed materials and authorized as feed for food producing animals. | ✓ | Regulation (EU) No 68/2013 |
| B1 | Food produced for human consumption, but which is no longer intended for human consumption for reasons such as expired use-by date or due to problems of manufacturing or packaging defects. Excluding meat and fish (processed animal proteins, PAPs). | ✓ | Former foodstuffs of vegetab origin: Regulation (EU) No 68/2013 Permitted former foodstuffs of Animal origin (non-PAPs): Regulation (EU) No 142/2011, Annex X, Chapter II, Section 10 |
| B2 | Meat and fish produced for human consumption, but which is no longer intended for human consumption for reasons such as expired use-by date or due to problems of manufacturing or packaging defects. | Х | Regulation (EC) No 999/2001, Article 7(2) Regulation (EU) No 142/2011, Annex X, Chapter II, Section 10 Regulation (EC) No 1069/2009, Article 10(f) |
| С | By-products from slaughterhouses (hides, hair, feathers, bones etc.) that do not enter the food chain but originate from animals fit for human consumption. | Х | Regulation (EC) No 999/2001, Article 7(2) Regulation (EC) No 1069/2009, Article 10(b) |
| D | Food waste from food for human consumption of both animal and non- animal origin from restaurants, catering and households. | Х | • Regulation (EC) No 1069/2009, Article 11(1) |
| E | Animal manure and intestinal content. | Х | • Regulation (EC) No 1069/2009, Article 9(a) |
| F | Other types of organic waste of vegetable nature such as gardening and forest material. | ✓/X | Regulation (EC) No 767/2009, Annex III Regulation (EC) No 68/2013 |
| G | Human manure and sewage sludge. | Х | Directive 2008/98/EC Regulation (EC) No 767/2009, Article 6 Directive 91/271/EEC Directive 86/278/EEC |

Table 1. Groups of insect feeding source options and their legal status.

Table 2. Rearing conditions and timing of harvest for studies evaluating dry matter (DM) and N conversion of organic resources by black soldier

610 fly larvae.

| Rearing | | | | | | | Harvest | Conv | ersion | Reference |
|---|----------------|----------|-----|-------|-------|-----------|--------------------------|------|--------|-------------------------|
| Feed type (group ¹) | Amount of feed | Larvae | Age | Temp. | Light | RH | Life stage | DM | Ν | |
| | (g FM) | (number) | (d) | (°C) | (h) | (%) | | (%) | (%) | |
| Municipal organic waste (D) | NR | NR | 0 | 31.8 | NR | NR | Prepupae | + | — | Diener et al. (2011) |
| Dairy cow manure (E) | 1249 | ~1200 | 10 | 27 | Env. | 60- 75 | Prepupae | + | + | Li et al. (2011) |
| Solid residual fraction of defatted raw waste from Chinese restaurants (D) | 1000 | 1000 | 8 | 26-29 | NR | 65- 75 | 50% Prepupae | + | _ | Zheng et al. (2012) |
| Four mixtures of spent grains, beer yeast, cookie remains, bread remains, potato steam peelings, beet molasses (all A) | 13-19 | 100 | 0 | 28 | 12 | 70 | 1 st Prepupae | + | + | Oonincx et al. (2015a) |
| Dairy cow manure (E), pig manure (E), chicken manure (E) | 111-165 | 100 | 0 | 27 | 12 | 70 | 1 st Prepupae | + | + | Oonincx et al. (2015b) |
| Vegetal (plantain, potato, cabbage) and fruit (banana, papaya) refuse (A) | 96-1194 | 59-333 | NR | 26-28 | NR | NR | 50% Prepupae | + | _ | Parra Paz et al. (2015) |

| Wheat middlings (A), DDGS (A), beet pulp (A) | 19,200-20,000 | ~16,000 | 8 | NR | NR | NR | 5-6 instar larvae | + | + | Tschirner and Simon (2015) |
|--|---------------|---------|---|------|----|-----------|--------------------------|---|---|----------------------------|
| Seaweed (A) | 3000-12,000 | ~15,000 | 8 | 30 | 0 | 65 | 16 d old larvae | + | _ | Liland et al. (2017) |
| Dairy cow manure (E), chicken manure (E), and mixtures thereof | 1000 | 1000 | 6 | 27 | NR | 60- 70 | 1 st Prepupae | + | _ | Rehman et al. (2017a) |
| (E) Soybean curd residue (A), dairy cow manure (E), and mixtures | 1000 | 1000 | 6 | 27 | NR | 60- 70 | 1 st Prepupae | + | _ | Rehman et al. (2017b) |
| thereof (E) Sorghum (A), cowpeas (A), and mixtures thereof (A) | 93-297 | 300 | 4 | 28±2 | 14 | 70 | Prepupae | + | + | Tinder et al. (2017) |

611 Abbreviations: Temp., temperature; RH, relative humidity; FM, fresh matter; NR, not reported; DDGS, dried distillers grains with solubles.

⁶12 ¹Groups of insect feeding source options according to legislation in European Union (see Table 1).

613 **Table 3.** Environmental impact of resources¹ for global warming potential (GWP, kg CO₂-

| Resource | GWP | Energy use | Land use |
|--|------|------------|----------|
| Alfalfa | 0.38 | 2.04 | 1.69 |
| Beer yeast | 0.47 | 7.30 | 0.00 |
| Beet molasses | 0.33 | 3.70 | 0.22 |
| Beet pulp | 0.37 | 5.60 | 0.00 |
| Bread remains* | 0.00 | 0.00 | 0.00 |
| Cookie remains* | 0.00 | 0.00 | 0.00 |
| Corn meal | 0.64 | 6.50 | 1.20 |
| Cowpea | 0.67 | 5.51 | 3.20 |
| Dried distillers grains with solubles (DDGS) | 0.30 | 4.60 | 0.00 |
| Electricity 1 kWh | 0.75 | 11.80 | 0.01 |
| Grain semolina | 0.52 | 3.26 | 1.19 |
| Maize | 0.60 | 5.20 | 1.30 |
| Manure chicken | 0.04 | 0.00 | 0.00 |
| Manure dairy | 0.04 | 0.00 | 0.00 |
| Palm kernel expeller | 0.55 | 3.20 | 0.30 |
| Palm oil | 3.90 | 11.00 | 3.00 |
| Potato steam peeling* | 0.00 | 0.00 | 0.00 |
| Rapeseed expeller | 0.53 | 3.50 | 1.40 |
| Sorghum | 0.56 | 5.30 | 2.40 |
| Soybean meal | 0.41 | 6.10 | 3.20 |
| Vegetable oils | 1.59 | 11.00 | 3.00 |
| Spent grains | 0.38 | 7.37 | 0.00 |
| Vegetal and fruit refuse* | 0.00 | 0.00 | 0.00 |
| Water | 0.00 | 0.00 | 0.00 |
| Wheat | 0.40 | 2.90 | 1.10 |
| Wheat bran | 0.43 | 4.80 | 0.53 |
| Wheat middlings | 0.25 | 2.20 | 0.60 |

614 eq), energy use (MJ) and land use (m^2) per kg of product unless defined differently.

615 ¹Obtained from ecoinvent database v3.3 except for those products indicated with *, which

616 were considered to be wasted and have no environmental impact.

617 **Table 4.** Environmental impact of black soldier fly larvae production in terms of global

618 warming potential (GWP; kg CO₂-eq), energy use (MJ) and land use (m²) per kg of fresh

| Group ¹ | GWP | | Energy us | se | Land use | Land use | | |
|----------------------|---------|-------|-----------|--------|----------|----------|--|--|
| | Average | Range | Average | Range | Average | Range | | |
| A (10 values) | 1 | 0-3 | 17 | 2 - 24 | 5 | 0 – 11 | | |
| B (4 values) | 1 | 0 - 1 | 6 | 2 - 10 | 0 | 0 - 0 | | |
| D (1 value) | 0 | - | 1 | - | 0 | - | | |
| E (12 values) | 0 | 0 - 1 | 2 | 0-3 | 0 | 0 - 0 | | |
| Total (27 values) | 1 | 0-3 | 8 | 0 - 24 | 2 | 0-11 | | |

619 larvae reared on a resource per legal group.

 1 Groups of insect feeding source options according to legislation in European Union (see

621 Table 1).

- 622 **Table 5.** Comparison between soybean meal and fishmeal and black soldier fly (BSF) larvae
- 623 per kg protein for global warming potential (GWP; CO₂-eq), energy use (MJ) and land use
- 624 (m²).

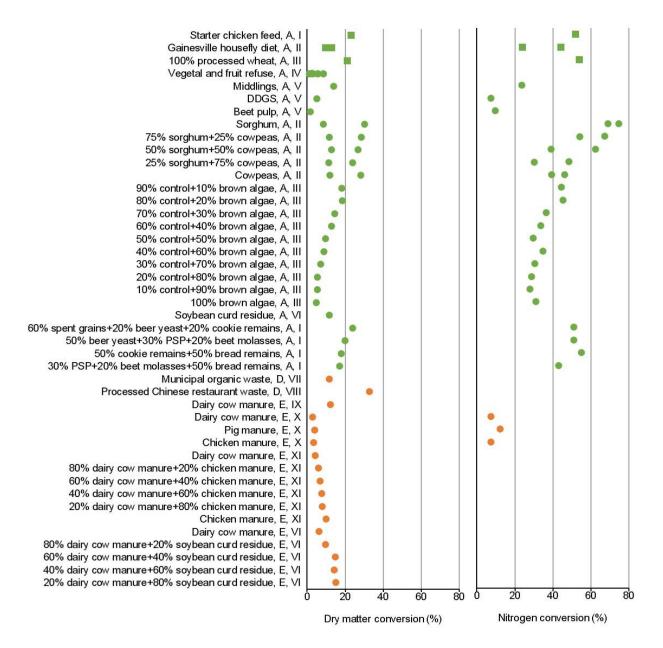
| Parameter | Fishmeal | Soybean meal | BSF ¹ | BSF ¹ | | | | |
|-------------------------|----------|--------------|------------------|------------------|----------|--|--|--|
| | | | Food | Feed | Residual | | | |
| GWP ² | 2.8 | 1.1 | 19 | 3 | 6 | | | |
| Energy use ² | 44 | 9 | 174 | 84 | 26 | | | |
| Land use | 0.0 | 3.4 | 67 | 3 | 0 | | | |

⁶²⁵ ¹BSF-Food are larvae reared on products that humans can consume, BSF-Feed are larvae

626 reared on co-products that are generally fed to livestock, and BSF-Residual are larvae reared

627 on products that are not used as food and feed.

⁶28 ²Drying is excluded, which, depending on the method, would increase GWP and energy use.





630 **Figure 1.** Dry matter (DM) and nitrogen (N) conversion efficiency¹ for various (mixtures of)

631 organic resources² as reported or calculated from data provided in scientific literature³.

⁶³² ¹Conversion efficiency was defined as collected insect biomass divided by amount of diet

- 633 provided, both in grams DM or N.
- ⁶³⁴ ²Groups of resources are indicated (A to E) as well as their legal status according to the
- European Union (allowed in green, not allowed in orange; for details see Table 1). The first
- three resources (data points depicted as squares) were used as reference resources in the

- 637 studies. For some studies, multiple data points exist per diet, which reflects replication of the
- 638 study or variations in feeding level (i.e. g diet provided per larva).
- ³References, I, Oonincx et al. (2015a); II, Tinder et al. (2017); III, Liland et al. (2017); IV,
- 640 Parra Paz et al. (2015) with the vegetal refuse and fruits consisting out of 21% plantain, 17%
- 641 potato, 20% banana, 6% papaya and 36% cabbage; V, Tschirner and Simon (2015); VI,
- Rehman et al. (2017b); VII, Diener et al. (2011); VIII, Zheng et al. (2012) with processed
- 643 material being the solid residual fraction of defatted raw waste from Chinese restaurants; IX,
- 644 Li et al. (2011); X, Oonincx et al. (2015b); XI, Rehman et al. (2017a).

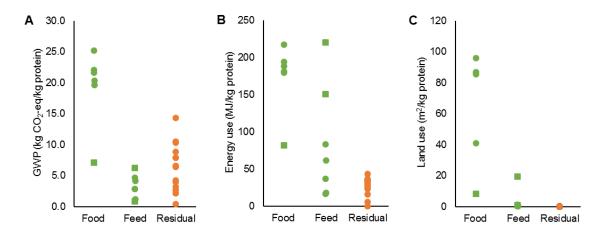


Figure 2. Environmental impact of black soldier fly larvae production in terms of global
warming potential (GWP; panel A), energy use (B) and land use (C) reared on a resource
being a food, feed or residual product¹.

645

⁶⁴⁹ ¹Food products are products that humans can consume (n=6 values); feed products are co-

650 products that are generally fed to livestock (n=7); and residual products are products that are

not used as food and feed (n=14). Colours refer to the legal status according to the European

Union, i.e. allowed in green, not allowed in orange (for details see Table 1). The data points

653 depicted as squares were used as reference resources in studies (see Figure 1).