## Current issues in the aquatic risk assessment for pesticides: HARAP, CLASSIC and ELINK revisited









## Aquatic Guidance Document (EFSA PPR, 2013)

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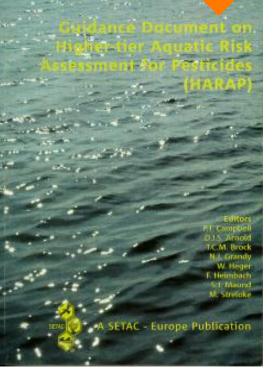
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#### HARAP Workshop 1998 (SETAC, 1999) Recommendations for higher-tier effects assessment

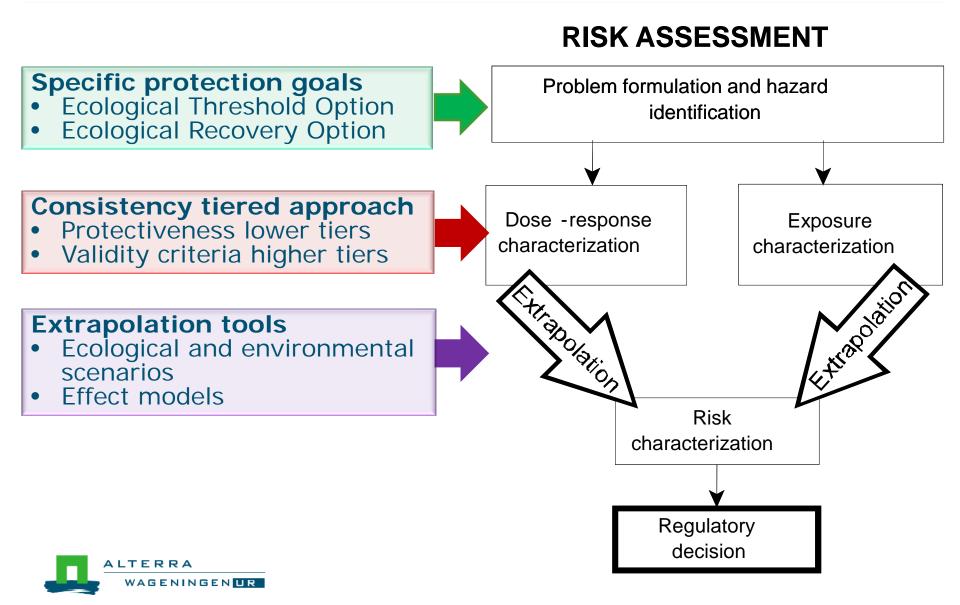
CLASSIC Workshop 1999 (SETAC, 2002) Further guidance on interpreting micro/mesocosm studies Linking Aquatic Exposure and Effects

> Risk Assessment of Pesticides

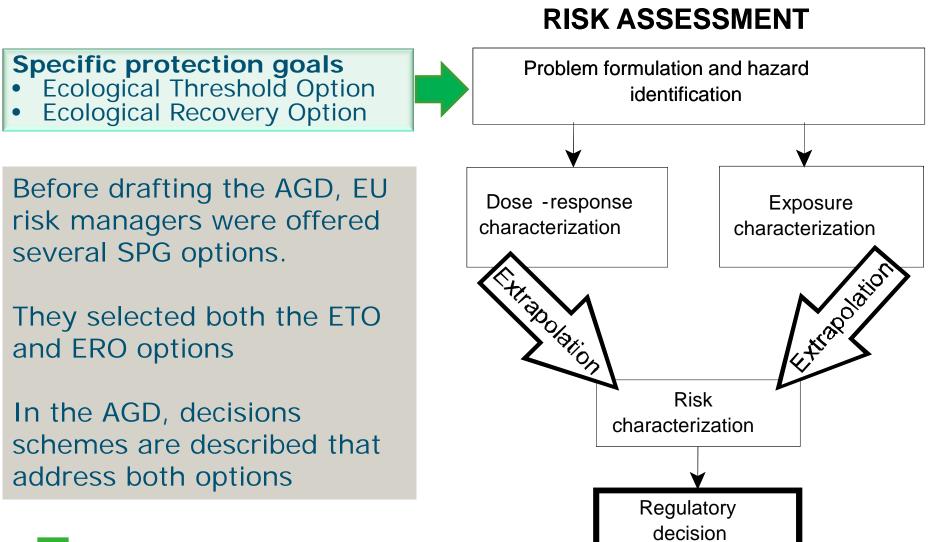
> > Editors Theo C.M. Brack Anne Alix Colin D. Brown Encer Capri kenthard EE Gonesbision Fred Heinsbach Ohris M. Lychgo Balf Schole Marris Stelleke

ELINK Workshop 2008 (SETAC, 2009) Recommendations for linking exposure and effect estimates

### Issues discussed in this presentation



### Specific Protection goals in AGD (EFSA PPR, 2013)





### Specific protection goals for water organisms

### Ecological threshold option (ETO)

Organism group	Ecological entity	Attribute	Magnitud	e Time
Algae	population	abundance/ biomass		
Aquatic plants	population	biomass negligible abundance/ effect		not
Aquatic invertebrates	population			applicable
	individual	survival		
Vertebrates	population	abundance/ biomass		
Aquatic microbes			RA is not c since Tier- requirement defined	•

Specific protection goals for water organisms

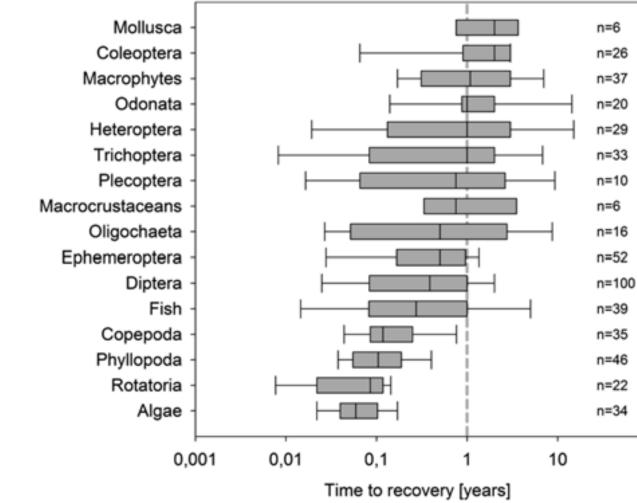
### Ecological recovery option (ERO)

Organism group	Ecological entity	Attribute	Duration and magnitude of effect on sensitive and vulnerable populations		
Algae	Population	Abundance/ Biomass	Total effect period < 8 weeks (also for repeated applications)		
Aquatic plants	plantsPopulationabundance/ BiomassAquaticPopulationabundance/ Biomass		Usually not possible for vulnerable populations with long life cycles and low		
Aquatic invertebrates			dispersal abilities Not leading to ecologically important indirect effects		
Vertebrates	No recovery option				

### Ecological Recovery Option (ERO)

- ERO may be addressed by micro-/mesocosm experiments and effect models for vulnerable taxa at risk
- Reluctance of regulatory authorities to accept an ERO-RAC derived from a micro-/mesocosm experiment
  - Representativeness of test system for vulnerable species
  - Possible risks due to simultaneous or repeated use of different PPPs
- Reluctance of regulatory authorities to accept population models in absence of EFSA guidance
  - Lack of experience and expertise in interpreting results of population-level models at MS level

Reported recovery times (years) for different groups of aquatic organisms independent of stressor



Gergs et al. 2016. Reviews Environ Contam Toxicol 236, 259-294

*Boxes* represent quartiles and whiskers symbolize 95 % confidence intervals. *n* = number of recovery endpoints

Recovery times differ between taxonomic groups and variability within taxonomic groups and between habitats high

### Species traits affecting recovery

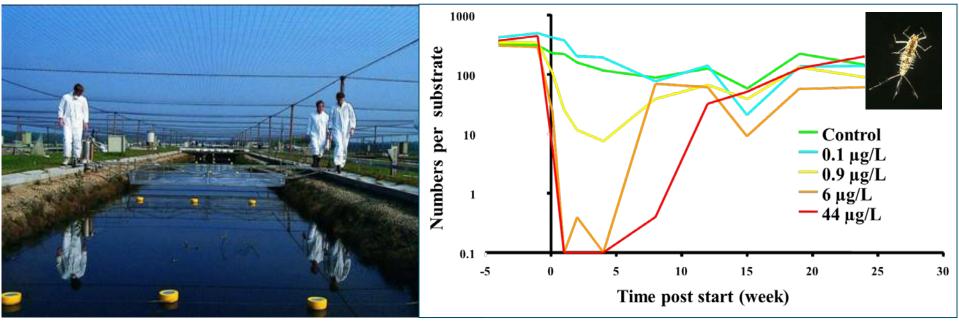
### Life-history traits determining internal recovery:

- Voltinism (number of generations per year)
- Reproduction rate
- Resistant life stages
- Development time

Additional traits determining external recovery:

- Dispersal ability (active or passive)
- Other important traits
  - Habitat choice and chance to become exposed
  - Susceptibility to indirect effects
  - Genetic diversity / population fitness (particularly important for small populations of endangered species)

### Evaluation of micro-/mesocosm tests (EFSA PPR, 2013)



- ETO and ERO Option: Possibility to demonstrate treatment-related effects for a sufficient number of **potentially sensitive populations** (at least 8)
- ERO Option: The observation period is long enough to demonstrate effects and recovery for representative **vulnerable taxa**

Vulnerable populations in micro-/mesocosms

### Criteria

Chance to become exposed to the pesticide(s)

- Habitat preference (e.g. an epi-benthic arthropod and exposure to an insecticide that accumulates at the water-sediment interface)
- Intrinsic sensitivity
  - Specific toxic mode-of-action
- Recovery potential
  - Species traits (*e.g. uni-/semi-voltine*)
  - Properties of test system/habitat (e.g. isolated; no refuges)

Motivate that either the potentially sensitive species with a poor recovery potential are not impacted or that the conditions for recovery were not 'best case' in the test systems used. Alternatively, use effect models for extrapolation.

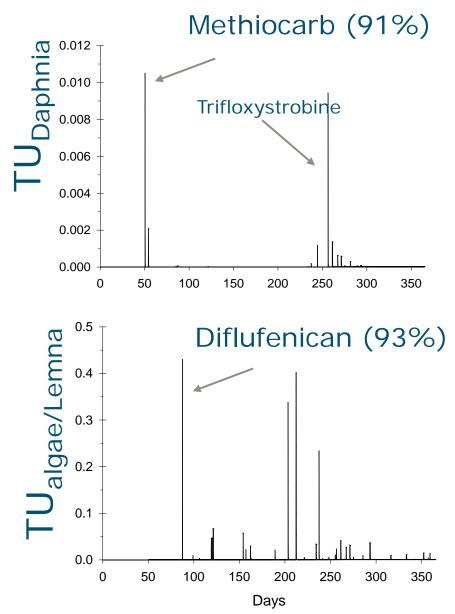
### Ecological recovery and intensive PPP use

EFSA PPR (2013) states: "...it is more uncertain if the ERO option can be achieved when assessing risks for individual PPPs for their use in crop protection programmes characterised by intensive PPP use."

Insight in potential impact of cumulative PPP-stress due to normal agricultural practise is required to address this concern

- Crop-oriented approach
- Landscape-oriented approach

### Crop approach in ERA for pesticides



Experimental and modelling studies simulating the crop approach for edge-of-field surface water reveal:

- Individual pulse exposures are generally dominated by single substances
- Different pulses may concern different substances
- Major effects observed usually are caused by one or a few pesticides

Van Wijngaarden et al. (2004) ET&C 23:1479-1498 Arts et al (2006) IEAM 2:105-125 Auber et al. (2011) Ecotoxicology 20:2042-2055 Focks et al. (2014a) ET&C 33:1489-1498

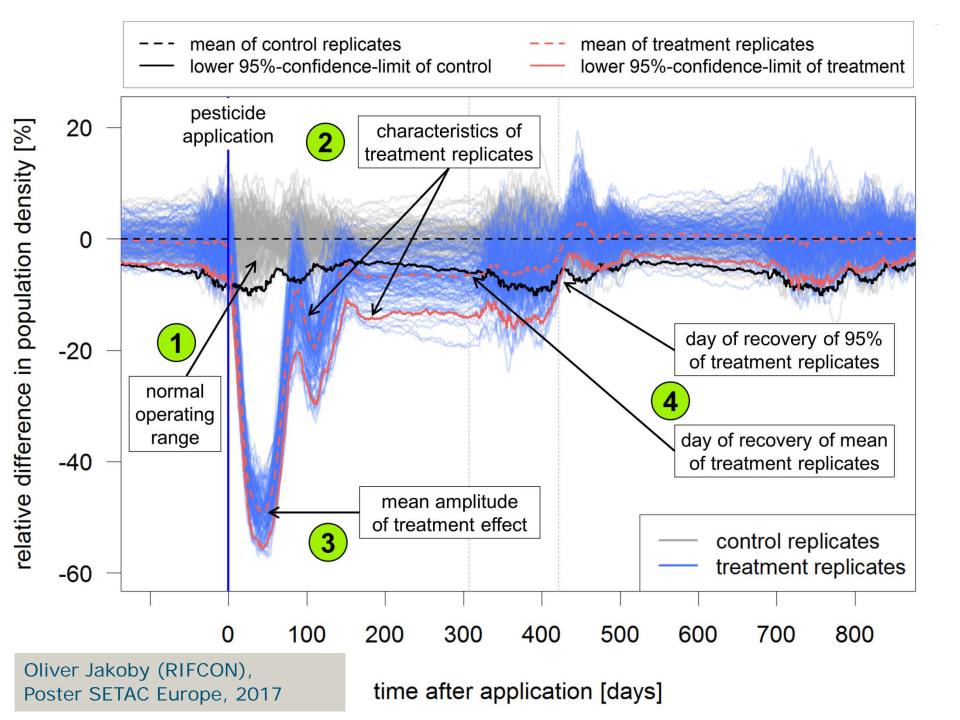
## ERO option requires a systems approach since many factors affect ecological recovery of non-target taxa

Extinction risk low Internal recovery key Short Many Yes Low Local ERA **Recovery option** Possibility/ability to escape stressor in time High (Semi-)field experiments always feasible Number of offspring Resistant life stages Stressor persistence Focal communities, species & landscapes Generation time Direct effects & ecological interactions External recovery key Landscape ERA Recovery Spatially structured Few No High option not feasible WO. **Extinction risk high** Population models Low High Possibility/ability of species to escape stressor in space Small \_ \_\_\_\_\_ Individual home range Large EFSA SC, 2016 Specialised \_\_\_\_\_ Habitat/food preference \_\_\_\_\_ Flexible **Recovery in ERA** ——— Refuge connectivity/availibility —— High Small -

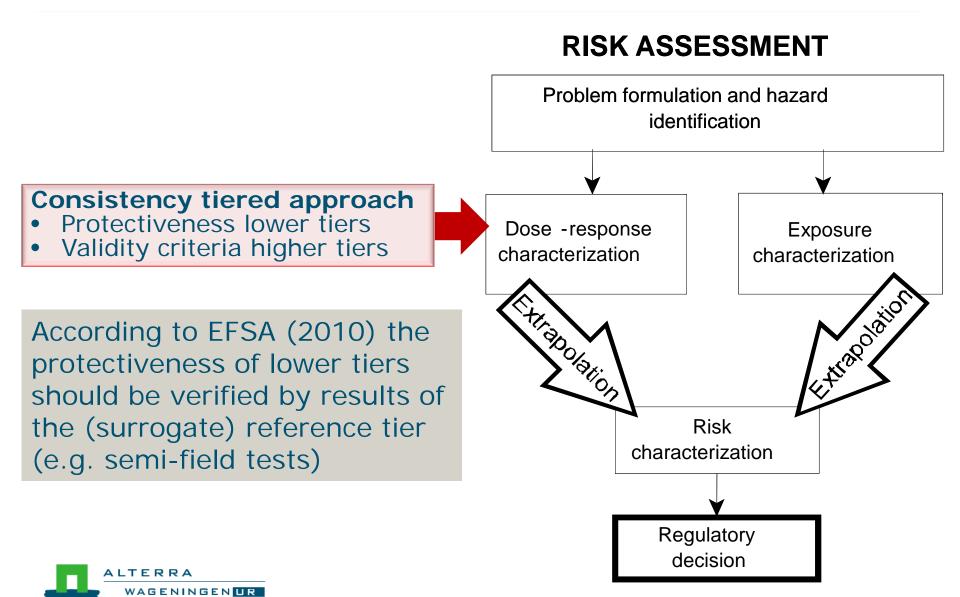
Large -

— Spatial scale of exposure \_\_\_\_\_ Small

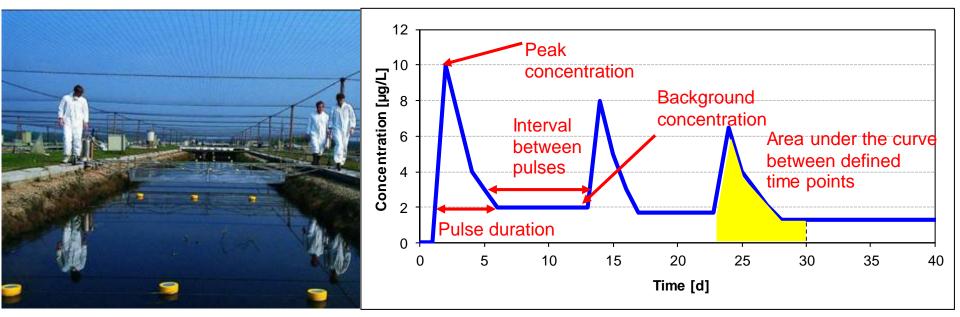
at EFSA



### Consistency of tiered approach



### Validity criteria micro-/mesocosm tests (EFSA 2013)



The **exposure** in the test system is **relatively worst case** to that predicted for edge-of-field surface water (pulse height and duration, number of pulses, interval between pulses)

### Information on the statistical power of test

 Appropriate Minimum Detectable Difference (MDD) values for at least 8 taxa of the potentially sensitive taxonomic groups

### Minimum Detectable Difference (MDD)

- The MDD defines the mean amount of difference between a treatment and the control that must exist to detect a statistically significant effect
- The lower the MDD, the less severe treatmentrelated declines in population abundance between controls and treatments need to be, to calculate a NOEC/LOEC

$$MDD = (\bar{x}_0 - \bar{x})^* = t_{1-\alpha,df,k} \sqrt{\frac{s_0^2}{n_0} + \frac{s^2}{n}}$$

$$MDD = (\bar{x}_0 - \bar{x})^* = t_{1-\alpha,df,k} \sqrt{\frac{1}{n_0} + \frac{1}{n}}$$

$$(\bar{x}_0 - \bar{x})^* = corresponding difference between control and treatment mean t_{1-a} = quantile of the t-distribution degrees of freedom number of comparisons s^2 = residual variance one-way ANOVA n_1, n = sample sizes$$

# MDD classes are given in EFSA AGD (but further guidance not given)

MDD	MDD%	Comment
Class		
0	>100%	No effects can be determined statistically
1	90-100%	Only large effects can be determined
		statistically
11	70-90 %	Large to medium effects can be determined
		statistically
111	50-70 %	Medium effects can be determined statistically
IV	< 50%	Small effects can be determined statistically

Proposal how to use MDD information in the evaluation of micro-/mesocosm tests for te derivation of ETO-RAC and ERO-RAC values is given by Brock et al. (2015. Environ Sci Pollut Res 22:1160-1174)

### How to report MDDs

## Minimum Detectable Difference (MDD) should be reported in concert with NOEC/LOEC values

#### Geometric mean abundance of Daphnia galeata

day	Controls	2 µg/L	6 μg/L	18 µg/L	54 μg/L	162 μg/L	Williams	%MDD <sub>abu</sub>
-5	94.3	93.3	88.8	139.3	86.2	108.5	- NOEC>=162 μg/L (incr.)	40.9
3	121.1	131.2	97.2	158.7	87.9	16.0	* NOEC=54 μg/L (decr.)	42.6
9	114.0	107.4	32.9	49.2	26.4	1.1	* NOEC=18 μg/L (decr.)	70.5
23	98.1	142.1	143.6	147.9	36.4	2.6	* NOEC=18 μg/L (decr.)	44.4
37	50.2	44.0	49.7	49.2	42.7	10.0	* NOEC=54 μg/L (decr.)	68.4
51	35.0	50.2	28.3	45.4	43.2	16.6	- NOEC>=162 μg/L (decr.)	57.6
65	35.0	87.9	29.2	32.9	108.5	18.6	- NOEC>=162 μg/L (decr.)	67.2
79	54.9	122.3	39.1	66.4	218.5	45.8	- NOEC>=162 μg/L (decr.)	82.9

The MDD<sub>abu</sub> needs to be below 100% to allow a statistical evaluation on treatment-related declines in abundance, and subsequent recovery.

The lower the MDD the larger is the power of the test.

### How to report MDDs

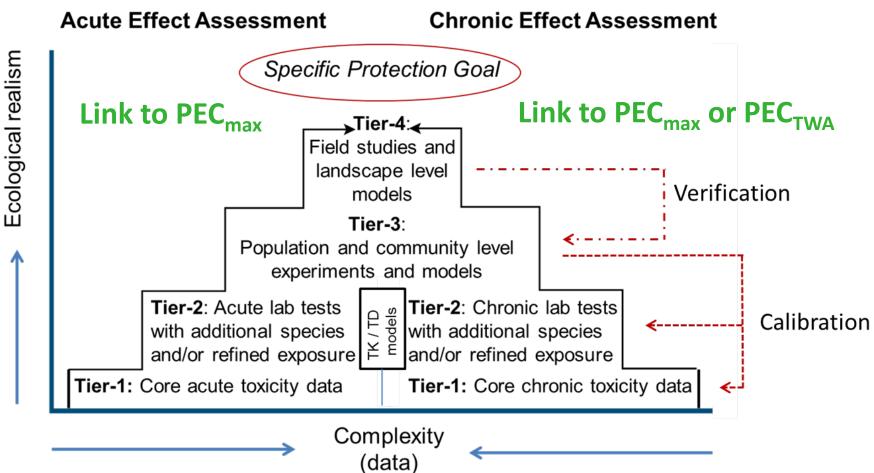
Geometric mean abundance of Stylaria lacustris

day	Controls	2 μg/L	6 μg/L	18 µg/L	54 µg/L	162 μg/L	Williams	%MDD <sub>abu</sub>
-5	7.9	5.0	13.8	15.3	6.1	6.2	- NOEC>=162 μg/L (incr.)	94.0
9	2.5	2.6	2.1	2.0	1.6	1.4	- NOEC>=162 μg/L (decr.)	107.1
23	5.3	4.5	5.3	6.0	3.8	2.4	<mark>* NOEC&gt;=18 μg/L (decr.)</mark>	71.9
37	2.1	3.1	3.2	4.2	2.3	2.0	- NOEC>=162 μg/L (decr.)	104.9
51	0.5	1.0	1.5	2.2	0.5	0.0	- NOEC>=162 μg/L (decr.)	173.2
65	1.7	0.4	1.7	1.0	6.0	0.0	- NOEC>=162 μg/L (decr.)	114.8
79	0.8	1.0	1.5	0.4	5.2	1.4	- NOEC>=162 μg/L (incr.)	144.2

If the MDD is consistently larger than 100% then

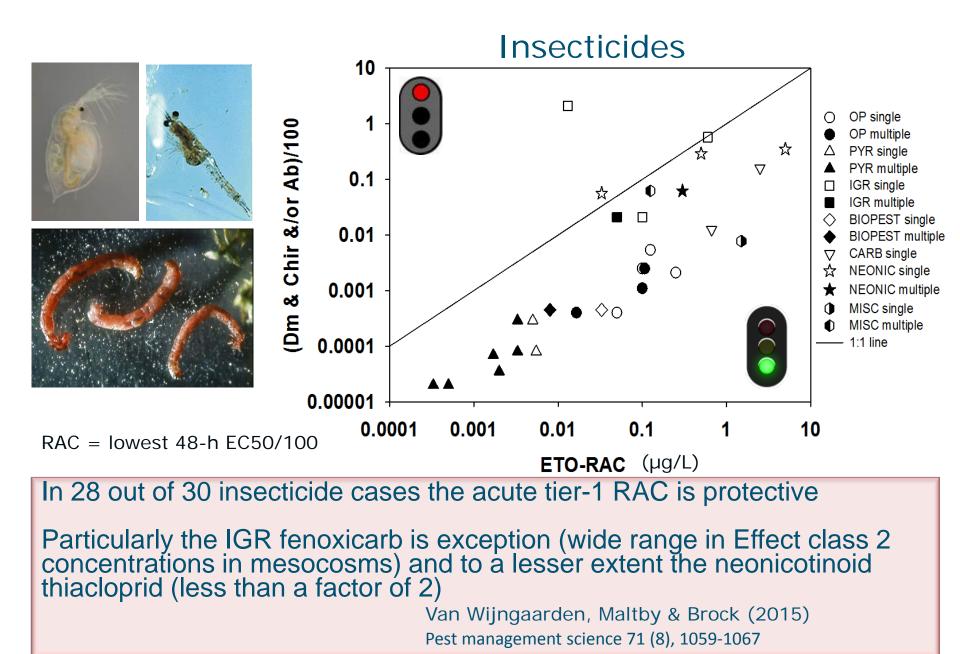
- the statistical power is too low to demonstrate treatment-related declines
- it will be difficult to draw firm conclusions on recovery if on isolated samplings a NOEC can be calculated

## Tiered risk assessment schemes

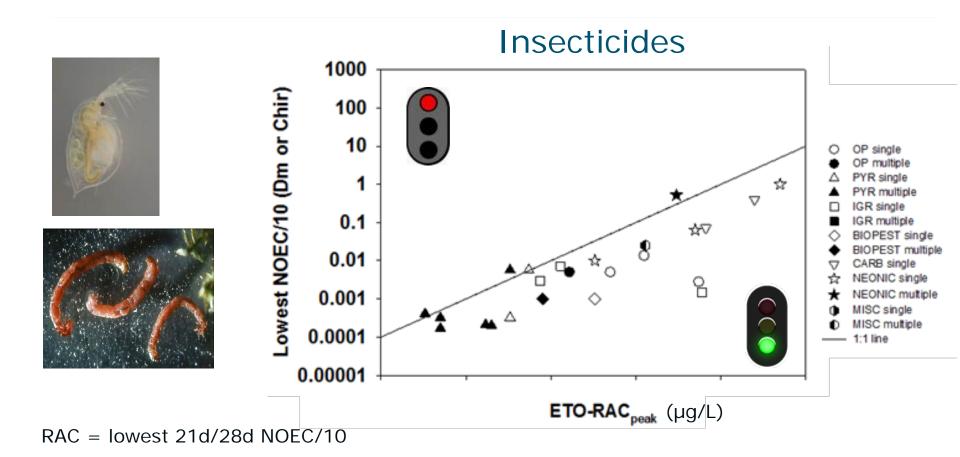


Micro-/mesocosm tests can be used to evaluate the protectiveness of lower tiers

### Calibration/verification acute Tier-1



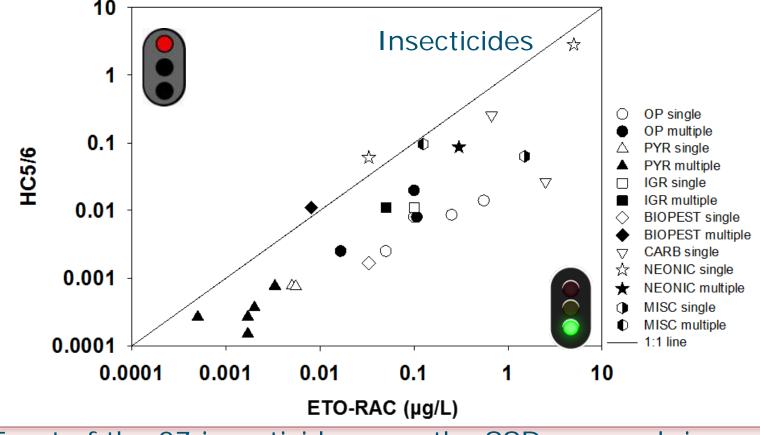
### Calibration/verification chronic Tier-1



In 21 out of 24 insecticide cases the chronic tier-1 RAC is protective Brock et al. (2016) IEAM 12:747-758

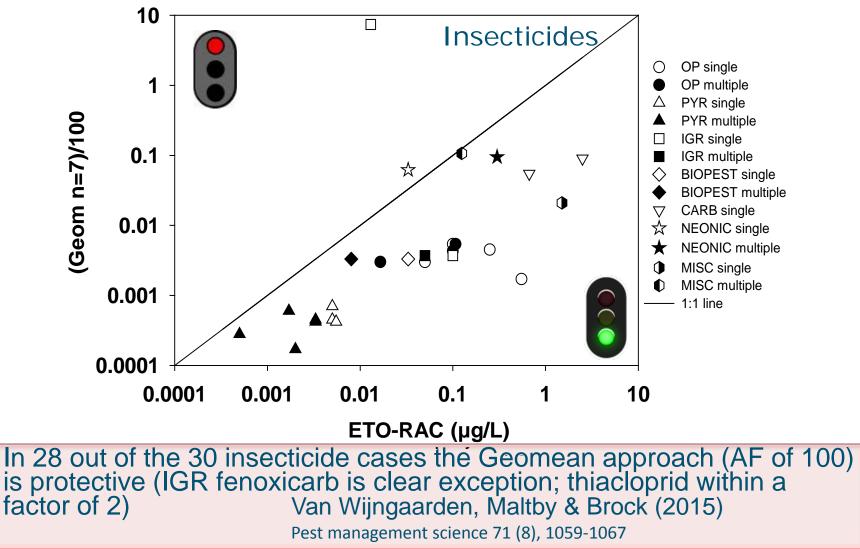
## SSD-RAC calibration with micro-/mesocosm RACs (ecological threshold option)

Assessment on basis of SSDs for arthropods (acute HC5/6)



In 25 out of the 27 insecticide cases the SSD approach is protective, but two borderline cases within a factor of 2 (thiacloprid and abamectin) Van Wijngaarden, Maltby & Brock (2015) Pest management science 71 (8), 1059-1067 Geom-RAC calibration with micro-/mesocosm RACs (ecological threshold option)

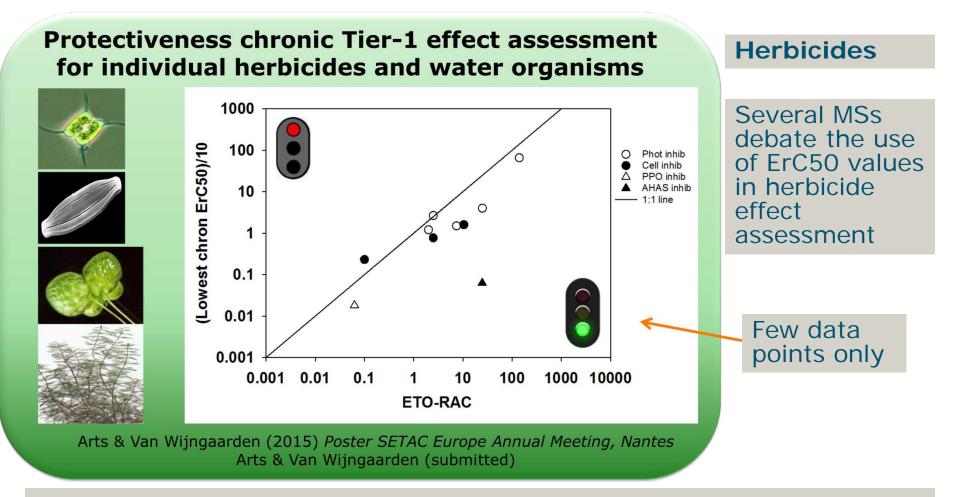
Lowest acute Geomean/100 value for insects and crustaceans



Geom-RAC calibration with micro-/mesocosm RACs (ecological threshold option)

- Geom-RACs could only be calibrated for acute toxicity data and insecticides
- The EFSA Guidance document recommends to use toxicity data of the same taxonomic group and of a similar endpoint and similar test duration (e.g. 48hmortality or 96h-immobility) in the Geomean approach
- The requirement to use similar endpoints and test durations may hamper the use of the Geomean approach in chronic effect assessments
- Currently, EU Member States do not use the Geomean approach in chronic risk assessments (a weight of evidence approach is advocated)

### Consistency of tiered approach



More efforts needed to verify the consistency of the tiered approach for aquatic ERA of herbicides and fungicides, and sediment ERA for pesticides in general

### Extrapolation tools

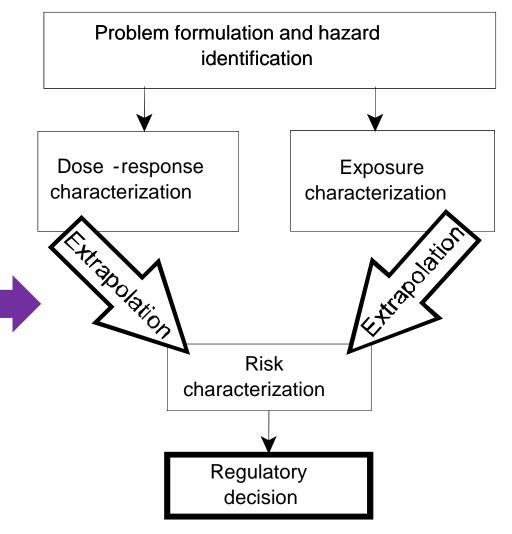
Currently exposure assessment mainly depends on modelling approaches while the effect assessment is based on experiments

Several developments to promote modelling approaches in effect assessment

#### **Extrapolation tools**

- Ecological and environmental scenarios
- Effect models

### **RISK ASSESSMENT**





### Experimentation and modelling

	Model ecosystems	Population / Ecosystems models
Pros	<ul> <li>Design easier to understand &amp; communicate</li> <li>Represents a 'real' system</li> <li>Guidance available</li> </ul>	<ul> <li>Controlled by the modeller</li> <li>No population and spatial/temporal limitations</li> <li>Easier to extrapolate</li> </ul>
Cons	<ul> <li>Not all focal populations present</li> <li>Spatial/temporal scale limited</li> <li>Extrapolate problems</li> </ul>	<ul> <li>Black box problem</li> <li>Validation issues</li> <li>Guidance not yet available</li> </ul>

Promising modelling approaches

- TK-TD models as tools to assess the effects of timevariable exposures
  - GUTS approach for acute risk
  - DEBTOX approach for chronic risks (validation issues)
  - EFSA scientific opinion under development
- Population and community-level models for spatialtemporal extrapolation of experimental studies
  - Recovery option
  - Integrated exposure and effect assessment
  - Requires appropriate scenarios

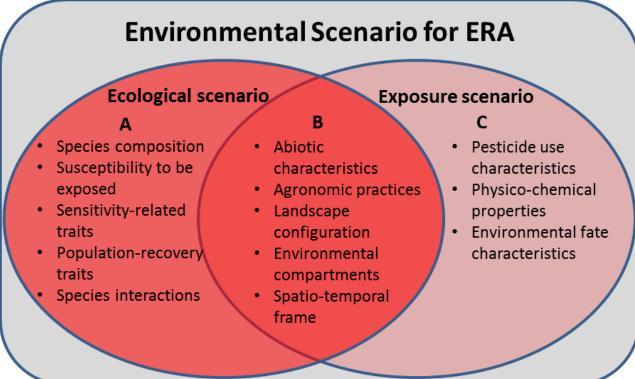
Population and community-level assessments require environmental scenarios to which both fate and effect experts can refer (EFSA PPR, 2014)

# Integration of exposure and ecological scenario in an environmental scenario

- Environmental scenarios: combination of biotic and abiotic parameters that are required to provide a realistic worst-case representation of the exposure, effects and recovery in the ecological entities (focal species) that are evaluated
- FOCUS exposure scenarios available, but ecological scenarios are urgently needed

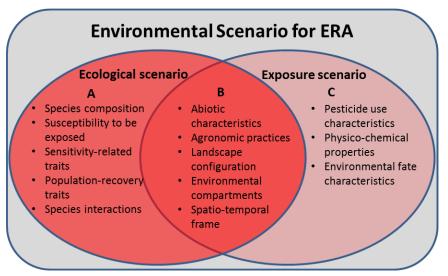
Rico et al., 2015 IEAM 12, 510-521





## Population and community-level modelling

- Selection of focal species and ecological scenarios should be developed in a standardized way for the main landscape units and climatic regions in EU
- The implementation of ecological scenarios may require a refinement of exposure scenarios (e.g. allowing spatially explicit exposure assessments)



- Environmental scenarios will contribute to an increase in 'ecological realism' of ERA
- But new expert knowledge required
- EFSA PPR will probably start with a scientific opinion on population and community-level modelling in 2018 - 2020

### Thank you for your attention Questions ?

