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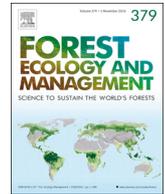
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# Effects of wild ungulates on the regeneration, structure and functioning of temperate forests: A semi-quantitative review



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## ABSTRACT

Wild ungulates such as red deer, roe deer and wild boar are key drivers of forest ecosystems. Across the northern hemisphere, their range and abundance is increasing, while at the same time forest conversion and habitat fragmentation have led to a large variation in ungulate density and composition among areas. Understanding ungulate density impacts are important in order to prevent shifts towards undesired states, such as from forest to heathland. Here, we assess the effects of ungulate density on forest regeneration, development and functioning. We carried out a systematic literature review of 433 published studies in temperate forests, and used the data to model dose-response curves of the effects of ungulate density on three sets of forest attributes; tree regeneration (abundance, species richness and composition), forest structure (horizontal and vertical), and forest functioning (nutrient cycling in soil, timber and food production). Ungulate density averaged  $23.6 \text{ km}^{-2}$  across studies. Ungulates had a negative effect on forest regeneration, structure and functioning in 70% of the evaluated cases. The dose-response curves had a sigmoidal, rather than a unimodal shape. Critical tipping points, where ungulates started to have a negative effect on forest regeneration, were found at an ungulate metabolic weight density of  $115 \text{ kg km}^{-2}$  for forest regeneration,  $141 \text{ kg km}^{-2}$  for forest structure, and  $251 \text{ kg km}^{-2}$  for forest functioning, which is roughly equivalent to 10, 13 and 23 roe deer per  $\text{km}^{-2}$ . Forest regeneration was most sensitive to immediate browsing and trampling impacts of small seedlings, while forest functioning was least sensitive because of time lags. However, these effects may build up over time. We suggest research priorities for studying ungulate-plant interactions in temperate forests, and make management recommendations how to balance wildlife with a functioning forest.

## 1. Introduction

Wild ungulates are increasing in density across the northern hemisphere (Glutton-Brock and Albon, 1992; Reimoser, 2003; Pellerin et al., 2010) because of ungulate reintroduction, abandonment of agricultural land, competitive release from domestic ungulates, absence of top predators, stricter hunting regulations and improvement of habitat quality (Kuiters et al., 1996; Rooney, 2001; Côté et al., 2004). Apart from inter-annual fluctuation, habitat quality and predation, conversion of natural forests to managed forests (Gordon and Prins, 2008) has led to the isolation of ungulates in different forest fragments (Kuiters et al., 1996). This has resulted in a large variation in ungulate density between fragments, and hence, a large variation in ungulate effects on the environment. Many temperate forests currently harbour large ungulate

populations resulting in intensive plant-animal interactions. Whether these population levels are acceptable is the subject of intensive debate among stakeholders (Horsley et al., 2003).

Ungulates affect ecosystems through browsing, trampling, fraying, stripping, uprooting, defecation and seed dispersal (Bruinderink and Hazebroek, 1996; Reimoser, 2003; Pellerin et al., 2010). These interactions are key determinants of the structure and dynamics of woody ecosystems. For instance, when ungulates are regulated naturally by predation and intraspecific competition, large herbivores can remove up to 10% of the above ground bottom-up control, where plants limit ungulate populations by chemical and structural defences that prevent herbivory, and top-down control, where predators regulate ungulates by predation (Terborgh et al., 2001).

Ungulates affect ecosystems at different organizational, spatial, and

Abbreviations: MWD, metabolic weight density

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temporal scales, ranging from the local patch up to the landscape scale. For instance, removal of the vegetation biomass has implications for local plant composition and structure, and with time this may lead to changes in soil fertility and the landscape, thus affecting the entire food web (Gordon and Prins, 2008; Prins and Fritz, 2008; Svenning et al., 2015). Their effect can enhance or reduce natural processes and patterns; which may directly affect species regeneration, forest structure and ecosystem functioning (Reimoser, 2003).

Whether ungulates have a positive or negative impact on ecosystems depends on their density, browsing intensity, local biotic and abiotic conditions, and forest management (Van Hees et al., 1996; Heckel et al., 2010; Pellerin et al., 2010). Because of their important role in ecosystem functioning, ungulates are considered to be keystone species and landscape modifiers (Rooney, 2001). Hence understanding the relationship between ungulate abundance and forest regeneration is fundamental.

Whether ungulates have positive, neutral, or negative effects on forest attributes depends on the shape of the dose-response curve and lag times (Nuttall et al., 2014). The most plausible shape of this dose-response curve is unimodal (Rooney and Waller, 2003). At low ungulate density, forests have a low plant species diversity because of a structurally uniform and dense vegetation with little heterogeneity in light and habitats. At intermediate ungulate density, forests have higher habitat heterogeneity due to vegetation removal, seed bed preparation (by litter removal and soil disturbance), and seed dispersal (Lucas et al., 2013). At the same time, ungulates may steer plant competition and succession through selective browsing on palatable species, thus facilitating the establishment of other plant species (Kuiters et al., 1996; Fuller and Gill, 2001; Brullhardt et al., 2015). Furthermore, as a result from low ungulate trampling, soil compaction is low and thus facilitating seedling establishment. At high ungulate density, however, tree regeneration is hampered by over-browsing, while selective browsing may lead to suppression of palatable species thus reducing tree species diversity (Tyler et al., 2008; Schippers et al., 2014). In ecotone transitions, it is even possible that forest shifts towards an alternative stable grassland state (Côté et al., 2004), with a high degree of soil compaction, low stem density and almost no canopy cover.

Understanding ungulate density impacts are important in order to prevent shifts towards undesired states, such as from forest to heathland (Scheffer et al., 2001; Folke et al., 2004). However, despite a wealth of studies that have assessed the effect of ungulates on vegetation, we know very little about the shape of the dose response curve between ungulate density and forest attributes, and whether there are critical thresholds and tipping points (Putman et al., 2011; Reimoser and Putman, 2011).

Here, we aim to (1) provide a synthetic review on the effects of ungulate density on the regeneration, structure and functioning of temperate forests, from both northern and southern hemispheres, (2) quantify the dose-response relationships between ungulate densities and forest attributes, and (3) identify potential thresholds and tipping points for each dose-response. Our study provides a first average global estimate of what ungulate densities may be critical for forest development in temperate zone. We then discuss ungulate management strategies and identify research priorities for animal-plant interactions in temperate forests.

## 2. Methods

We searched three literature data bases (CAB Abstract, Web of Science and Scopus) for scientific publications on the effects of wild ungulates on temperate deciduous forests. Although ungulate species are currently increasing in range and density especially in Central Europe and North America, ungulate species from other continents were included as well to have a wider overview. We used the following combination of keywords: “(mammals or mammal or mammalia or deer or mouflon or wild boar) and (forest or trees or forests) and (seed

dispersal or browsing or trampling or stripping or defecating or rooting or fraying) and (structure or species richness or abundance or functioning) and (temperate or seasonal or deciduous)”. We retrieved 469 articles, from which 164 studies were utilized for our analysis because they yielded information on the study region, ungulate species, ungulate abundance and effects on different forest response variables. All extracted information was organized in an Excel file by response variable. For each reported case, corresponding information on the type of effect, ungulate species, ungulate abundance/density and research area was incorporated.

### 2.1. Forest responses

All response effects were than grouped into three broad response attributes, each consisting of 2–3 similar forest components. This was done in order to have enough data points and statistical power for the analysis. The three forest response attributes consist of forest regeneration (i.e., the amount, diversity, and seedlings and saplings composition), forest structure (i.e., horizontal patch structure and vertical forest layering), and forest functioning (i.e., nutrient cycling in soil, tree growth and provision of wild forest food). Because studies differed widely in the temporal scale, survey design, environmental conditions, response variables used and measurements procedures, we quantified ungulate effects in a qualitative way mainly to enhance comparability across a large array of heterogeneous studies. For each response variable, it was evaluated whether ungulate presence or density had a significantly positive (1), significantly negative (–1), or no significant effect (0). In total, we compiled 435 cases.

### 2.2. Ungulate density

For each study, the ungulate species and density were recorded. To be able to assess the combined effects of different species, ungulate density was expressed as metabolic weight density (MWD) using the following formula:  $MWD = mass^{0.75}$  (Kleiber, 1947). This allowed the standardization of ungulate density based on their nutritional needs. A standard body mass (in kg) was used for the different ungulates species (Table 1). If a mix of ungulates was presented without identification we used 96 kg as an average for the most common *cervidae* species presented across the studies.

**Table 1**

Ungulate standard weight and representation of ungulate species from total number of reported species.

Common name	Species	Weight (kg)	%	Source
Buffalo	<i>Bison bison</i>	650	1,1	Lott and Galland (1987)
Moose	<i>Alces alces</i>	425	3,6	Stephenson et al. (1998)
Elk	<i>Cervus canadensis</i>	250	1,9	Cook et al. (2003)
Red deer	<i>Cervus elaphus</i>	190	11,2	Gill and Morgan (2010)
Sika deer	<i>Cervus nippon</i>	110	5,8	Suzuki et al. (2001)
Wild boar	<i>Sus scrofa</i>	80	9,4	Genov and Massei (2004)
Fallow deer	<i>Dama dama</i>	65	2,8	Gill and Morgan (2010)
White-tailed deer	<i>Odocoileus virginianus</i>	60	28,3	Hefley et al. (2013)
Chamois	<i>Rupicapra rupicapra</i>	35	1,1	Garcia-Gonzalez and Cuartas (1996)
Roe deer	<i>Capreolus capreolus</i>	25	7,1	De Jong et al. (1995)
Black-tailed deer	<i>Muntiacus reevesi</i>	18	2,6	Parker et al. (1993)
Pudu	<i>Pudu pudu</i>	12	1,1	Merino et al. (2005)
Ungulates	Mix of species	96	24	

### 2.3. Statistical analysis

We compared the relative frequency of response types (positive, neutral or negative) for each forest component with Chi-square tests. We used Multinomial Logistic Regression analysis to fit dose-response curves, in which ungulate density was expressed in terms of MWD, and ungulate effects were classified as being significantly positive (1), not significant (0) or significantly negative (−1). All model assumptions are met, such as: an independent continuous variable, a nominal continuous variable, independent observation, low multicollinearity and large sample size. The independent variable, MWD, was transformed to log<sub>10</sub>. These models have a sigmoid shape. To allow for unimodal responses to ungulate density, we evaluated whether adding the square of ungulate density significantly improved the model (Augustin et al., 2001; Merino et al., 2005). Dose-response curves were fitted for all three forest attribute responses (forest regeneration, forest structure and forest functioning) to increase power and obtain generalities, as well as for each of the individual 8 response variables to obtain more detailed insight. For each analysis, we calculated Chi<sup>2</sup>, the coefficient of determination, the probability value and the degrees of freedoms of the model. To identify critical ungulate densities that reflect tipping points in forest response, we quantified at what value of ungulate density the average response value as fitted by the curve became greater than 0.5 (from neutral to positive) or smaller than −0.5 (from neutral to negative).

### 3. Results

#### 3.1. Ungulate effects on forest regeneration, structure, and functioning

From the total of 435 cases compiled, 50.5% were situated in North America, 31.9% in Europe, 7.2% in Oceania, 6.8% in Asia and 3.6% in South America. Ungulates had a significantly negative effect on forest attributes on most (70%) of the cases. In 11% of the cases a significantly positive effect was reported, ranging from food production with 0% to horizontal structure with 22% (Fig. 1). In 19% of the cases no effect was

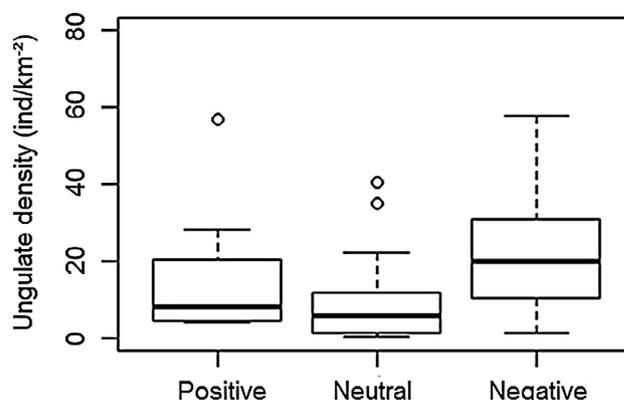


Fig. 2. Positive, negative and neutral effects of wild ungulates on different forest attributes are associated with different ungulate densities (Kruskal-Wallis test, Chi square = 16.78, P < 0.001). The reported number of cases are 20 for positive, 39 for neutral and 162 for negative. Boxplots of ungulate densities are shown and their significance test; positive vs. neutral P = 0.081, positive vs. negative P < 0.000, neutral vs. negative P < 0.000 (pairwise-Mann-Whitney U test). Horizontal “line” indicates the median. The “box” represents the middle 50% of scores for the group. The upper and lower “whiskers” represent scores outside the middle 50%. Outliers are plotted as individual “points”.

detected.

Negative effects of ungulates were most frequently reported for wild food provision (in 100% of 8 cases), on regeneration abundance (in 79% of 139 cases) and forest composition (in 78% of 108 cases). Positive effects were most frequently reported for horizontal forest structure (in 22% of 45 cases) and on nutrient cycling in soil (in 21% of 29 cases). Finally, ungulates had most often non-significant effects on timber production (in 31% of the cases) and nutrient cycling in soil (in 35% of the cases).

Ungulate densities varied from < 1 to > 300 km<sup>−2</sup> across studies. Average ungulate density was greater for studies that reported a negative effect on forest attributes (on average 28 km<sup>−2</sup>), than for studies

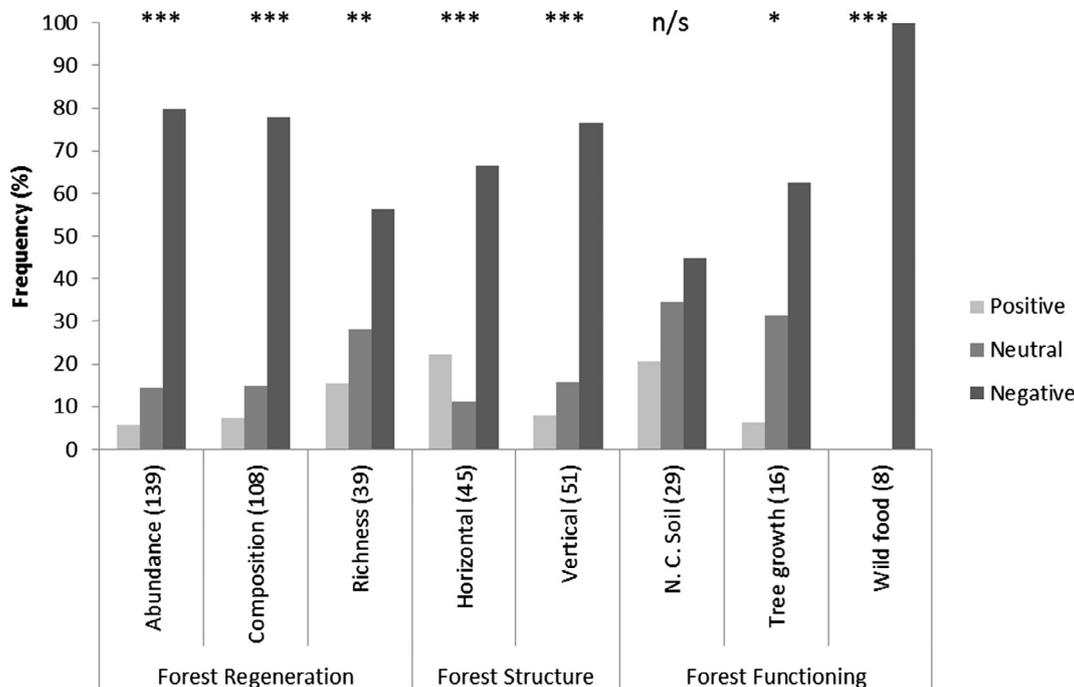


Fig. 1. Relative frequency of ungulate effects (significantly positive, neutral, or significantly negative) on forest regeneration (abundance, composition and richness), forest structure (horizontal and vertical) and forest functioning (nutrient cycling in soil and the provision of food and timber). The sample size is given in parentheses. Significant differences in the relative frequency of response types (positive, neutral or negative) within a forest component are tested with a Chi-square test (d.f. = 2) and indicated by asterisks. \*P < 0.05, \*\*P < 0.01, \*\*\*P < 0.001 and n.s. P > 0.05.

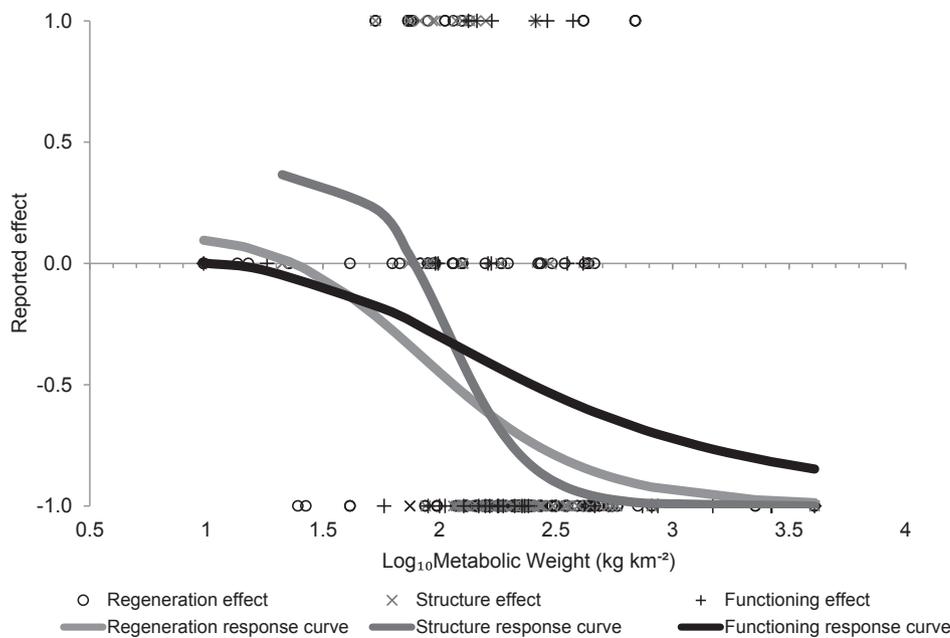


Fig. 3. Dose-response relationship between Metabolic Weight Density and the probability of studies reporting a significantly negative (−1), neutral (0) or significantly positive (1) effect on each forest attribute – Forest regeneration (N = 221 cases), forest structure (N = 84), forest functioning (N = 42) – summarized with multinomial logistic regression curves. Each of the forest attributes are shown with different shades of grey and different symbols.

that reported positive or neutral effects on the different forest attributes (on average 15 km<sup>-2</sup> and 11 km<sup>-2</sup> respectively) (Fig. 2).

### 3.2. Dose-response relationships

The dose-response curves (Fig. 3, Appendix A1) reported effects on forest attributes, which were more negative as ungulate abundance increased. This was the case for forest regeneration (Multinomial logistic regression,  $X^2 = 36.80$ ,  $R^2 = 0.20$ ,  $P < 0.001$ ,  $df = 2$ , Fig. 3), forest structure ( $X^2 = 24.88$ ,  $R^2 = 0.33$ ,  $P < 0.001$ ,  $df = 2$  Fig. 3), as well as forest functioning ( $X^2 = 7.84$ ,  $R^2 = 0.22$ ,  $P = 0.02$ ,  $df = 2$  Fig. 3). The dose-response curves were generally sigmoidal declining, and no unimodal curves were found. The thresholds, where ungulates tended to switch from generally neutral effects towards significantly negative effects was observed at a value of −0.5. The threshold was lowest ( $\text{Log}_{10}2.06$ , or 115 kg of metabolic weight km<sup>-2</sup>) for forest regeneration, the threshold occurred at intermediate ungulate density ( $\text{Log}_{10}2.15$  or 141 kg km<sup>-2</sup>) for forest structure and the threshold occurred at highest ungulate densities ( $\text{Log}_{10}2.40$  or 251 kg km<sup>-2</sup>) in the case of forest functioning. Across the evaluated studies, the average MWD was  $\text{Log}_{10}2.306$  or 202 kg km<sup>-2</sup>.

## 4. Discussion

How ungulates affect temperate forests is a major concern of forest and nature managers, especially in regions where ungulates are increasing in abundance (Pellerin et al., 2010). This study is the first semi-quantitative review on the effects of ungulate abundance on temperate forests at a global scale. Dose-response curves that we fitted to link reported effects of ungulates to ungulate density revealed that effects on a variety of forest characteristics changed from neutral to generally negative as ungulate density increased. There was no optimum density for a positive effect on forests, but rather a gradual decline with an increasing density, followed by a steep decline (inverted sigmoidal curve). As ungulate density increases, there is a rapid shift from neutral to negative effects on the regeneration, structure and functioning of forests around a MWD of 169 kg km<sup>-2</sup>, which is equivalent to 15 roe deer per km<sup>2</sup> or 8 white-tailed deer per km<sup>2</sup>.

### 4.1. Ungulate effects on forests

Negative effects were clearly more frequent than positive and neutral effects (Fig. 1). From all response variables, ungulates had the strongest negative effect on food production (in 100% of the cases), although the sample is too small (N = 8) to draw strong conclusions. Regeneration abundance and forest composition decreased in 79% and 78% of the cases, as selective browsing reduced the abundance of palatable species and increased the abundance of less palatable species through competitive release (Schippers et al., 2014).

Ungulates had the strongest positive effect on horizontal forest structure (in 22% of the cases). This is explained by the fact that ungulates create horizontal heterogeneity by generating open areas by vegetation removal and by dispersing seeds, thus promoting tree species distribution across the forest (Eycott et al., 2007). Nutrient cycling in soil increased (in 21% of the cases), as ungulates increase nutrient availability and heterogeneity through defecation and litter removal from the forest floor (Murray et al., 2014).

Ungulates had most often neutral, i.e., non-significant effects on timber production (31% of the cases) and nutrient cycling in soil (in 35% of the cases), probably because a larger time span (> 50 years) is needed for ungulates to affect wood production (Welch and Scott, 1998) and the entire trophic levels in order to change local soil conditions.

### 4.2. Thresholds: Ungulate densities

Our results indicate that ungulates negatively affect regeneration in temperate forests beyond densities of 115 kg km<sup>-2</sup> (MWD), which is equivalent to 2.3 red deer, 4.3 wild boar, 5.3 white-tailed deer or 10.2 roe deer per km<sup>2</sup>. The high apparent sensitivity of regeneration to ungulates is probably because seedlings and saplings are relatively vulnerable to intermediate browsing and trampling, and because seedlings and saplings are highly dynamic and respond rapidly to environmental change. Our results are similar to the threshold density of 4 red deer suggested in a Scottish forest study (Putman et al., 2011), but below the suggested threshold density of 14 deer per km<sup>2</sup> for a mix of deer species (roe, muntjac and fallow deer) in a lowland UK forest (Gill and Morgan, 2010). This variation can be explained by a combination of primary productivity and ungulate composition. Our estimates are not representative for one particular area, but represent the general pattern

across all temperate forests globally.

We found that ungulates negatively affect forest structure beyond a MWD of  $141 \text{ kg km}^{-2}$ , which is equivalent to 2.7 red deer, 5.2 wild boar, 6.5 white-tailed deer or 12.5 roe deer. The fact that the threshold is higher than forest regeneration reflects that stem density and development of less palatable species is relatively insensitive to ungulates. Our estimates are similar to stem density thresholds reported in earlier studies, e.g., 8 white-tailed deer per  $\text{km}^2$  in a Pennsylvanian forest (Horsley et al., 2003) and 4–8 white-tailed deer in a Wisconsin forest (Alverson et al., 1988).

Finally, forest functioning was negatively affected beyond a MWD of  $251 \text{ kg km}^{-2}$ , which is equivalent to 4.9 red deer, 9.4 wild boar, 11.6 white-tailed deer or 22.5 roe deer. The threshold is again higher for forest functioning, probably because browsing limits regeneration growth and reproduction. These threshold estimates are similar to 4 red deer per  $\text{km}^2$  in a British forest (Holloway, 1967) and 7–15 white-tailed deer per  $\text{km}^2$  in a Pennsylvanian forest (Tilghman, 1989), but much higher than 4–12 roe deer reported in a British commercial forest (Reimoser and Putman, 2011). This variation can be explained by different methods used to estimate ungulate density and their effect on vegetation, as well as other factors that mediate these interactions such as primary productivity. For example, more fertile soils and higher primary productivity allow for higher ungulate densities (Cromsigt and Kuijper, 2011).

Our multinomial logistical regression analysis was only based on whether a significant ungulate effect (positive or negative) was detected in the original studies, and not based on effect sizes. This allowed us to include a greater number of studies with a variety of methodologies. Future research could study effect sizes, to get a more precise estimate of ungulate impacts.

Although our data came from temperate forest studies from all continents, most cases (82%) came from North American and European forests, reflecting the common bias in research efforts. Similarly, due to our search criteria, we recognise that our study has bias towards deer species, relegating other wild ungulate species, such as: moose and elk.

We acknowledge that not all temperate forests respond equally to different ungulates densities and composition, hence that a single dose-response relationship cannot represent the reality of all temperate deciduous forests. We used metabolic weight density rather than absolute density, taking the metabolic properties of these different species into account. We acknowledge that such an approach is still not perfect because ungulate species also have different behaviour, that is not entirely explained by their metabolic weight. The tipping points we have identified are likely to be higher in more fertile systems, and lower in more infertile systems (Cromsigt and Kuijper, 2011).

#### 4.3. Managing ungulate effects on forests; towards an optimal ungulate density

Our results suggest that switching from negative to neutral or even positive ungulate effects it is necessary to reduce ungulate densities, at least temporarily. Such a release of plants from natural enemies provides a time window of opportunity for tree regeneration (Nuttall et al., 2014). Once saplings attain a minimum size (e.g., 1.6 or 2.0 m in height, depending on the particular ungulate assemblage) they can escape herbivore pressure, because their top shoot is out of ungulate reach, and because plants are sufficiently robust to recover from herbivory (Motta, 2003; Renaud et al., 2003) and fraying. The time window needed for forest recovery depends on seed availability and tree species growth rate. Hence, for lowland central Europe it will be relatively short (i.e., 3 years) for fast-growing light-demanding species (e.g., *Betula pendula* and *Pinus sylvestris*), the time window of recovery

will be relatively long (i.e., 10 years) for slow-growing and palatable species (*Quercus robur*, *Quercus petraea*, and *Tilia cordata*). Whereas if the system experiences neutral or even positive effects from ungulates, it is expected that the three main plant defence mechanisms (resistance, tolerance and escape) are enough to secure their establishment and reproduction (Lindroth and St. Clair, 2013).

Forest managers and stakeholders can enhance forest heterogeneity at different spatiotemporal scales by rotating ungulate exclosures at different frequencies (e.g., 3–10 years) and sizes (e.g., 0.1–1 ha) (Augustine and Frelich, 1998), by offering alternative foliage for ungulates, changing landscape structure such as shelter provision or even by modifying the forest edge (Reimoser and Putman, 2011). Forest managers can also steer the landscape of fear by reintroducing top predators (Sergio et al., 2008), varying hunting pressure over space and time (Cromsigt et al., 2013), or managing ungulate sex ratio within a population, for instance limiting the number of reproducing females (Clutton-Brock et al., 2004).

Rewilding natural areas with apex predators, might bring ungulates density into balance again and thus secure regeneration establishment and development. Predators can eventually restore top-down interactions and associated cascading effects that culling and game hunting have failed to achieve. On natural areas trophic rewilding gives the opportunity of self-regulation, thus providing new opportunities for the entire ecological community and reducing human interference (Svenning et al., 2015).

#### 4.4. Research priorities and outlook

We used qualitative data to characterize the responses from a wide variety of temperate forests and showed a non-linear relationship between ungulate density and forest regeneration (Gill and Morgan, 2010). However, the coefficient of determination was rather low, indicating a high variability among independent and dependent variables. To better understand how ungulate density affects forest development, the dose-response curves should be produced more accurately. Our study is a first important step based on data from literature review, but results were combined from studies in many different forest systems, which used different methods to quantify ungulate densities and forest responses. Modern techniques, such as surveys with camera traps, can provide more accurate and more standardized ungulate density estimates, detecting animals with a wide range in body sizes and different levels of shyness, 24 hrs per day (Rowcliffe et al., 2011).

### 5. Conclusions

Understanding effects of wild ungulates is essential to the management and conservation of temperate forests. Much of the recent debate on ungulate management has focused on the carrying capacity of forests from an animal perspective. However, for managers, it is also important to identify how many ungulates a forest can handle before forest regeneration is compromised. Our results suggest a crucial role of ungulate density in forest regeneration, structure, and functioning. At low ungulate density forest regeneration, structure and functioning are maintained, whereas at intermediate to high ungulate densities the different forest attributes are negatively affected.

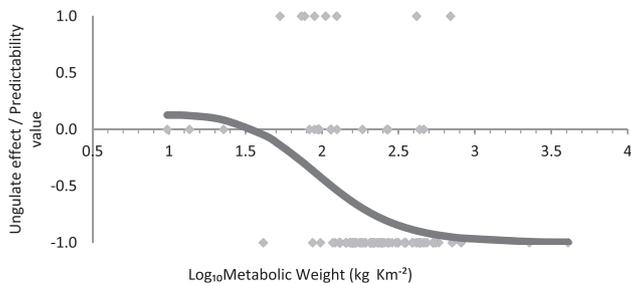
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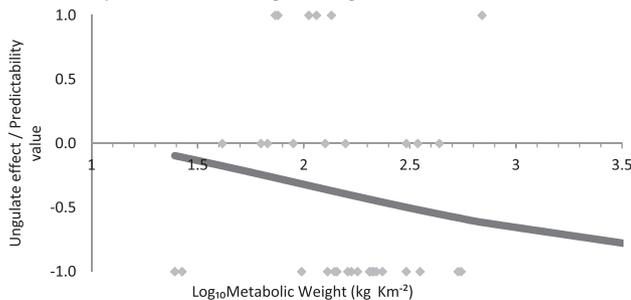
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A.1 Multinomial regression curves for forest regeneration

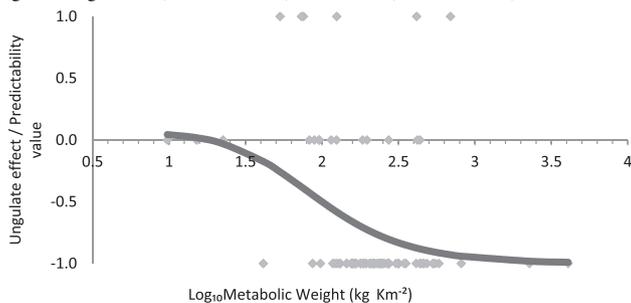
Abundance refers to the amount of seedling, saplings and poles in the forest. Multinomial logistic regression,  $X^2 = 21.33$ ,  $R^2 = 0.25$ ,  $P < 0.001$ ,  $df = 2$ .



Diversity refers to species richness and diversity. Multinomial logistic regression,  $X^2 = 1.54$ ,  $R^2 = 0.05$ ,  $P < 0.463$ ,  $df = 2$ .

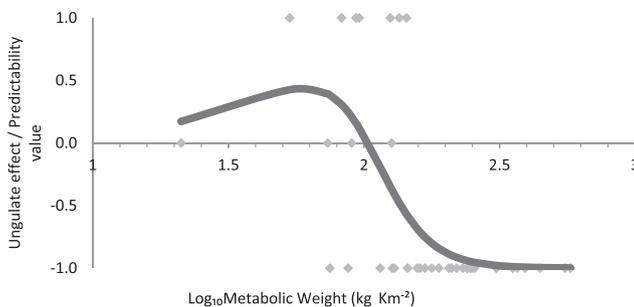


Composition refers to the proportion of a group of tree species compared to another group of tree species. Literature often compares conifer and broadleaved proportions. Multinomial logistic regression,  $X^2 = 15.74$ ,  $R^2 = 0.23$ ,  $P < 0.001$ ,  $df = 2$ .

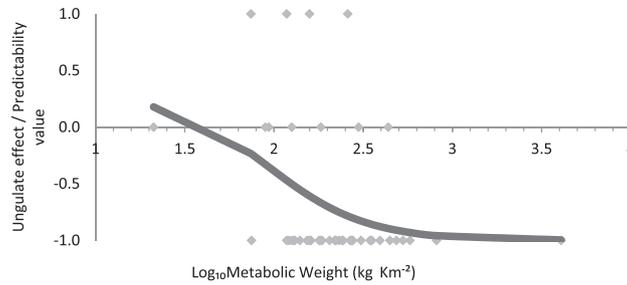


A.2. Multinomial regression curves for forest structure

Horizontal structure refers to the range and distribution of tree species in a forest. Multinomial logistic regression,  $X^2 = 22.66$ ,  $R^2 = 0.57$ ,  $P < 0.001$ ,  $df = 2$ .

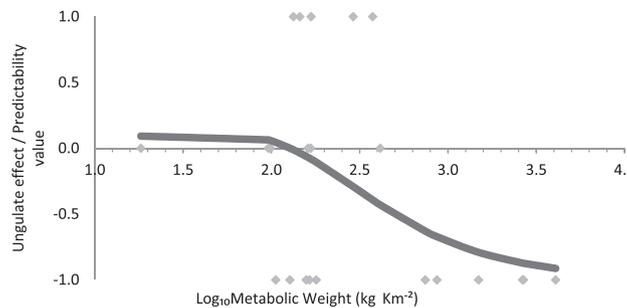


Vertical structure refers to the vertical layering of the forest, which includes undergrowth, understory, canopy and emergent layer. Multinomial logistic regression,  $X^2 = 7.28$ ,  $R^2 = 0.20$ ,  $P = 0.026$ ,  $df = 2$ .

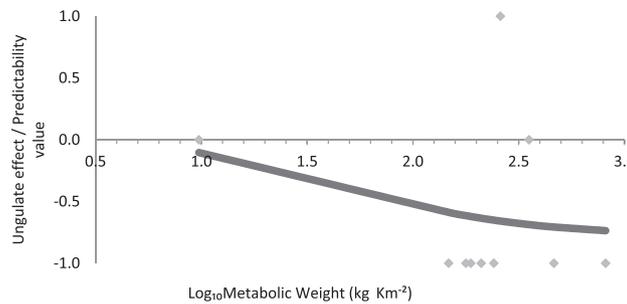


**A.3. Multinomial regression curves for forest functioning**

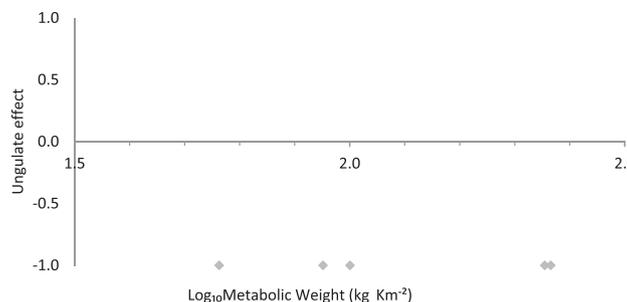
Nutrient cycling in soil refers to disruption of the processes related to the formation of soil and nutrient availability. Multinomial logistic regression,  $X^2 = 9.57$ ,  $R^2 = 0.39$   $P = 0.008$ ,  $df = 2$ .



Tree growth refers mainly to the amount of woody biomass produced by trees. Multinomial logistic regression,  $X^2 = 2.55$ ,  $R^2 = 0.28$   $P = 0.279$ ,  $df = 2$ .



Wild food refers to the provision of wild food on forests, example fruits and nuts. Predicted regression curve does not apply due to insufficient sample size.



**B**

**B.1. List of papers reviewed**

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