## 11

# Analytical Methods for Monitoring Biodegradation Processes of Environmentally Degradable Polymers

Maarten van der Zee

# 11.1 Introduction

This chapter presents an overview of the current knowledge on experimental methods for monitoring the biodegradability of polymeric materials. The focus is, in particular, on the biodegradation of materials under environmental conditions. Examples of *in vivo* degradation of polymers used in biomedical applications are not covered in detail but have been extensively reviewed elsewhere, e.g., [1–3]. Nevertheless, it is good to realize that the same principles of the methods for monitoring biodegradability of environmental polymers are also used for the evaluation of the degradation behavior of biomedical polymers.

A number of different aspects of assessing the potential, the rate, and the degree of biodegradation of polymeric materials are discussed. The mechanisms of polymer degradation and erosion receive attention and factors affecting enzymatic and nonenzymatic degradation are briefly addressed. Particular attention is given to the various ways for measuring biodegradation, including complete mineralization to gasses (such as carbon dioxide and methane), water, and possibly microbial biomass. Finally, some general conclusions are presented with respect to measuring biodegradability of polymeric materials.

# 11.2 Some Background

There is a worldwide research effort to develop biodegradable polymers for agricultural applications or as a waste management option for polymers in the environment. Until the end of the 20th century, most of the efforts were synthesis

oriented, and not much attention was paid to the identification of environmental requirements for, and testing of, biodegradable polymers. Consequently, many unsubstantiated claims to biodegradability were made, and this has damaged the general acceptance.

An important factor is that the term biodegradation has not been applied consistently. In the medical field of sutures, bone reconstruction, and drug delivery, the term biodegradation has been used to indicate degradation into macromolecules that stay in the body but migrate (e.g., UHMW polyethylene from joint prostheses), or hydrolysis into low-molecular-weight molecules that are excreted from the body (bioresorption), or dissolving without modification of the molecular weight (bioabsorption) [4, 5]. On the other hand, for environmentally degradable plastics, the term biodegradation may mean fragmentation, loss of mechanical properties, or sometimes degradation through the action of living organisms [6]. Deterioration or loss in physical integrity is also often mistaken for biodegradation [7]. Nevertheless, it is essential to have a universally acceptable definition of biodegradability to avoid confusion as to where biodegradable polymers can be used in agriculture or fit into the overall plan of polymer waste management. Many groups and organizations have endeavored to clearly define the terms "degradation," "biodegradation," and "biodegradability." But there are several reasons why establishing a single definition among the international communities has not been straightforward, including:

- 1) the variability of an intended definition given the different environments in which the material is to be introduced and its related impact on those environments.
- 2) the differences of opinion with respect to the scientific approach or reference points used for determining biodegradability,
- the divergence of opinion concerning the policy implications of various definitions, and
- 4) challenges posed by language differences around the world.

As a result, many different definitions have officially been adopted, depending on the background of the defining organization and their particular interests. However, of more practical importance are the criteria for calling a material "biodegradable." A demonstrated potential of a material to biodegrade does not say anything about the time frame in which this occurs, nor the ultimate degree of degradation. The complexity of this issue is illustrated by the following common examples.

Low-density polyethylene has been shown to biodegrade slowly to carbon dioxide (0.35% in 2.5 years) [8], and according to some definitions can thus be called a biodegradable polymer. However, the degradation process is so slow in comparison with the application rate that accumulation in the environment will occur. The same applies for polyolefin–starch blends which rapidly loose strength, disintegrate, and visually disappear if exposed to microorganisms [9–11]. This is due to

utilization of the starch component, but the polyolefin fraction will nevertheless persist in the environment. Can these materials be called "biodegradable"?

# 11.3 **Defining Biodegradability**

In 1992, an international workshop on biodegradability was organized to bring together experts from around the world to achieve areas of agreement on definitions, standards, and testing methodologies. Participants came from manufacturers, legislative authorities, testing laboratories, environmentalists, and standardization organizations in Europe, United States, and Japan. Since this fruitful meeting, there is a general agreement concerning the following key points [12].

- 1) For all practical purposes of applying a definition, material manufactured to be biodegradable must relate to a specific disposal pathway such as composting, sewage treatment, denitrification, and anaerobic sludge treatment.
- The rate of degradation of a material manufactured to be biodegradable has to be consistent with the disposal method and other components of the pathway into which it is introduced, such that accumulation is controlled.
- 3) The ultimate end products of aerobic biodegradation of a material manufactured to be biodegradable are CO<sub>2</sub>, water, and minerals and that the intermediate products include biomass and humic materials. (Anaerobic biodegradation was discussed in less detail by the participants.)
- Materials must biodegrade safely and not negatively impact the disposal process or the use of the end product of the disposal.

As a result, specified periods of time, specific disposal pathways, and standard test methodologies were incorporated into definitions. Standardization organizations such as CEN, ISO, and ASTM were consequently encouraged to rapidly develop standard biodegradation tests so these could be determined. Society further demanded nondebatable criteria for the evaluation of the suitability of polymeric materials for disposal in specific waste streams such as composting or anaerobic digestion. Biodegradability is usually just one of the essential criteria, besides ecotoxicity, effects on waste treatment processes, etc.

In the following sections, biodegradation of polymeric materials is looked upon form the chemical perspective. The chemistry of the key degradation process is represented by Eq. (11.1) and (11.2), where C<sub>polymer</sub> represents either a polymer or a fragment from any of the degradation processes defined earlier. For simplicity here, the polymer or fragment is considered to be composed only of carbon, hydrogen, and oxygen; other elements may, of course, be incorporated in the polymer, and these would appear in an oxidized or reduced form after biodegradation depending on whether the conditions are aerobic or anaerobic, respectively.

Aerobic biodegradation:

$$C_{\text{polymer}} + O_2 \rightarrow CO_2 + H_2O + C_{\text{residue}} + C_{\text{biomass}}$$
 (11.1)

Anaerobic biodegradation:

$$C_{\text{polymer}} \rightarrow CO_2 + CH_4 + H_2O + C_{\text{residue}} + C_{\text{biomass}}$$
 (11.2)

Complete biodegradation occurs when no residue remains, and complete mineralization is established when the original substrate, C<sub>polymer</sub> in this example, is completely converted into gaseous products and salts. However, mineralization is a very slow process under natural conditions because some of the polymer undergoing biodegradation will initially be turned into biomass [13, 14]. Therefore, complete biodegradation, and not mineralization, is the measurable goal when assessing removal from the environment.

# 11.4 Mechanisms of Polymer Degradation

When working with biodegradable materials, the obvious question is why some polymers biodegrade and others do not. To understand this, one needs to know about the mechanisms through which polymeric materials are biodegraded. Although biodegradation is usually defined as degradation caused by biological activity (especially enzymatic action), it will usually occur simultaneously with – and is sometimes even initiated by - abiotic degradation such as photodegradation and simple hydrolysis. The following paragraphs give a brief introduction about the most important mechanisms of polymer degradation.

#### 11.4.1

## Nonbiological Degradation of Polymers

A great number of polymers is subject to hydrolysis, such as polyesters, polyanhydrides, polyamides, polycarbonates, polyurethanes, polyureas, polyacetals, and polyorthoesters. Different mechanisms of hydrolysis have been extensively reviewed not only for backbone hydrolysis but also for the hydrolysis of pendant groups [15-17]. The necessary elements for a wide range of catalysis, such as acids and bases, cations, nucleophiles and micellar, and phase transfer agents are usually present in most environments. In contrast to enzymatic degradation, where a material is degraded gradually from the surface inward (primarily because macromolecular enzymes cannot diffuse into the interior of the material), chemical hydrolysis of a solid material can take place throughout its cross section except for few hydrophobic polymers.

Important features affecting chemical polymer degradation and erosion include (i) the type of chemical bond, (ii) the pH, (iii) the temperature, (iv) the copolymer composition, and (v) water uptake (hydrophilicity). These features will not be discussed here, but have been covered in detail by Göpferich [4].

#### 11.4.2

## **Biological Degradation of Polymers**

Polymers represent major constituents of the living cells which are most important for the metabolism (enzyme proteins and storage compounds), the genetic information (nucleic acids), and the structure (cell wall constituents and proteins) of cells [18]. These polymers have to be degraded inside cells in order to be available for environmental changes and to other organisms upon cell lysis. It is therefore not surprising that organisms, during many millions of years of adaptation, have developed various mechanisms to degrade naturally occurring polymers. For the many different new synthetic polymers that have found their way into the environment only in the last 70 years, however, these mechanisms may not as yet have been developed.

There are many different degradation mechanisms that combine synergistically in nature to degrade polymers. Microbiological degradation can take place through the action of enzymes or by-products (such as acids and peroxides) secreted by microorganisms (bacteria, yeasts, fungi, etc.). Also macroorganisms can eat and, sometimes, digest polymers and cause mechanical, chemical, or enzymatic aging [19, 20].

Two key steps occur in the microbial polymer degradation process: first, a depolymerization or chain cleavage step, and second, mineralization. The first step normally occurs outside the organism due to the size of the polymer chain and the insoluble nature of many of the polymers. Extracellular enzymes are responsible for this step, acting either endo (random cleavage on the internal linkages of the polymer chains) or exo (sequential cleavage on the terminal monomer units in the main chain).

Once sufficiently small-size oligomeric or monomeric fragments are formed, they are transported into the cell where they are mineralized. At this stage, the cell usually derives metabolic energy from the mineralization process. The products of this process, apart from ATP, are gasses (e.g., CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>, and H<sub>2</sub>), water, salts and minerals, and biomass. Many variations of this general view of the biodegradation process can occur, depending on the polymer, the organisms, and the environment. Nevertheless, there will always be, at one stage or another, the involvement of enzymes.

# 11.5 Measuring Biodegradation of Polymers

As can be imagined from the various mechanisms described above, biodegradation does not only depend on the chemistry of the polymer but also on the presence of the biological systems involved in the process. When investigating the

		(1)	(2)
		aquatic	high solids
a)	aerobic	aerobic wastewater treatment plants	surface soils
		surface waters, e.g., lakes and rivers	organic waste composting plants
		marine environments	littering
b)	anaerobic	anaerobic wastewater	deep sea sediments
		treatment plants	anaerobic sludge
		rumen of herbivores	anaerobic digestion/ biogasification
			landfill

Figure 11.1 Schematic classification of different biodegradation environments for polymers.

biodegradability of a material, the effect of the environment cannot be neglected. Microbial activity and hence biodegradation is influenced by

- 1) the presence of microorganisms
- 2) the availability of oxygen
- 3) the amount of available water
- 4) the temperature
- 5) the chemical environment (pH, electrolytes, etc.).

In order to simplify the overall picture, the environments in which biodegradation occurs are basically divided in two environments: (a) aerobic (with oxygen available) and (b) anaerobic (no oxygen present). These two in turn can be subdivided into (1) aquatic and (2) high-solids environments. Figure 11.1 schematically presents the different environments, with examples in which biodegradation may occur [21, 22].

The high-solids environments will be the most relevant for measuring environmental biodegradation of polymeric materials, since they represent the conditions during biological municipal solid waste treatment, such as composting or anaerobic digestion (biogasification). However, possible applications of biodegradable materials other than in packaging and consumer products, for example, in fishing nets at sea, or undesirable exposure in the environment due to littering, explain the necessity of aquatic biodegradation tests.

Numerous ways for the experimental assessment of polymer biodegradability have been described in the scientific literature. Because of slightly different definitions or interpretations of the term "biodegradability," the different approaches are therefore not equivalent in terms of information they provide or the practical significance. Since the typical exposure to environment involves incubation of a polymer substrate with microorganisms or enzymes, only a limited number of

measurements are possible: those pertaining to the substrates, to the microorganisms, or to the reaction products. Four common approaches available for studying biodegradation processes have been reviewed in detail by Andrady [13, 14]:

- 1) monitoring accumulation of biomass
- 2) monitoring the depletion of substrates
- 3) monitoring reaction products
- 4) monitoring changes in substrate properties.

In the following sections, different test methods for the assessment of polymer biodegradability are presented. Measurements are usually based on one of the four approaches given above, but combinations also occur. Before choosing an assay to simulate environmental effects in an accelerated manner, it is critical to consider the closeness of fit that the assay will provide between substrate, microorganisms, or enzymes, and the application or environment in which biodegradation should take place [23].

### 11.5.1

## **Enzyme Assays**

## 11.5.1.1 **Principle**

In enzyme assays, the polymer substrate is added to a buffered or pH-controlled system, containing one or several types of purified enzymes. These assays are very useful in examining the kinetics of depolymerization, or oligomer or monomer release from a polymer chain under different assay conditions. The method is very rapid (minutes to hours) and can give quantitative information. However, mineralization rates cannot be determined with enzyme assays.

## 11.5.1.2 Applications

The type of enzyme to be used, and quantification of degradation, will depend on the polymer being screened. For example, Mochizuki et al. [24] studied the effects of draw ratio of polycaprolactone fibers on enzymatic hydrolysis by lipase. Degradability of PCL fibers was monitored by dissolved organic carbon (DOC) formation and weight loss. Similar systems with lipases have been used for studying the hydrolysis of broad ranges of aliphatic polyesters [25-30], copolyesters with aromatic segments [26, 31-33], and copolyesteramides [34, 35]. Other enzymes such as  $\alpha$ -chymotrypsin and  $\alpha$ -trypsin have also been applied for these polymers [36, 37]. Biodegradability of poly(vinyl alcohol) segments with respect to block length and stereochemical configuration has been studied using isolated poly(vinyl alcohol)-dehydrogenase [38]. Cellulolytic enzymes have been used to study the biodegradability of cellulose ester derivatives as a function of degree of substitution and the substituent size [39]. Similar work has been performed with starch esters using amylolytic enzymes such as α-amylases, β-amylases, glucoamylases, and amyloglucosidases [40]. Enzymatic methods have also been used to study the biodegradability of starch plastics or packaging materials containing cellulose [41-46].

#### 11.5.1.3 **Drawbacks**

Caution must be taken in extrapolating enzyme assays as a screening tool for different polymers since the enzymes have been paired to only one polymer. The initially selected enzymes may show significantly reduced activity toward modified polymers or different materials, even though more suitable enzymes may exist in the environment. Caution must also be taken if the enzymes are not purified or appropriately stabilized or stored, since inhibitors and loss of enzyme activity can occur [23].

## 11.5.2

#### **Plate Tests**

## 11.5.2.1 Principle

Plate tests have initially been developed in order to assess the resistance of plastics to microbial degradation. Several methods have been standardized by standardization organizations such as the ASTM and the ISO [47–49]. They are now also used to see if a polymeric material will support growth [23, 50]. The principle of the method involves placing the test material on the surface of a mineral salts agar in a petri dish containing no additional carbon source. The test material and agar surface are sprayed with a standardized mixed inoculum of known bacteria and/or fungi. The test material is examined after a predetermined incubation period at constant temperature for the amount of growth on its surface and the rating is given.

## 11.5.2.2 Applications

Potts [51] used the method in his screening of 31 commercially available polymers for biodegradability. Other studies where the growth of either mixed or pure cultures of microorganisms is taken to be indicative for biodegradation have been reported [6]. The validity of this type of test and the use of visual assessment alone have been questioned by Seal and Pantke [52] for all plastics. They recommended that mechanical properties should be assessed to support visual observations. Microscopic examination of the surface can also give additional information.

A variation of the plate test is the "clear zone" technique [53], sometimes used to screen polymers for biodegradability. A fine suspension of polymer is placed in an agar gel as the sole carbon source, and the test inoculum is placed in wells bored in the agar. After incubation, a clear zone around the well, detected visually or instrumentally, is indicative of utilization of the polymer. The method has, for example, been used in the case of starch plastics [54], various polyesters [55–57], and polyurethanes [58].

## 11.5.2.3 Drawbacks

A positive result in an agar plate test indicates that an organism can grow on the substrate, but does not mean that the polymer is biodegradable, since growth may appear on contaminants, plasticizers present, oligomeric fractions still present in the polymer, and so on. Therefore, these tests should be treated with caution when extrapolating the data to field situations.

## 11.5.3

# **Respiration Tests**

# 11.5.3.1 Principle

Aerobic microbial activity is typically characterized by the utilization of oxygen. Aerobic biodegradation requires oxygen for the oxidation of compounds to its mineral constituents such as CO<sub>2</sub>, H<sub>2</sub>O, SO<sub>2</sub>, P<sub>2</sub>O<sub>5</sub>, etc. The amount of oxygen utilized during incubation, also called the biochemical (or biological) oxygen demand (BOD), is therefore a measure of the degree of biodegradation. Several test methods are based on measurement of the BOD, often expressed as a percentage of the theoretical oxygen demand (TOD) of the compound. The TOD, which is the theoretical amount of oxygen necessary for completely oxidizing a substrate to its mineral constituents, can be calculated by considering the elemental composition and the stoichiometry of oxidation [13, 59–62] or based on experimental determination of the chemical oxygen demand (COD) [13, 63].

# 11.5.3.2 Applications

The closed bottle BOD tests were designed to determine the biodegradability of detergents [61, 64]. These have stringent conditions due to the low level of inoculum (in the order of 10<sup>5</sup> microorganisms/L) and the limited amount of test substance that can be added (normally between 2 and 4 mg/L). These limitations originate from the practical requirement that the oxygen demand should not be more than half the maximum dissolved oxygen level in water at the temperature of the test, to avoid the generation of anaerobic conditions during incubation.

For nonsoluble materials such as polymers, less stringent conditions are necessary and alternative ways for measuring BOD were developed. Two-phase (semi) closed bottle tests provide higher oxygen content in the flasks and permit a higher inoculum level. Higher test concentrations are also possible, encouraging higher accuracy with directly weighing in of samples. The oxygen demand can alternatively be determined by periodically measuring the oxygen concentration in the aquatic phase by opening the flasks [60, 65–67], by measuring the change in volume or pressure in incubation flasks containing CO<sub>2</sub>-absorbing agents [59, 68, 69], or by measuring the quantity of oxygen produced (electrolytically) to maintain constant gas volume/pressure in specialized respirometers [59, 62, 65, 66, 68].

### 11.5.3.3 Suitability

BOD tests are relatively simple to perform and sensitive, and are therefore often used as screening tests. However, the measurement of oxygen consumption is a nonspecific, indirect measure for biodegradation, and it is not suitable for determining anaerobic degradation. The requirement for test materials to be the sole carbon/energy source for microorganisms in the incubation media eliminates the use of oxygen measurements in complex natural environments.

11.5.4

## Gas (CO<sub>2</sub> or CH<sub>4</sub>) Evolution Tests

## 11.5.4.1 Principle

The evolution of carbon dioxide or methane from a substrate represents a direct parameter for mineralization. Therefore, gas evolution tests can be important tools in the determination of biodegradability of polymeric materials. A number of well-known test methods have been standardized for aerobic biodegradation, such as the (modified) Sturm test [70–75] and the laboratory-controlled composting test [76–79], as well as for anaerobic biodegradation, such as the anaerobic sludge test [80, 81] and the anaerobic digestion test [82, 83]. Although the principles of these test methods are the same, they may differ in medium composition, inoculum, the way substrates are introduced, and in the technique for measuring gas evolution.

# 11.5.4.2 Applications

Anaerobic tests generally follow biodegradation by measuring the increase in pressure and/or volume due to gas evolution, usually in combination with gas chromatographic analysis of the gas phase [84, 85]. Most aerobic standard tests apply continuous aeration; the exit stream of air can be directly analyzed continuously using a carbon dioxide monitor (usually infrared detectors) or titrimetrically after sorption in dilute alkali. The cumulative amount of carbon dioxide generated, expressed as a percentage of the theoretically expected value for total conversion to  $CO_2$ , is a measure for the extent of mineralization achieved. A value of 60% carbon conversion to  $CO_2$ , achieved within 28 days, is generally taken to indicate ready degradability. Taking into account that in this system there will also be incorporation of carbon into the formation of biomass (growth), the 60% value for  $CO_2$  implies almost complete degradation. While this criterion is meant for water-soluble substrates, it is probably applicable to very finely divided moderately degradable polymeric materials as well [13]. Nevertheless, most standards for determining biodegradability of plastics consider a maximum test duration of 6 months.

Besides the continuously aerated systems, described above, several static respirometers have been described. Bartha and Yabannavar [86] describe a two-flask system; one flask, containing a mixture of soil and the substrate, is connected to another chamber holding a quantity of carbon dioxide sorbant. Care must be taken to ensure that enough oxygen is available in the flask for biodegradation. Nevertheless, this experimental setup and modified versions thereof have been successfully applied in the assessment of biodegradability of polymer films and food packaging materials [87–89].

The percentage of carbon converted to biomass instead of carbon dioxide depends on the type of polymer and the phase of degradation. Therefore, it has been suggested to regard the complete carbon balance to determine the degree of degradation [90]. This implies that besides the detection of gaseous carbon, also the amount of carbon in soluble and solid products needs to be determined. Soluble products, oligomers of different molecular size, intermediates, and proteins secreted from microbial cells can be measured as COD or as DOC. Solid

products, biomass, and polymer remnants require a combination of procedures to separate and detect different fractions. The protein content of the insoluble fraction is usually determined to estimate the amount of carbon converted to biomass, using the assumptions that dry biomass consists of 50% protein, and that the carbon content of dry biomass is 50% [90–92].

# 11.5.4.3 Suitability

Gas evolution tests are popular test methods because they are relatively simple to perform and sensitive. A direct measure for mineralization is determined, and water-soluble or -insoluble polymers can be tested as films, powders, or objects. Furthermore, the test conditions and inoculum can be adjusted to fit the application or environment in which biodegradation should take place. Aquatic synthetic media are usually used, but also natural sea water [93, 94] or soil samples [86, 88, 89, 95] can be applied as biodegradation environments. A prerequisite for these media is that the background CO<sub>2</sub> evolution is limited, which excludes the application of real composting conditions. Biodegradation under composting conditions is therefore measured using an inoculum derived from matured compost with low respiration activity [76–78, 96, 97].

A drawback of using complex degradation environments such as mature compost is that simultaneous characterization of intermediate degradation products of determination of the carbon balance is difficult due to the presence of a great number of interfering compounds. To overcome this, an alternative test was developed based on an inoculated mineral bed-based matrix [98, 99].

## 11.5.5

## Radioactively Labeled Polymers

## 11.5.5.1 Principle and Applications

Some materials tend to degrade very slowly under stringent test conditions without an additional source of carbon. However, if readily available sources of carbon are added, it becomes impossible to tell how much of the evolved carbon dioxide can be attributed to the decomposition of the plastic. The incorporation of radioactive  $^{14}\mathrm{C}$  in synthetic polymers gives a means of distinguishing between  $\mathrm{CO}_2$  or  $\mathrm{CH}_4$  produced by the metabolism of the polymer, and that generated by other carbon sources in the test environment. By comparison of the amount of radioactive  $^{14}\mathrm{CO}_2$  or  $^{14}\mathrm{CH}_4$  with the original radioactivity of the labeled polymer, it is possible to determine the percent by weight of carbon in the polymer which was mineralized during the duration of the exposure [51, 100–102]. Collection of radioactively labeled gasses or low-molecular-weight products can also provide extremely sensitive and reproducible methods to assess the degradation of polymers with low susceptibility to enzymes, such as polyethylene [8, 103] and cellulose acetates [104, 105].

### 11.5.5.2 Drawbacks

Problems with handling the radioactively labeled materials and their disposal are issues on the down side to this method. In addition, in some cases, it is difficult

to synthesize the target polymer with the radioactive labels in the appropriate locations, with representative molecular weights, or with representative morphological characteristics.

#### 11.5.6

## Laboratory-Scale Simulated Accelerating Environments

#### 11.5.6.1 **Principle**

Biodegradation of a polymer material is usually associated with changes in the physical, chemical, and mechanical properties of the material. It is indeed these changes, rather than the chemical reactions, which make the biodegradation process so interesting from an application point of view. These useful properties might be measured as a function of the duration of exposure to a biotic medium, to follow the consequences of the biodegradation process on material properties. The biotic media can be specifically designed in a laboratory scale as to mimic natural systems but with a maximum control of variables such as temperature, pH, microbial community, mechanical agitation, and supply of oxygen. Regulating these variables improves the reproducibility and may accelerate the degradation process. Laboratory simulations can also be used for the assessment of long-term effects due to continuous dosing on the activity and the environment of the disposal system [50].

## 11.5.6.2 Applications

The OECD Coupled Unit test [106] simulates an activated sludge sewage treatment system, but its application for polymers would be difficult as DOC is the parameter used to assess biodegradability. Krupp and Jewell [107] described well-controlled anaerobic and aerobic aquatic bioreactors to study degradation of a range of commercially available polymer films. A relatively low loading rate of the semicontinuous reactors and a long retention time were maintained to maximize the efficiency of biodegradation. Experimental setups have also been designed to simulate marine environments [108], soil burial conditions [108–110], composting environments [111–114], and landfill conditions [115] at laboratory scale, with controlled parameters such as temperature and moisture level, and a synthetic waste, to provide a standardized basis for comparing the degradation kinetics of films.

A wide choice of material properties can be followed during the degradation process. However, it is important to select one which is relevant to the end-use of the polymer material or provides fundamental information about the degradation process. Weight loss is a parameter frequently followed because it clearly demonstrates the disintegration of a biodegradable product [116–118]. Tensile properties are also often monitored, due to the interest in the use of biodegradable plastics in packaging applications [54, 119, 120]. In those polymers where the biodegradation involves a random scission of the macromolecular chains, a decrease in the average molecular weight and a general broadening of the molecular weight distribution provide initial evidence of a breakdown process [86, 121, 122]. However, no significant changes in material characteristics may be observed in recovered

material if the mechanism of biodegradation involves bioerosion, that is, enzymatic or hydrolytic cleavage at the surface. Visual examination of the surface with various microscopic techniques can also give information on the biodegradation process [123–126]. Likewise, chemical and/or physical changes in the polymer may be followed by (combinations of) specific techniques such as infrared [10, 127] or UV spectroscopy [84, 128], nuclear magnetic resonance measurements [122–129], X-ray diffractometry [130, 131], and differential scanning calorimetry [132, 133].

#### 11.5.6.3 **Drawbacks**

An inherent drawback in the use of mechanical properties, weight loss, molecular weights, or any other property which relies on the macromolecular nature of the substrate is that in spite of their sensitivity, these can only address the early stages of the biodegradation process. Furthermore, these parameters can give no information on the extent of mineralization. Especially in material blends or copolymers, the hydrolysis of one component can cause significant disintegration (and thus loss of weight and tensile properties), whereas other components may persist in the environment, even in disintegrated form [13]. Blends of starch, poly(3hydroxybutyrate) or poly(\varepsilon-caprolactone) with polyolefins are examples of such systems [11, 43, 134].

#### 11.5.7

## Natural Environments, Field Trials

Exposures in natural environments provide the best true measure of the environmental fate of a polymer, because these tests include a diversity of organisms and achieve a desirable natural closeness of fit between the substrate, microbial agent, and the environment. However, the results of that exposure are only relevant to the specific environment studied, which is likely to differ substantially from many other environments. An additional problem is the timescale for this method, since the degradation process, depending on the environment, may be very slow (months to years) [23]. Moreover, little information on the degradation process can be gained other than the real time required for weight loss or total disintegration.

Nevertheless, field trials in natural environments are still used to extrapolate results acquired in laboratory tests to biodegradation behavior under realistic outdoor conditions [123, 135, 136].

# 11.6 Conclusions

The overview presented above makes clear that there is no such thing as a single optimal method for determining biodegradation of polymeric materials. First of all, biodegradation of a material is not only determined by the chemical composition and corresponding physical properties; the degradation environment in which the material is exposed also affects the rate and degree of biodegradation.

Furthermore, the method or test to be used depends on what information is requested.

One should realize that biodegradability is usually not of interest by itself. It is often just one aspect of health and environmental safety issues or integrated waste management concepts. It is fairly obvious but often neglected that one should always consider why a particular polymeric material should be (or not be) biodegradable when contemplating how to assess its biodegradability. After all, it is the intended application of the material that governs the most suitable testing environment, the parameters to be measured during exposure, and the corresponding limit values. For example, investigating whether biodegradation of a plastic material designed for food packaging could facilitate undesired growth of (pathogenic) microorganisms requires a completely different approach from investigating whether its waste can be discarded via composting (i.e., whether it degrades sufficiently rapid to be compatible with existing biowaste composting facilities).

In most cases, it will not be sufficient to ascertain macroscopic changes, such as weight loss and disintegration, or growth of microorganisms, because these observations may originate from biodegradation of just one of separate components. The ultimate fate of all individual components and degradation products must be included in the investigations. This implies that it is essential that both the polymeric materials and also intermediate degradation products have to be well characterized in order to understand the degradation process. For a good number of biodegradable materials, this means that a lot of work still needs to be done.

#### References

- 1 Hayashi, T. (1994) Prog. Polym. Sci., 19, 663.
- 2 Williams, D.F. and Zhong, S.P. (1994) Int. Biodeterior. Biodegradation, 34, 95.
- 3 Buchanan, F. (ed.) (2008) Degradation Rate of Bioresorbable Materials—Prediction and Evaluation, Woodhead Publishing Limited, Cambridge, p. 397.
- **4** Göpferich, A. (1996) *Biomaterials*, **17**, 103.
- 5 Mabilleau, G. and Albertsson (2008) Sabokbar, in *Degradation Rate of Bioresorbable Materials—Prediction and Evaluation* (ed. F. Buchanan), Woodhead Publishing Limited, Cambridge, p. 145.
- 6 Albertsson, A.-C. and Karlsson, S. (1990) Degradable Materials – Perspectives, Issues and Opportunities, (eds S.A. Barenberg, J.L. Brash, R. Narayan, and A.E. Redpath), CRC Press, Boston, p. 263.

- **7** Palmisano, A.C. and Pettigrew, C.A. (1992) *Bioscience*, **42**, 680.
- 8 Albertsson, A.-C. and Rånby, B. (1979) J. Appl. Polym. Sci. Appl. Polym. Symp., 35, 423.
- 9 Austin, R.G. (1990) Degradable Materials – Perspectives, Issues and Opportunities (eds S.A. Barenberg, J.L. Brash, R. Narayan, and A.E. Redpath), CRC Press, Boston, p. 209.
- 10 Goheen, S.M. and Wool, R.P. (1991) J. Appl. Polym. Sci., 42, 2691.
- 11 Breslin, V.T. (1993) J. Environ. Polym. Degrad., 1, 127.
- 12 Anonymous (1992) Towards Common Ground—Meeting Summary of the International Workshop on Biodegradability, Annapolis, MD, USA, 20–21 October, 1992.
- 13 Andrady, A.L. (1994) *J.M.S.-Rev. Macromol. Chem. Phys.*, C34, 25.

- 14 Andrady, A.L. (2000) Handbook of Polymer Degradation, 2nd edn (ed. S.H. Hamid), Marcel Dekker, New York, p. 441
- 15 St.Pierre, T. and Chiellini, E. (1986) *Bioact. Compat. Polym.*, 1, 467.
- 16 St.Pierre, T. and Chiellini, E. (1987) *Bioact. Compat. Polym.*, 2, 4.
- 17 Cameron, R.E. and Kamvari-Moghaddam, A. (2008) Degradation Rate of Bioresorbable Materials – Prediction and Evaluation (ed. F. Buchanan), Woodhead Publishing Limited, Cambridge, p. 43.
- 18 Stryer, L. (1981) Biochemistry, 2nd edn. W.H. Freeman and Company, New York, USA.
- 19 Anderson, T.A., Tsao, R., and Coats, J.R. (1993) J. Environ. Polym. Degrad., 1, 301.
- 20 Whitney, P.J., Swaffield, C.H., and Graffam, A.J. (1993) *Int. Biodeter. Biodegrad.*, 31, 179.
- 21 Van der Zee, M., Stoutjesdijk, J.H., Van der Heijden, P.A.A.W., and De Wit, D. (1995) J. Environ. Polym. Degrad., 3, 235.
- 22 Eggink, G., Van der Zee, M., and Sijtsma, L. (1995) International edition of the IOP on Environmental Biotechnology, 7–8.
- 23 Mayer, J.M. and Kaplan, D.L. (1993) Biodegradable Polymers and Packaging (eds C. Ching, D.L. Kaplan , and E.L. Thomas), Technomic Publishing, Lancaster-Basel, p. 233.
- 24 Mochizuki, M., Hirano, M., Kanmuri, Y., Kudo, K., and Tokiwa, Y. (1995) *J. Appl. Polym. Sci.*, 55, 289.
- 25 Tokiwa, Y. and Suzuki, T. (1981) J. Appl. Polym. Sci., 26, 441.
- **26** Tokiwa, Y., Suzuki, T., and Takeda, K. (1986) *Agric. Biol. Chem.*, **50**, 1323.
- 27 Arvanitoyannis, I., Nakayama, A., Kawasaki, N., and Yamamoto, N. (1995) Polymer, 36, 2271.
- 28 Nakayama, A., Kawasaki, N., Arvanitoyannis, I., Iyoda, J., and Yamamoto, N. (1995) *Polymer*, 36, 1295.
- 29 Walter, T., Augusta, J., Müller, R.-J., Widdecke, H., and Klein, J. (1995) Enzym. Microb. Technol., 17, 218.
- 30 Nagata, M., Kiyotsukuri, T., Ibuki, H., Tsutsumi, N., and Sakai, W. (1996) React. Funct. Polym., 30, 165.

- 31 Jun, H.S., Kim, B.O., Kim, Y.C., Chang, H.N., and Woo, S.I. (1994) J. Environ. Polym. Degrad., 2, 9.
- 32 Chiellini, E., Corti, A., Giovannini, A., Narducci, P., Paparella, A.M., and Solaro, R. (1996) J. Environ. Polym. Degrad., 4, 37.
- 33 Nagata, M., Kiyotsukuri, T., Minami, S., Tsutsumi, N., and Sakai, W. (1996) Polym. Int., 39, 83.
- **34** Nagata, M. and Kiyotsukuri, T. (1994) *Eur. Polym. J.*, **30**, 1277.
- **35** Nagata, M. (1996) *Macromol. Rap. Commun.*, **17**, 583.
- 36 Arvanitoyannis, I., Nikolaou, E., and Yamamoto, N. (1994) *Polymer*, 35, 4678.
- 37 Arvanitoyannis, I., Nikolaou, E., and Yamamoto, N. (1995) Macromol. Chem. Phys., 196, 1129.
- 38 Matsumura, S., Shimura, Y., Toshima, K., Tsuji, M., and Hatanaka, T. (1995) Macromol. Chem. Phys., 196, 3437.
- 39 Glasser, W.G., McCartney, B.K., and Samaranayake, G. (1994) *Biotechnol. Prog.*, 10, 214.
- 40 Rivard, C., Moens, L., Roberts, K., Brigham, J., and Kelley, S. (1995) Enzym. Microb. Technol., 17, 848.
- **41** Strantz, A.A. and Zottola, E.A. (1992) *J. Food Protect.*, **55**, 736.
- 42 Coma, V., Couturier, Y., Pascat, B., Bureau, G., Cuq, J.L., and Guilbert, S. (1995) Enzyme Microb. Technol., 17, 524.
- 43 Imam, S.H., Gordon, S.H., Burgess-Cassler, A., and Greene, R.V. (1995) *J. Environ. Polym. Degrad.*, 3, 107.
- 44 Imam, S.H., Gordon, S.H., Shogren, R.L., and Greene, R.V. (1995) J. Environ. Polym. Degrad., 3, 205.
- 45 Vikman, M., Itävaara, M., and Poutanen, K. (1995) J.M.S.-Pure Appl. Chem., A32, 863
- 46 Vikman, M., Itävaara, M., and Poutanen, K. (1995) J. Environ. Polym. Degrad., 3,
- 47 ASTM (2009) G21-96. Standard Practice for Determining Resistance of Synthetic Polymeric Materials to Fungi, American Society for Testing and Materials (ASTM), Philadelphia, PA, USA.
- **48** ASTM (1996) G22-76. Standard Practice for Determining Resistance of Plastics to Bacteria, American Society for

- Testing and Materials (ASTM), Philadelphia, PA, USA (withdrawn in 2002).
- 49 International Standard (1997) ISO 846. Plastics – Evaluation of the action of micro-organisms, International Organization for Standardization (ISO), Genève, Switzerland.
- 50 Seal, K.J. (1994) Chemistry and Technology of Biodegradable Polymers (ed. G.J.L. Griffin), Blackie Academic and Professional, London, p. 116.
- 51 Potts, J.E. (1978) Aspects of Degradation and Stabilization of Polymers (ed. H.H.G. Jellinek), Elsevier Scientific Publishing Co., Amsterdam, p. 617.
- **52** Seal, K.J. and Pantke, M. (1986) *Mater. Org.*, **21**, 151.
- 53 Delafield, F.P., Doudoroff, M., Palleroni, N.J., Lusty, C.J., and Contopoulos, R. (1965) J. Bacteriol., 90, 1455.
- 54 Gould, J.M., Gordon, S.H., Dexter, L.B., and Swanson, C.L. (1990) Agricultural and Synthetic Polymers—Biodegradability and Utilization (eds J.E. Glass and G. Swift), American Chemical Society, Washington, DC, ACS Symposium Series 433. p. 65.
- 55 Augusta, J., Müller, R.-J., and Widdecke, H. (1993) Appl. Microbiol. Biotechnol., 39, 673
- **56** Nishida, H. and Tokiwa, Y. (1994) *Chem. Lett.*, **3**, 421.
- **57** Nishida, H. and Tokiwa, Y. (1994) *Chem. Lett.*, **7**, 1293.
- 58 Crabbe, J.R., Campbell, J.R., Thompson, L., Walz, S.L., and Schultz, W.W. (1994) Int. Biodeter. Biodegrad., 33, 103.
- 59 International Standard (1999) ISO 9408:1999(E). Water quality—Evaluation of ultimate aerobic biodegradability of organic compounds in aqueous medium by determination of oxygen demand in a closed respirometer, International Organization for Standardization (ISO), Genève, Switzerland.
- 60 International Standard (1997) ISO 10708:1997(E). Water quality Evaluation in an aqueous medium of the ultimate aerobic biodegradability of organic compounds Determination of biochemical oxygen demand in a two-phase closed bottle test, International Organization for

- Standardization (ISO), Genève, Switzerland.
- 61 OECD (1993) 301D. Ready
  Biodegradability: Closed Bottle Test,
  Guidelines for Testing of Chemicals,
  Organization for Economic Cooperation
  and Development (OECD), Paris, France.
- 62 OECD (1993) 302C. Inherent Biodegradability: Modified MITI Test (II), Guidelines for Testing of Chemicals, Organization for Economic Cooperation and Development (OECD), Paris, France.
- 63 International Standard (1989) ISO 6060:1989(E). Water quality – Determination of the chemical oxygen demand, International Organization for Standardization (ISO), Genève, Switzerland.
- 64 International Standard (1997) ISO 10707:1997(E). Water quality—Evaluation in an aqueous medium of the "ultimate" aerobic biodegradability of organic compounds—Method by analysis of biochemical oxygen demand (closed bottle test), International Organization for Standardization (ISO), Genève, Switzerland.
- 65 International Standard (2004) ISO 14851:2004(E). Determination of the ultimate aerobic biodegradability of plastic materials in an aqueous medium—Method by measuring the oxygen demand in a closed respirometer, International Organization for Standardization (ISO), Genève, Switzerland.
- 66 ASTM (2002) ASTM D5271-02 Standard
  Test Method for Determining the
  Aerobic Biodegradation of Plastic
  Materials in an Activated-SludgeWastewater-Treatment System.
  American Society for Testing and
  Materials (ASTM), Philadelphia, PA,
  USA (withdrawn in 2011).
- 67 European Standard (2003) EN
  14048:2003. Packaging—Determination
  of the ultimate aerobic biodegradability
  of packaging materials in an aqueous
  medium—Method by measuring the
  oxygen demand in a closed
  respirometer, European Committee for
  Standardization (CEN), Brussels,
  Belgium.

- 68 OECD (1993) 301F. Manometric Respirometry Test, Guidelines for Testing of Chemicals, Organization for Economic Cooperation and Development (OECD), Paris, France.
- 69 Tilstra, L. and Johnsonbaugh, D. (1993) J. Environ. Polym. Degrad., 1, 247.
- 70 ASTM (1992) D5209-92. Standard Test Method for Determining the Aerobic Biodegradation of Plastic Materials in the Presence of Municipal Sewage Sludge, American Society for Testing and Materials (ASTM), Philadelphia, PA, USA (withdrawn in 2004).
- 71 International Standard (2000) ISO 9439:2000(E). Water quality—Evaluation of ultimate aerobic biodegradability of organic compounds in aqueous medium—Carbon dioxide evolution test, International Organization for Standardization (ISO), Genève, Switzerland.
- 72 International Standard (2004) ISO 14852:2004(E). Determination of the ultimate aerobic biodegradability of plastic materials in an aqueous medium—Method by analysis of evolved carbon dioxide, International Organization for Standardization (ISO), Genève, Switzerland.
- 73 European Standard (2003) EN
  14047:2003. Packaging Determination
  of the ultimate aerobic biodegradability
  of packaging materials in an aqueous
  medium Method by analysis of evolved
  carbon dioxide, European Committee for
  Standardization (CEN), Brussels,
  Belgium.
- 74 OECD (1993) 301B. Ready
  Biodegradability: Modified Sturm Test,
  Guidelines for Testing of Chemicals,
  Organization for Economic Cooperation
  and Development (OECD), Paris,
  France.
- 75 ASTM (2009) D6691-09. Standard Test Method for Determining Aerobic Biodegradation of Plastic Materials in the Marine Environment by a Defined Microbial Consortium or Natural Sea Water Inoculum, American Society for Testing and Materials (ASTM), Philadelphia, PA, USA.
- **76** ASTM (2003) D5338-98(2003). Standard Test Method for Determining Aerobic

- Biodegradation of Plastic Materials Under Controlled Composting Conditions, American Society for Testing and Materials (ASTM), Philadelphia, PA, USA.
- 77 International Standard (2007) 14855-1:2007(E). Determination of the ultimate aerobic biodegradability of plastic materials under controlled composting conditions—Method by analysis of evolved carbon dioxide—Part 1: General method, International Organization for Standardization (ISO), Genève, Switzerland.
- 78 International Standard (2009) 14855-2:2009(E). Determination of the ultimate aerobic biodegradability of plastic materials under controlled composting conditions—Method by analysis of evolved carbon dioxide—Part 2: Gravimetric measurement of carbon dioxide evolved in a laboratory-scale test, International Organization for Standardization (ISO), Genève, Switzerland.
- 79 European Standard (2003) EN 14046:2003. Packaging – Evaluation of the ultimate aerobic biodegradability of packaging materials under controlled composting conditions – Method by analysis of released carbon dioxide, European Committee for Standardization (CEN), Brussels, Belgium.
- 80 ASTM (2007) D5210-92. Standard Test Method for Determining the Anaerobic Biodegradation of Plastic Materials in the Presence of Municipal Sewage Sludge, American Society for Testing and Materials (ASTM), Philadelphia, PA, USA.
- 81 International Standard (1998) ISO 11734:1998(E). Water quality—Evaluation of the "ultimate" anaerobic biodegradability of organic compounds in digested sludge—Method by measurement of the biogas production, International Organization for Standardization (ISO), Genève, Switzerland.
- 82 ASTM (2011) D5511-11. Standard Test Method for Determining Anaerobic Biodegradation of Plastic Materials Under High-Solids Anaerobic-Digestion

- Conditions, American Society for Testing and Materials (ASTM), Philadelphia, PA, USA.
- 83 ASTM (2011) D5526-94(2011)e1.
  Standard Test Method for Determining
  Anaerobic Biodegradation of Plastic
  Materials Under Accelerated Landfill
  Conditions, American Society for
  Testing and Materials (ASTM),
  Philadelphia, PA, USA.
- 84 Day, M., Shaw, K., and Cooney, J.D. (1994) J. Environ. Polym. Degrad., 2, 121.
- 85 Puechner, P., Mueller, W.-R., and Bardtke, D. (1995) J. Environ. Polym. Degrad., 3, 133.
- 86 Bartha, R. and Yabannavar, A. (1995)
  Proceedings of the Fourth International
  Workshop on Biodegradable Plastics and
  Polymers and Fourth Annual Meeting of
  the Bio-Environmentally Degradable
  Polymer Society, Durham, NH, USA,
  October 11–14, 1995.
- 87 Andrady, A.L., Pegram, J.E., and Tropsha, Y. (1993) J. Environ. Polym. Degrad., 1, 171.
- **88** Yabannavar, A. and Bartha, R. (1993) *Soil Biol. Biochem.*, **25**, 1469.
- **89** Yabannavar, A.V. and Bartha, R. (1994) *Appl. Environ. Microbiol.*, **60**, 3608.
- 90 Urstadt, S., Augusta, J., Müller, R.-J., and Deckwer, W.-D. (1995) *J. Environ.* Polym. Degrad., 3, 121.
- **91** Itävaara, M. and Vikman, M. (1995) *Chemosphere*, **31**, 4359.
- 92 Spitzer, B., Mende, C., Menner, M., and Luck, T. (1996) J. Environ. Polym. Degrad., 4, 157.
- 93 Allen, A.L., Mayer, J.M., Stote, R., and Kaplan, D.L. (1994) *J. Environ. Polym. Degrad.*, 2, 237.
- 94 Courtes, R., Bahlaoui, A., Rambaud, A., Deschamps, F., Sunde, E., and Dutriex, E. (1995) *Ecotoxicol. Environ. Saf.*, 31, 142.
- 95 Barak, P., Coquet, Y., Halbach, T.R., and Molina, J.A.E. (1991) J. Environ. Qual., 20, 173.
- 96 Pagga, U., Beimborn, D.B., Boelens, J., and De Wilde, B. (1995) *Chemosphere*, 31, 4475.
- 97 Pagga, U., Beimborn, D.B., and Yamamoto, M. (1996) J. Environ. Polym. Degrad., 4, 173.

- 98 Tosin, M., Degli-Innocenti, F., and Bastioli, C. (1998) J. Environ. Polym. Degr., 6, 79.
- 99 Bellia, G., Tosin, M., Floridi, G., and Degli-Innocenti, F. (1999) Polym. Degrad. Stabil., 66, 65.
- 100 ASTM (2007) D6340-98. Standard Test Methods for Determining Aerobic Biodegradation of Radiolabeled Plastic Materials in an Aqueous or Compost Environment, American Society for Testing and Materials (ASTM), Philadelphia, PA, USA.
- 101 ASTM (2001) D6692-01. Standard Test Method for Determining the Biodegradability of Radiolabeled Polymeric Plastic Materials in Seawater, American Society for Testing and Materials (ASTM), Philadelphia, PA, USA (withdrawn in 2010).
- 102 ASTM (2002) D6776-02. Standard Test Method for Determining Anaerobic Biodegradability of Radiolabeled Plastic Materials in a Laboratory-Scale Simulated Landfill Environment, American Society for Testing and Materials (ASTM), Philadelphia, PA, USA.
- 103 Albertsson, A.-C., Barenstedt, C., and Karlsson, S. (1993) J. Environ. Polym. Degrad., 1, 241.
- 104 Komarek, R.J., Gardner, R.M., Buchanan, C.M., and Gedon, S. (1993) J. Appl. Polym. Sci., 50, 1739.
- 105 Buchanan, C.M., Dorschel, D., Gardner, R.M., Komarek, R.J., Matosky, A.J., White, A.W., and Wood, M.D. (1996) J. Environ. Polym. Degrad., 4, 179.
- 106 OECD (1993) 303A. Simulation Test-Aerobic Sewage Treatment: Coupled Units Test. Guidelines for Testing of Chemicals, Organization for Economic Cooperation and Development (OECD), Paris, France.
- 107 Krupp, L.R. and Jewell, W.J. (1992) Environ. Sci. Technol., 26, 193.
- 108 Kaplan, D.L., Mayer, J.M., Greenberger, M., Gross, R., and McCarthy, S. (1994) Polym. Degrad. Stab., 45, 165.
- **109** Dale, R. and Squirrell, D.J. (1990) *Int. Biodeterior.*, **26**, 355.
- 110 Seal, K.J. and Pantke, M. (1990) *Mater. Org.*, 25, 87.

- 111 Gardner, R.M., Buchanan, C.M., Komarek, R., Dorschel, D., Boggs, C., and White, A.W. (1994) J. Appl. Polym. Sci., 52, 1477.
- 112 Buchanan, C.M., Dorschel, D.D., Gardner, R.M., Komarek, R.J., and White, A.W. (1995) J.M.S. – Pure Appl. Chem., A32, 683.
- 113 Gross, R.A., Gu, J.-D., Eberiel, D., and McCarthy, S.P. (1995) J.M.S.—Pure Appl. Chem., A32, 613.
- 114 European Standard (2003) EN 14045:2003. Packaging – Evaluation of the disintegration of packaging materials in practical oriented tests under defined composting conditions, European Committee for Standardization (CEN), Brussels, Belgium.
- 115 Smith, G.P., Press, B., Eberiel, D., McCarthy, S.P., Gross, R.A., and Kaplan, D.L. (1990) *Polym. Mater. Sci. Eng.*, **63**, 862.
- 116 Coma, V., Couturier, Y., Pascat, B., Bureau, G., Guilbert, S., and Cuq, J.L. (1994) *Pack. Techn. Sci.*, 7, 27.
- 117 Buchanan, C.M., Boggs, C.N., Dorschel, D., Gardner, R.M., Komarek, R.J., Watterson, T.L., and White, A.W. (1995) J. Environ. Polym. Degrad., 3, 1.
- **118** Goldberg, D. (1995) *J. Environ. Polym. Degrad.*, **3**, 61.
- 119 Iannotti, G., Fair, N., Tempesta, M., Neibling, H., Hsieh, F.H., and Mueller, M. (1990) Degradable Materials – Perspectives, Issues and Opportunities (eds S.A. Barenberg, J.L. Brash, R. Narayan, and A.E. Redpath), CRC Press, Boston, p. 425.
- 120 Mergaert, J., Webb, A., Anderson, C., Wouters, A., and Swings, J. (1993) Appl. Environ. Microbiol., 59, 3233.

- 121 Tilstra, L. and Johnsonbaugh, D. (1993)

  J. Environ. Polym. Degrad., 1, 257.
- **122** Hu, D.S.G. and Liu, H.J. (1994) *J. Appl. Polym. Sci.*, **51**, 473.
- 123 Greizerstein, H.B., Syracuse, J.A., and Kostyniak, P.J. (1993) *Polym. Degrad.* Stab., 39, 251.
- **124** Lopez-Llorca, L.V. and Colom Valiente, M.F. (1993) *Micron*, **24**, 457.
- 125 Nishida, H. and Tokiwa, Y. (1993) J. Environ. Polym. Degrad., 1, 227.
- 126 Bastioli, C., Cerutti, A., Guanella, I., Romano, G.C., and Tosin, M. (1995) I. Environ. Polym. Degrad., 3, 81.
- 127 Kay, M.J., McCabe, R.W., and Morton, L.H.G. (1993) *Int. Biodeter. Biodegrad.*, 31, 209.
- 128 Allen, N.S., Edge, M., Mohammadian, M., and Jones, K. (1994) *Polym. Degrad. Stab.*, 43, 229.
- 129 Löfgren, A. and Albertsson, A.-C. (1994) *J. Appl. Polym. Sci.*, **52**, 1327.
- 130 Albertsson, A.-C. and Karlsson, S. (1995) Macromol. Symp., 98, 797.
- 131 Schurz, J., Zipper, P., and Lenz, J. (1993) *J.M.S. – Pure Appl. Chem.*, **A30**, 603.
- 132 Albertsson, A.-C., Barenstedt, C., and Karlsson, S. (1994) J. Appl. Polym. Sci., 51, 1097.
- 133 Santerre, J.P., Labow, R.S., Duguay, D.G., Erfle, D., and Adams, G.A. (1994) *J. Biomed. Mat. Res.*, 28, 1187.
- 134 Iwamoto, A. and Tokiwa, Y. (1994) *Polym. Degrad. Stab.*, 45, 205.
- 135 Leonas, K.K., Cole, M.A., and Xiao, X.-Y. (1994) *J. Environ. Polym. Degrad.*, 2, 253.
- 136 Halley, P., Rutgers, R., Coombs, S., Kettels, J., Gralton, J., Christie, G., Jenkins, M., Beh, H., Griffin, K., Jayasekara, R., and Lonergan, G. (2001) Starch – Stärke, 53, 362.