

Climate and soil drive forest structure in Bolivian lowland forests

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Abstract: Climate is one of the most important factors determining variation in forest structure, but whether soils have independent effects is less clear. We evaluate how climate and soil independently affect forest structure, using 89 200 stems ≥ 10 cm dbh from 220 1-ha permanent plots distributed along environmental gradients in lowland Bolivia. Fifteen forest structural variables, related to vertical structure (forest height and layering), horizontal structure (basal area, median and the 99th percentile of the stem diameter and size-class distribution) and density of life forms (tree, palm and liana), were evaluated. Environmental variables were summarized in four multivariate axes, related to rainfall, temperature, soil fertility and soil texture. Multiple regression indicates that all structural variables were affected by one or more of the environmental axes, but the explained variation was generally low (median $R^2 = 0.15$). Rainfall and soil texture affected most forest structural variables (respectively 87% and 80%) and had qualitatively similar effects. This suggests that plant water availability, as determined by rainfall and soil water retention capacity, is the strongest driver of forest structure, whereas soil fertility was a weaker driver of forest structure, affecting 53% of the variables. Maximum forest height, palm density, total basal area and liana infestation showed the strongest responses to environmental variation (with R^2 ranging from 0.31–0.82). Forest height, palm density and total basal area increased with plant water availability, while liana infestation decreased with plant water availability. Therefore, multiple rather than single environmental factors must be used to explain the structure of tropical forests.

Key Words: climate, environmental gradient, forest structure, liana, lowland Bolivia, palm, soil, tree

INTRODUCTION

Tropical lowland forests are very complex in structure (Richards 1996). In general, the term ‘forest structure’ is used to describe the architecture, organization, composition or abundance of the different assemblages occurring in forests. Forest structure encompasses many components that can be described in numerous ways (Bongers 2001, Spies 1998); however, most papers describing forest structure have focused on few tree variables, mainly maximum canopy height, stem density and basal area (Ashton & Hall 1992, Clark & Clark 2000, DeWalt & Chave 2004, Killeen *et al.* 1998, Poorter *et al.* 2008). In this paper we analyse 15 forest structural variables related to tree height, crown position, diameter distribution and liana load and focus on three woody life forms (trees, palms and lianas). These life-forms are

key components of tropical forests because they largely determine the structure, biomass and diversity of these forests (Gentry 1991, Schnitzer & Bongers 2002).

Several factors have been proposed to be strong drivers of forest structure in tropical regions, climate and soil being the most important ones (Clark & Clark 2000, Malhi *et al.* 2002, Murphy & Lugo 1986). For example, canopy height tends to be higher in wetter and less seasonal forests, but tends to be independent of soil fertility (Ashton & Hall 1992, Swaine *et al.* 1990). Stem density and basal area of trees tend to be higher with a shorter dry season (Losos *et al.* 2004, Malhi *et al.* 2002). A comparison of four Neotropical lowland forests showed that stem density and basal area tended to be higher on less fertile soils while richer soils were characterized by a low tree but high palm density (DeWalt & Chave 2004). Relationships between soil fertility and density of lianas are equivocal, as a positive relationship was found in Amazonian and Malaysian forests (DeWalt *et al.* 2006, Laurance *et al.* 2001, Putz & Chai 1987) but no relationship in other

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tropical forests (DeWalt & Chave 2004, Ibarra-Manríquez & Martínez-Ramos 2002).

Large-scale patterns in forest structure and their underlying driving factors are less well-known (but see DeWalt & Chave 2004, Lewis *et al.* 2004, Proctor *et al.* 1983), because studies describing forest structure included only one site (Bongers *et al.* 1988, Milliken 1998, Newbery *et al.* 1992, Poulsen *et al.* 2006) or studies quantifying the drivers included mostly one driver only, e.g. climate (Takyu *et al.* 2005, Vieira *et al.* 2004) or soil (Faber-Langendoen & Gentry 1991, Nebel *et al.* 2001, Paoli *et al.* 2008). Understanding the patterns and causes of spatial variation in forest structure is important to understand the history and function of forest ecosystems.

Bolivia provides an ideal setting to study vegetation–environment relationships because it covers an extraordinary display of vegetation types and soil heterogeneity across a rainfall gradient (Toledo *et al.* 2011a). We use data from 89 200 stems ≥ 10 cm dbh from 220 1-ha plots, distributed over an area of c. 160 000 km², to analyse the independent effects of climate and soil on the structure of tropical lowland forests. We predict that stem density and basal area of trees and palms will increase with water and nutrient availability because higher resource availability permits more stems to coexist; and we predict that liana density will increase in drier forests due to its dry-season advantage (Schnitzer 2005). To our knowledge, this is the largest study in the Neotropics doing broad-scale comparisons of forest structure considering both environmental factors.

METHODS

Study area

The 220 1-ha plots were established in old-growth forests in lowland Bolivia (10°–18°S, 59°–69°W) between 1995 and 2007 by various projects and forestry concessionaries (see Acknowledgements for more details). For a map of Bolivia and the spatial location of the plots, see Toledo *et al.* (2011a). Currently, the network of plots is being coordinated and managed by the Instituto Boliviano de Investigación Forestal (IBIF). Nearly all plots are in upland forests (terra firme; only 5% of the plots were found in areas with seasonal flooding), generally on flat terrain (20% in slopes of hilly areas), and in an altitude range of 100–500 m asl. The selected plots are distributed along a climatic gradient in lowland Bolivia where the precipitation can vary from 600 to 3000 mm y⁻¹ from the driest to wettest areas (based on at least 30 y data, Servicio Nacional de Meteorología e Hidrología – SENAMHI, unpubl. data). In general, this area experiences a 4–7-mo-long dry period (e.g. precipitation < 100 mm mo⁻¹), mostly from April to October, corresponding to

the austral winter. Mean annual temperature is between 24 °C and 26 °C. Soils in lowland Bolivia are variable (Gerold 2003), due to differences in geomorphology and geological history. The north contains the wide fluvial plain of the Amazon basin, in the west the relatively young landscapes of the Andean foothills occur, while in the east ancient rocks of the Pre-Cambrian Shield abound (Suárez-Soruco 2000). For the average and range of environmental conditions, see Toledo *et al.* (2011b).

Environmental data

For each plot we obtained five climatic variables, interpolated from available data from 45 weather stations in the region, and 14 edaphic variables obtained from a composite soil sample from the first 30 cm of soil depth, from 20 locations in each plot. More details of soil analysis and climate interpolation can be found in Toledo (2010). To summarize these often highly correlated environmental data we performed two Principal Component Analyses (PCAs). One PCA was done using five climatic variables; annual temperature, annual precipitation, the sum of precipitation from three driest months and the length of the dry period (< 100 mm of precipitation mo⁻¹) and the drought period (< 50 mm mo⁻¹). A second PCA was done using 12 edaphic variables (CEC, Ca, K, Mg, Na, P, OM, N, acidity, sand, silt and clay). The first two axes resulting from these PCAs were used as the four main environmental axes in the analysis below. The first climatic axis (hereafter rainfall axis) explained 65% of the variation and correlated positively with the annual precipitation and negatively with the dry period. The second climatic axis (hereafter temperature axis) explained 29% and correlated positively with the mean annual temperature and negatively with the precipitation in the driest months. The first edaphic axis (hereafter soil fertility axis) explained 48% and correlated positively with soil fertility variables (CEC, Ca, Mg, Na, K, P, OM and N), and negatively with acidity. The second edaphic axis (hereafter soil texture axis) explained 20% and correlated positively with clay and silt, and negatively with sand content. The rainfall axis was weakly and negatively related to the soil fertility axis, but strongly and positively to the soil texture axis (Appendix 1). These results indicate that in high rainfall areas the soil were less fertile and had a higher silt and clay content. Although in general drier forests tended to have higher soil fertility than moister forests, some plots in moist areas had also soils with high fertility. These latter plots were all situated in the foothills of the Andes, which are from a younger geological origin, and hence more fertile (Toledo *et al.* 2011a).

Forest data

Most plots were square (100 × 100 m) but 11 were rectangular (20 × 500 m). Each tree and palm ≥ 10 cm diameter at breast height (dbh; measured at 130 cm or higher when buttresses were present) was measured for its diameter, tagged and identified following standard protocols (Alder & Synnott 1992, Contreras *et al.* 1999). The crown position of each individual was scored into one of five categories: (1) no direct light, (2) some side light, (3) some overhead light, (4) full overhead light and (5) emergent crown (Dawkins & Field 1978). Each individual stem was scored into one of four categories of liana infestation: (0) without lianas, (1) lianas only on the trunk, (2) liana partially in trunk and crown, and (3) completely covered with lianas (Contreras *et al.* 1999). In addition to these variables, the height of the tallest individual in the plot (hereafter height_{max}) was measured with a clinometer for 90% of the plots.

Based on the field data we calculated 15 variables to describe forest structure. The vertical structure was described using four variables: the height_{max} and the percentage of individuals in three forest layers (emergent, canopy and subcanopy). The emergent layer consists of all individuals with emergent crown, the canopy layer consists of all individuals with full and some overhead light, and the subcanopy layer consists of all individuals with either some or no side light. The horizontal structure was described using six variables: total basal area, tree basal area, palm basal area, median stem diameter (dbh₅₀), the 99 percentile of the stem diameter (dbh₉₉) and the slope of the size-class frequency distribution at the stand level (hereafter size-class distribution – SCD). This slope was calculated by regressing the (log-transformed) number of individuals per diameter class of 10 cm width to the average diameter of each size class. Finally, the density of life forms was described using five variables: the abundance of all individuals per plot (total density), the tree density, the palm density, the density of stems with liana infestation (hereafter ‘liana-density’) and the mean liana infestation (calculated by averaging the degree of liana infestation of all trees in the plot). Although we did not measure the abundance of lianas directly, we inferred it from the latter two variables.

Data analysis

To investigate the effects of the environmental factors on forest structural variables we used a backward multiple regression, with the four main environmental axes as independent variables and each structural parameter as a dependent variable. Pearson correlations were used to evaluate how structural variables were associated amongst themselves (Appendix 2), and with

Table 1. Mean (±SD) and ranges (minimum–maximum) of 15 structural variables (related to vertical structure, horizontal structure and density of life forms) from 220 1-ha permanent plots located in the lowlands of Bolivia. The ratio was calculated by dividing the maximum value by the minimum value, except for variables with zero values. Liana density refers to the number of trees infested with lianas, and is an indicator of liana abundance of the stand. Liana infestation indicates the average liana load of the trees.

Variables	Mean ± SD	Range	Ratio
Vertical structure			
Height _{max} (m)	30.3 ± 7.4	20.0–54.0	2.7
Emergent layer (%)	12.2 ± 9.2	0–62.2	–
Canopy layer (%)	50.3 ± 13.9	20.4–97.9	4.8
Subcanopy layer (%)	36.8 ± 15.8	0.13–70.4	541
Horizontal structure			
Total basal area (m ² ha ⁻¹)	20.7 ± 4.6	9.6–32.9	3.4
Tree basal area (m ² ha ⁻¹)	19.9 ± 4.2	9.5–30.5	3.2
Palm basal area (m ² ha ⁻¹)	0.79 ± 1.2	0–8.5	–
Size class distribution (slope)	–0.03 ± 0.0	–0.04 to –0.02	0.4
dbh ₅₀ (cm)	17.4 ± 2.2	14.2–30.2	2.1
dbh ₉₉ (cm)	78.1 ± 16.6	46.9–146.6	3.1
Density of life forms			
Total density (stems ha ⁻¹)	406 ± 83.3	124–763	6.2
Tree density (stems ha ⁻¹)	378 ± 70.4	124–536	4.3
Palm density (stems ha ⁻¹)	27.9 ± 43.5	0–353	–
Liana density (stems ha ⁻¹)	199 ± 77.7	36–371	10.3
Liana infestation (mean)	0.96 ± 0.4	0.12–2.1	17.2

environmental factors (Appendix 3). If necessary the data were logarithmic (log₁₀)-, square root-, or arcsine-transformed to obtain normality. All statistical analyses and PCAs were performed with SPSS 15.0 for Windows (SPSS Inc.).

RESULTS

Variation in forest structure

Forest structure varied considerably across the plots, with the largest variation in the subcanopy and emergent layers and in palm density (Table 1). In terms of vertical structure, the height_{max} was on average 30 m (range = 20–54 m); on average 50% of the stems were in the canopy layer and only 12% in the emergent layer (range = 0%–62%). In terms of horizontal structure, the total basal area averaged 21 m² ha⁻¹ (range = 10–33 m² ha⁻¹), the dbh₅₀ was 17 cm, and the dbh₉₉ was 78 cm (range = 47–147 cm). Average total density was 406 stems ha⁻¹ (range = 124–763 stems ha⁻¹). Palms presented the largest variation in density among life forms (range = 0–353 stems ha⁻¹) followed by lianas.

Forest–environment relationships

The backward multiple regression analysis showed that the rainfall axis was significantly related to 13 of the 15

Table 2. Backward multiple regression of 15 structural variables on four environmental factors of 220 1-ha permanent plots located in lowland Bolivia. The standardized regression coefficient, F-value and coefficient of determination (R^2) are provided. Significance levels are shown. * $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$.

Structural components and variables	Environmental factors (PCA Axes)					R^2
	Rainfall	Temperature	Fertility	Texture	F	
Vertical structure						
Height _{max} (m)	0.64***	0.19***	—	0.11*	86.4	0.57
Emergent layer (%)	-0.24**	-0.21*	-0.25**	0.30***	5.1	0.09
Canopy layer (%)	0.32***	0.16*	—	—	16.2	0.13
Subcanopy layer (%)	-0.21**	-0.13*	—	-0.17*	10.3	0.13
Horizontal structure						
Total basal area (m ² ha ⁻¹)	0.16*	—	—	0.28***	19.2	0.15
Tree basal area (m ² ha ⁻¹)	—	—	—	0.26***	16.4	0.07
Palm basal area (m ² ha ⁻¹)	0.70***	—	-0.14***	0.16***	158	0.69
Size class distribution (slope)	0.58***	0.39***	0.38***	-0.16*	22.2	0.29
Dbh ₅₀ (cm)	-0.39***	—	—	—	18.7	0.15
Dbh ₉₉ (cm)	0.42***	0.35***	0.32***	—	19.7	0.21
Density of life forms						
Total density (stems ha ⁻¹)	0.16*	-0.36***	-0.33***	0.27***	16.7	0.24
Tree density (stems ha ⁻¹)	-0.21*	-0.24**	-0.24**	0.21**	3.5	0.06
Palm density (stems ha ⁻¹)	0.80***	—	-0.13***	0.14***	331	0.82
Liana density (stems ha ⁻¹)	-0.19*	—	—	-0.26***	20.0	0.15
Mean liana infestation	—	0.21**	0.31***	-0.52***	32.8	0.31

forest structural variables studied, the soil texture axis to 12 of the variables and the temperature axis to nine variables (Table 2). To our surprise, the soil fertility axis was related to only eight forest structural variables. The variation explained by the backward regression models ranged from 6–82%. Palm density, palm basal area, height_{max} and mean liana infestation were the forest structural variables best explained by the models (31–

82%). Most of the forest structural variables were affected by a combination of both climatic and soil factors, tree basal area was only affected by soils and dbh₅₀ was only affected by climate. Overall, climatic and edaphic factors had more positive effects than negative effects on the forest structural variables (Table 2).

In Figures 1 and 2 we present bivariate relationships of selected forest structural variables with the rainfall

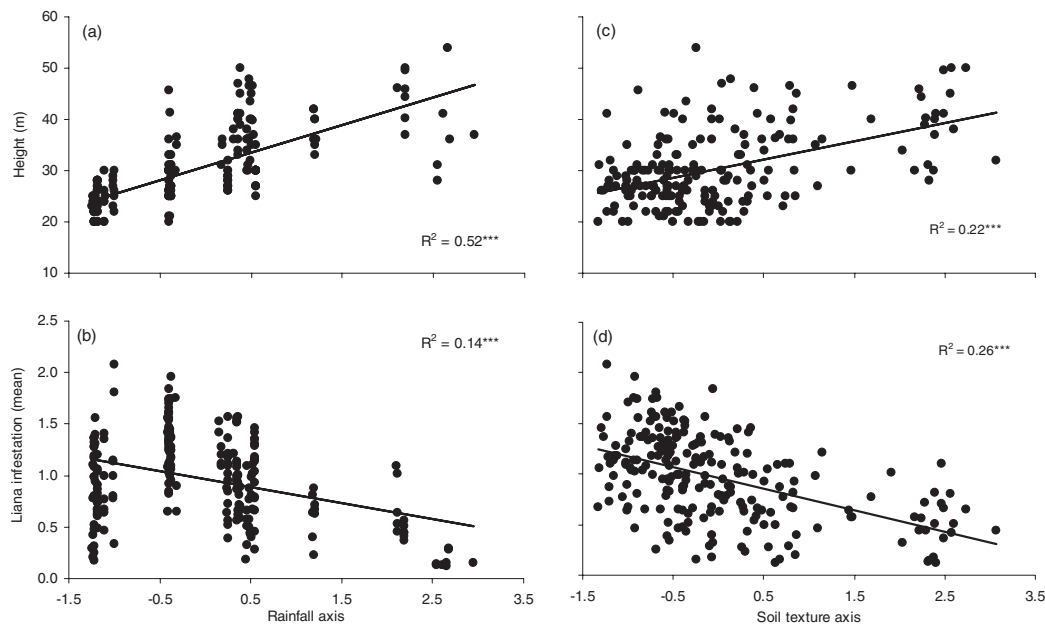


Figure 1. Relationships of the PCA rainfall axis with height_{max} (a) and liana infestation (b), and the PCA soil texture axis with height_{max} (c) and liana infestation (d) of 220 1-ha plots located in lowland Bolivia. Regression lines, corresponding coefficient of determination (R^2), and significance levels are shown. *** $P < 0.001$.

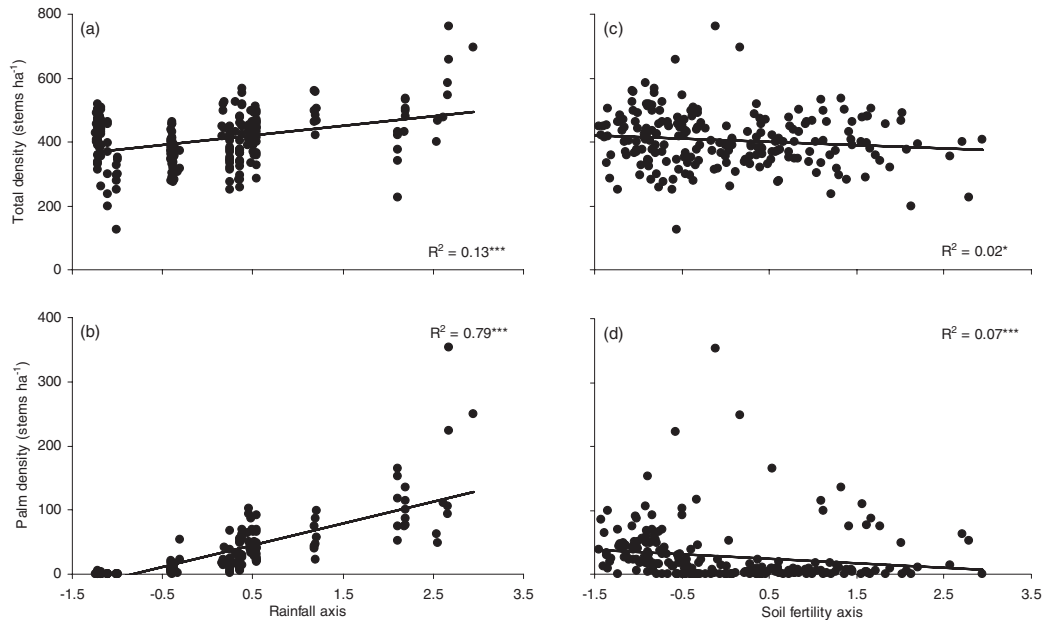


Figure 2. Relationships of the PCA rainfall axis with total density (a) and palm density (b), and the PCA soil fertility axes with total density (c) and palm density (d) of 220 1-ha plots located in lowland Bolivia. Regression lines, corresponding coefficient of determination (R^2) and significance levels are shown. * $P < 0.05$, *** $P < 0.001$.

and soil axes. Height_{\max} increased with rainfall and the silt content of the soils (Figure 1a, c). Liana infestation tended to decrease along the rainfall gradient and in plots with clay-silt soils (Figure 1b, d). Palm density increased significantly with rainfall, similar to total density (Figure 2a, b). While total and tree density were variable along the soil fertility gradient, the palm density was highest at intermediate level of soil fertility (Figure 2c, d).

Annual temperature (from 24–26 °C) and altitude (100–480 m asl) were negatively correlated (Pearson $r = -0.91$, $P < 0.001$) and the variation of these variables among plots was small.

DISCUSSION

In this study we described how forest structure differed among Bolivian forests and we analysed how climatic and edaphic factors affect forest structure. Although all forest structural variables were significantly related to at least one environmental axis, the explained variation was generally low. Height_{\max} , total basal area, palm density and liana infestation responded more to this environmental variation, and our discussion will focus mostly on these forest structural variables.

Patterns in forest structure

In general, moister forests had taller stems, higher total basal area and palm density and lower liana abundance than drier forests. The maximum tree height in Bolivian

lowland forests was 54 m; average maximum height in rain forests usually ranges between 45 and 55 m, although in some tropical rain forests individuals can reach over 60 m (Ashton & Hall 1992, Richards 1996). Dry forests are smaller in stature and tend to have an even canopy (Murphy & Lugo 1986, Richards 1996, Swaine *et al.* 1990).

In our study the total basal area for individuals ≥ 10 cm dbh averaged 21 $\text{m}^2 \text{ha}^{-1}$ and ranged from 10 to 33 $\text{m}^2 \text{ha}^{-1}$ which is at the lower end of the range (20–70 $\text{m}^2 \text{ha}^{-1}$) found for tropical forests worldwide (de Gouvenain & Silander 2003, Losos *et al.* 2004). The variation of total basal area in tropical forests can be due to variation in stem density combined with variation in tree thickness. A high basal area can be the result of many slender stems or few thick stems (Bongers *et al.* 1988). In contrast, a low basal area could also result from disturbance by logging, wind and fire, directly affecting forest structure or indirectly through changing the floristic composition and consequently the forest structure (Spies 1998). Occurrence of cyclones can also temporarily increase tree density as shown in Africa and Madagascar (de Gouvenain & Silander 2003). In lowland Bolivia, we found several plots with lower tree density and basal area due to the massive abundance of some understorey herbs or shrubs (e.g. *Phenakospermum guianense*, *Erythrochiton fallax*, *Metrodorea flavida* and *Pausandra trianae*).

Palms are a striking feature of tropical forests, being very abundant and often even dominant, forming ‘oligarchic’ forests (Vormisto *et al.* 2004). Some forests in lowland Bolivia had a relatively high palm density

(48–353 palms ha⁻¹) compared to other tropical forests: 103 palms ha⁻¹ in Cocha Cashu, Peru (Gentry & Terborgh 1990); 90–129 palms ha⁻¹ in Bajo Calima, Colombia (Faber-Langendoen & Gentry 1991) and 11–115 palms ha⁻¹ in La Selva, Costa Rica (Lieberman *et al.* 1996). Higher palm density in Bolivia could be due to local dominance of palm-rich habitats such as flooded forests, where we found the highest palm densities.

Around 50% of all trees measured in our plots had some degree of liana infestation. Drier forests in Bolivia, as reported for other studies (Carse *et al.* 2000, Pérez-Salicrup *et al.* 2001, Uslar *et al.* 2004), had between 50% and 80% of their trees infested by lianas while moister forest had less than 50% (Licón-Vasquez *et al.* 2007). Mascaro *et al.* (2004) hypothesized that high palm abundance could negatively affect the regeneration, and consequently, the abundance of lianas in wet forests. Although we have found that in lowland Bolivia moister forests have lower liana density and higher palm density, lianas and palms respond independently to rainfall. Other components of forest structure, such as the amount of small-diameter stems and branches are important for liana support and success (Putz 1984, Schnitzer & Bongers 2002). Liana density tended to increase with the percentage of trees in the subcanopy layer and to decrease with increasing canopy height (Appendix 2). Similar results, more lianas in short trees and low canopies, were found in South Africa (Balfour & Bond 1993) and Panama (DeWalt *et al.* 2000). This result suggests that connected crowns of trees in lower canopies facilitate liana support and success.

Environmental effects on forest structure and trees

In the tropics, water availability is one of the most important environmental drivers of forest structure, function and dynamics (Malhi *et al.* 2002, Murphy & Lugo 1986, Toledo *et al.* 2011b). Water availability is determined by the amount and seasonal distribution of rainfall, and the water-retention capacity of soils. The importance of rainfall and water-retention capacity is underscored by the fact that they were the most important environmental determinants of forest structure in our study. The rainfall axis was significant in 87% of the cases, and the soil texture axis in 80%. Both environmental factors worked often in a similar direction, as indicated by the sign of the regression coefficient (Table 2). For example, forest height increased with rainfall and the water-holding capacity of the soils (e.g. clay and silt content). This confirms the hypothesis of Ashton & Hall (1992) that canopy height is mostly related to soil water supply. On the other hand, stem density, and consequently total basal area, was not related to water availability as we had hypothesized because higher resource availability should allow more stems to coexist.

This lack of relationship was probably due to the fact that tree density and tree basal area were highly variable among both moist and dry forests.

Temperature and soil fertility factors had also the same direction for some structural variables. Both factors had positive effects on the size-class distribution, dbh₉₉ and liana infestation; and negative effects on the emergent layer, total density and tree density. The temperature axis represents both the annual temperature as well as the precipitation of the three driest months. Whereas the annual temperature showed a stronger correlation with maximum height and the canopy and subcanopy strata, the precipitation of the three driest months showed a stronger correlation with total basal area and total density. Plots located in northern and southern Bolivia with higher annual temperature tended to have a higher percentage of trees in the canopy layer and a lower percentage of trees in the subcanopy layer. From the data available, we can state that higher temperature (and hence decreasing altitude) may increase tree stature. Plots located in western Bolivia with lower seasonality tended to have a higher total basal area and total stem density, the latter clearly being the result of a higher palm abundance.

Environmental effects on palms

The results for palm density supported our hypothesis that moister forests had more palms than drier forests, with the highest palm density being found in seasonally flooded forests. Although palms are widely distributed in the tropics and grow in a wide range of habitats, from upland and cleared forests to the slopes of mountains, some species tend to have a higher density in seasonally flooded forests (Kahn & Henderson 1999, Velarde & Moraes 2008, Vormisto 2002). In lowland Bolivia, palm density peaked at an intermediate level of soil fertility. Similar results were found by Vormisto (2002) in the Peruvian Amazonia, where the lowest palm density was found on the richest soils and the highest palm density at an intermediate level of fertility. In contrast, other studies reported higher palm density on richer soils (Gentry & Terborgh 1990, Nebel *et al.* 2001, Sesnie *et al.* 2009). In conclusion, palm density was found to be strongly determined by rainfall but still no clear pattern was found in relation to edaphic factors. This lack of a clear pattern suggests either that palms show a stronger relationship with soil properties at the species level (Clark *et al.* 1995) than at the family level, or that other factor such as dispersal limitation are more important (Velarde & Moraes 2008).

Environmental effects on lianas

We hypothesized that lianas would have higher densities at low water availability due to their capacity for taking

water from deep soil layers (Schnitzer 2005). Liana density was indeed higher at low rainfall and on coarse soils with a low water-holding capacity. Although we did not measure liana abundance directly, our results are in line with those studies that show that lianas are more abundant in drier than in moister forests (DeWalt *et al.* 2010, Madeira *et al.* 2009, Parthasarathy *et al.* 2004, Pérez-Salicrup *et al.* 2001, Putz 1984, Schnitzer 2005, Swaine & Grace 2007). These results support the hypothesis that lianas have a growth advantage over trees in areas with a long dry season because of their deep and efficient root for taking water from deep soil layers and vascular systems (Schnitzer 2005). A more simple explanation for high liana abundance in drier forests is that lianas are light demanding and take advantage of the higher light availability in the more open dry-forest canopy. A weak positive relationship between liana abundance and soil fertility was found in lowland Bolivia, with the highly fertile terra preta plots in Eastern Bolivia having one of the highest levels of liana infestation of trees (60–80%). Other studies also found that liana abundance increases slightly with soil fertility (Balfour & Bond 1993, Poulsen *et al.* 2006, Proctor *et al.* 1983, Putz & Chai 1987). Recently, phosphorus concentrations were found to be high in liana litter (Cai & Bongers 2007), which suggests that lianas have the potential to enhance the availability of nutrients in areas where lianas are already abundant.

Concluding remarks

Rainfall and soil texture together determine plant water availability, and these were more important drivers of forest structure than soil fertility. Therefore, we consider that multiple, rather than single, environmental factors must be used to explain the forest structure in tropical forests. Compared with soil fertility (Clark & Clark 2000, DeWalt & Chave 2004, Nebel *et al.* 2001, Paoli *et al.* 2008, White & Hood 2004) modest attention has been given to soil texture as a driving factor (but see Jha & Singh 1990), although it is an important property that helps to determine the nutrient-supplying ability and the supply of water and air necessary for plant root activity (Brady 1990).

Our study supports earlier results indicating that forest structure is strongly influenced by climate (Clinebell *et al.* 1995, Gentry 1988, Murphy & Lugo 1986, ter Steege *et al.* 2003). Climatic conditions that determine the length of the growing season and intensity of rainfall may increase or decrease chemical and biological processes such as photosynthesis, respiration and soil nutrient availability (Saxe *et al.* 2001). In these Bolivian forests climate is the strongest driver of forest structure (this paper), composition (Toledo *et al.* 2011a) and dynamics

(Toledo *et al.* 2011b). As climate-change scenarios predict a decrease in rainfall and increase in dry-season length (IPCC 2007) we may expect potentially large changes in the structure and functioning of these and other tropical forests.

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LITERATURE CITED

- ALDER, D. & SYNNOTT, T. J. 1992. *Permanent sample plot techniques for mixed tropical forest*. Tropical Forestry Papers 25. Oxford Forestry Institute. 124 pp.
- ASHTON, P. S. & HALL, P. 1992. Comparisons of structure among mixed dipterocarp forests of north-western Borneo. *Journal of Ecology* 80:459–481.
- BALFOUR, D. A. & BOND, W. J. 1993. Factors limiting climber distribution and abundance in a southern African forest. *Journal of Ecology* 81:93–99.
- BONGERS, F. 2001. Methods to assess tropical rain forest canopy structure: an overview. *Plant Ecology* 153:263–277.
- BONGERS, F., POPMA, J., MEAVE DEL CASTILLO, J. & CARABIAS, J. 1988. Structure and floristic composition of the lowland rain forest of Los Tuxtlas, Mexico. *Vegetatio* 74:55–80.
- BRADY, N. C. 1990. *The nature and properties of soils*. (Tenth edition). MacMillan Publishing Company, New York. 621 pp.
- CAI, Z. Q. & BONGERS, F. 2007. Contrasting nitrogen and phosphorus resorption efficiencies in trees and lianas from a tropical montane rain forest in Xishuangbanna, southwest China. *Journal of Tropical Ecology* 23:115–118.

- CARSE, L. E., FREDERICKSEN, T. S. & LICONA, J. C. 2000. Liana-tree species associations in a Bolivian dry forest. *Tropical Ecology* 41:1–10.
- CLARK, D. A., CLARK, D. B., SANDOVAL, R. & CASTRO, M. V. 1995. Edaphic and human effects on landscape-scale distribution of tropical rain forest palms. *Ecology* 76:2581–2594.
- CLARK, D. B. & CLARK, D. A. 2000. Landscape-scale variation in forest structure and biomass in a tropical rain forest. *Forest Ecology and Management* 137:185–198.
- CLINEBELL, R. R. I., PHILLIPS, O. L., GENTRY, A. H., STARK, N. & ZUURING, H. 1995. Prediction of tropical tree and liana species richness from soil and climatic data. *Biodiversity and Conservation* 4:56–90.
- CONTRERAS, F., LEAÑO, C., LICONA, J. C., DAUBER, E., GUNNAR, L., HAGER, N. & CABA, C. 1999. *Guía para la instalación y evaluación de parcelas permanentes de muestreo (PPMs)*. BOLFOR-PROMABOSQUE, Santa Cruz, Bolivia. 50 pp.
- DAWKINS, H. C. & FIELD, D. R. B. 1978. *A long-term surveillance system for British woodland vegetation*. Department of Forestry, Oxford University, Oxford. 106 pp.
- DE GOUVENAIN, R. C. & SILANDER, J. A. 2003. Do tropical storm regimes influence the structure of tropical lowland rain forests? *Biotropica* 35:166–180.
- DEWALT, S. J. & CHAVE, J. 2004. Structure and biomass of four lowland Neotropical forests. *Biotropica* 36:7–19.
- DEWALT, S. J., SCHNITZER, S. A. & DENSLow, J. S. 2000. Density and diversity of lianas along a chronosequence in a central Panamanian lowland forest. *Journal of Tropical Ecology* 16:1–19.
- DEWALT, S. J., ICKES, K., NILUS, R., HARMS, K. E. & BURSLEM, D. F. R. P. 2006. Liana habitat associations and community structure in a Bornean lowland tropical forest. *Plant Ecology* 186:203–216.
- DEWALT, S. J., SCHNITZER, S. A., CHAVE, J., BONGERS, F., BURNHAM, R. J., CAI, Z., CHUYONG, G., CLARK, D. B., EWANGO, C. E. N., GERWING, J. J., GORTAIRE, E., HART, T., IBARRA-MANRIQUEZ, G., ICKES, K., KENFACK, D., MACÍA, M. J., MAKANA, J., MARTÍNEZ-RAMOS, M., MASCARO, J., MOSES, S., MULLER-LANDAU, H. C., PARREN, M. P. E., PARTHASARATHY, N., PÉREZ-SALICRUP, D. R., PUTZ, F. E., ROMERO-SALTOS, H. & THOMAS, D. 2010. Annual rainfall and seasonality predict pan-tropical patterns of liana density and basal area. *Biotropica* 42:309–317.
- FABER-LANGENDOEN, D. & GENTRY, A. H. 1991. The structure and diversity of rain forests at Bajo Calima, Choco Region, Western Colombia. *Biotropica* 23:2–11.
- GENTRY, A. H. 1988. Changes in plant community diversity and floristic composition on environmental and geographical gradients. *Annals of the Missouri Botanical Garden* 75:1–34.
- GENTRY, A. H. 1991. The distribution and evolution of climbing plants. Pp. 3–49 in Putz, F. E. & Mooney, H. A. (eds.). *The biology of vines*. Cambridge University Press, Cambridge.
- GENTRY, A. H. & TERBORGH, J. 1990. Composition and dynamics of the Cocha Cashu “mature” floodplain forest. Pp. 542–564 in Gentry, A. H. (ed.). *Four neotropical rainforests*. Yale University Press, New Haven.
- GEROLD, G. 2003. La base para la biodiversidad: el suelo. Pp. 18–31 in Ibsch, P. L. & Mérida, G. (eds.). *Biodiversidad: la riqueza de Bolivia. Estado de conocimiento y conservación*. Ministerio de Desarrollo Sostenible. Editorial FAN, Santa Cruz, Bolivia.
- IBARRA-MANRIQUEZ, G. & MARTÍNEZ-RAMOS, M. 2002. Landscape variation of liana communities in a neotropical rain forest. *Plant Ecology* 160:91–112.
- IPCC. 2007. *Climate change 2007: Impacts, Adaptation and Vulnerability*. Fourth Assessment Report (AR4). Published for the Intergovernmental Panel on Climate Change. Cambridge University Press, Cambridge. 976 pp.
- JHA, C. S. & SINGH, J. S. 1990. Composition and dynamics of dry tropical forest in relation to soil texture. *Journal of Vegetation Science* 1:609–614.
- KAHN, F. & HENDERSON, A. 1999. An overview of the palms of the várzea in the Amazon region. *Advances in Economic Botany* 13:187–193.
- KILLEEN, T. J., JARDIM, A., MAMANI, F. & ROJAS, N. 1998. Diversity, composition and structure of a tropical semideciduous forest in the Chiquitania region of Santa Cruz, Bolivia. *Journal of Tropical Ecology* 14:803–827.
- LAURANCE, W. F., PÉREZ-SALICRUP, D., DELAMÓNICA, P., FEARNside, P. M., D'ANGELO, S., JEROZOLINSKI, A., POHL, L. & LOVEJOY, T. E. 2001. Rain forest fragmentation and the structure of Amazonian liana communities. *Ecology* 82:105–116.
- LEWIS, S. L., PHILLIPS, O. L., BAKER, T. R., LLOYD, J., MALHI, Y., ALMEIDA, S., HIGUCHI, N., LAURANCE, W. F., NEILL, D. A., SILVA, J. N. M., TERBORGH, J., LEZAMA, A. T., MARTÍNEZ, R. V., BROWN, S., CHAVE, J., KUEBLER, C., VARGAS, P. N. & VINCETI, B. 2004. Concerted changes in tropical forest structure and dynamics: evidence from 50 South American long-term plots. *Philosophical Transactions of the Royal Society of London B Biological Sciences*. 359:421–436.
- LICONA-VASQUEZ, J. C., PEÑA-CLAROS, M. & MOSTACEDO, B. 2007. *Composición florística, estructura y dinámica de un bosque amazónico aprovechado a diferentes intensidades en Pando, Bolivia*. BOLFOR & IBIF, Santa Cruz, Bolivia. 49 pp.
- LIEBERMAN, D., LIEBERMAN, M., PERALTA, R. & HARTSHORN, G. 1996. Tropical forest structure and composition on a large-scale altitudinal gradient in Costa Rica. *Journal of Ecology* 84:137–152.
- LOSOS, E., ASHTON, P. S., BROKAW, N., BUNYAVEJCHEWIN, S., CONDIT, R., CHUYONG, G., CO, L., DATTARAJA, H. S., DAVIES, S., ESUFALI, S., EWANGO, C., FOSTER, R., GUNATILLEKE, N., GUNATILLEKE, S., HART, T., HERNANDEZ, C., HUBBELL, S., ITOH, A., JOHN, R., KANZAKI, M., KENFACK, D., KIRATIPRAYOON, S., LAFRANKIE, J., LEE, H. S., LIENGOLA, I., LAO, S., LOSOS, E., MAKANA, J. R., MANOKARAN, N., NAVARRETE, H., OHKUBO, T., PÉREZ, R., PONGPATTANANURAK, N., SAMPER, C., KRIANGSAK SRI-NGERNYUANG, S., SUKUMAR, R., SUN, I. F., SURESH, H. S., TAN, S., THOMAS, D., THOMPSON, J., VALLEJO, M., VILLA MUÑOZ, G., VALENCIA, R., YAMAKURA, T. & ZIMMERMAN, J. 2004. The structure of tropical forests. Pp. 69–78 in Losos, E. C. & Leigh, E. G. (eds.). *Tropical forest diversity and dynamism: findings from a large-scale plot network*. The University of Chicago Press, Chicago.

- MADEIRA, B. G., ESPÍRITO-SANTO, M. M., NETO, S. D., NUNES, Y. R. F., AZOFEITA, G. A. S., FERNANDES, G. W. & QUESADA, M. 2009. Changes in tree and liana communities along a successional gradient in a tropical dry forest in south-eastern Brazil. *Plant Ecology* 201:291–304.
- MALHI, Y., PHILLIPS, O. L., LLOYD, J., BAKER, T., WRIGHT, J., ALMEIDA, S., ARROYO, L., FREDERIKSEN, T., GRACE, J., HIGUCHI, N., KILLEEN, T., LAURANCE, W. F., LEANO, C., LEWIS, S., MEIR, P., MONTEAGUDO, A., NEILL, D., NÚÑEZ VARGAS, P., PANFIL, S. N., PATIÑO, S., PITMAN, N., QUESADA, C. A., RUDAS-LL, A., SALOMAO, R., SALESKA, S., SILVA, N., SILVEIRA, M., SOMBROEK, W. G., VALENCIA, R., VÁSQUEZ MARTÍNEZ, R., VIEIRA, I. C. G. & VINCETI, B. 2002. An international network to monitor the structure, composition and dynamics of Amazonian forests (RAINFOR). *Journal of Vegetation Science* 13:439–450.
- MASCARO, J., SCHNITZER, S. A. & CARSON, W. P. 2004. Liana diversity, abundance, and mortality in a tropical wet forest in Costa Rica. *Forest Ecology and Management* 190:3–14.
- MILLIKEN, W. 1998. Structure and composition of one hectare of central Amazonian terra firme forest. *Biotropica* 30:530–537.
- MURPHY, P. G. & LUGO, A. E. 1986. Ecology of tropical dry forest. *Annual Review of Ecology and Systematics* 17:67–88.
- NEBEL, G., KVIST, L. P., VANCLAY, J. K., CHRISTENSEN, H., FREITAS, L. & RUIZ, J. 2001. Structure and floristic composition of flood plain forests in the Peruvian Amazon. 1. Overstorey. *Forest Ecology and Management* 150:27–57.
- NEWBERY, D. M., CAMPBELL, E. J. F., LEE, Y. F., RIDSDALE, C. E. & STILL, M. J. 1992. Primary lowland dipterocarp forest at Danum Valley, Sabah, Malaysia: structure, relative abundance and family composition. *Philosophical Transactions: Biological Sciences, The Royal Society* 335:341–356.
- PAOLI, G. D., CURRAN, L. M. & SLIK, J. W. F. 2008. Soil nutrients affect spatial patterns of aboveground biomass and emergent tree density in southwestern Borneo. *Oecologia* 155:287–299.
- PARTHASARATHY, N., MUTHURAMKUMAR, S. & REDDY, M. S. 2004. Patterns of liana diversity in tropical evergreen forest of Peninsular India. *Forest Ecology and Management* 190:15–31.
- PÉREZ-SALICRUP, D. R., SORK, V. L. & PUTZ, F. E. 2001. Lianas and trees in a liana forest of Amazonian Bolivia. *Biotropica* 33:34–47.
- POORTER, L., HAWTHORNE, W., BONGERS, F. & SHEIL, D. 2008. Maximum size distributions in tropical forest communities: relationships with rainfall and disturbance. *Journal of Ecology* 96:495–504.
- POULSEN, A. D., TUOMISTO, H. & BALSLEV, H. 2006. Edaphic and floristic variation within a 1-ha plot of lowland Amazonian rain forest. *Biotropica* 38:468–478.
- PROCTOR, J., ANDERSON, J. M., CHAI, P. & VALLACK, H. W. 1983. Ecological studies in 4 contrasting lowland rain forests. *Journal of Ecology* 71:237–260.
- PUTZ, F. E. 1984. The natural history of lianas on Barro Colorado Island, Panama. *Ecology* 65:1713–1724.
- PUTZ, F. E. & CHAI, P. 1987. Ecological studies of lianas in Lambir National Park, Sarawak, Malaysia. *Journal of Ecology* 75:523–531.
- RICHARDS, P. W. 1996. *The tropical rain forest: An ecological study*. Cambridge University Press, Cambridge. 575 pp.
- SAXE, H., CANNELL, M. G. R., JOHNSEN, B., RYAN, M. G. & VOURLITIS, G. 2001. Tree and forest functioning in response to global warming. *New Phytologist* 149:369–399.
- SCHNITZER, S. A. 2005. A mechanistic explanation for global patterns of liana abundance and distribution. *American Naturalist* 166:262–276.
- SCHNITZER, S. A. & BONGERS, F. 2002. The ecology of lianas and their role in forests. *Trends in Ecology and Evolution* 17:223–230.
- SESNIE, S. E., FINEGAN, B., GESSLER, P. E. & RAMOS, Z. 2009. Landscape-scale environmental and floristic variation in Costa Rican old-growth rain forest remnants. *Biotropica* 41:16–26.
- SPIES, T. A. 1998. Forest structure: a key to the ecosystem. *Northwest Science* 72:34–39.
- SUÁREZ-SORUCO, M. 2000. *Compendio de la Geología de Bolivia*. Cochabamba, Bolivia. 114 pp.
- SWAINE, M. D. & GRACE, J. 2007. Lianas may be favored by low rainfall: evidence from Ghana. *Plant Ecology* 192:271–276.
- SWAINE, M. D., LIEBERMAN, D. & HALL, J. B. 1990. Structure and dynamics of a tropical dry forest in Ghana. *Vegetatio* 88:31–51.
- TAKYU, M., KUBOTA, Y., AIBA, S., SEINO, T. & NISHIMURA, T. 2005. Pattern of changes in species diversity, structure and dynamics of forest ecosystems along latitudinal gradients in East Asia. *Ecological Research* 20:287–296.
- TER STEEGE, H., PITMAN, N., SABATIER, D., CASTELLANOS, H., VAN DER HOUT, P., DALY, D. C., SILVEIRA, M., PHILLIPS, O., VÁSQUEZ, R., VAN ANDEL, T., DUIVENVOORDEN, J., DE OLIVEIRA, A. A., EK, R., LILWAH, R., THOMAS, R., VAN ESSEN, J., BAIDER, C., MAAS, P., MORI, S., TERBORGH, J., VARGAS, P. N., MOGOLLÓN, H. & MORAWETZ, W. 2003. A spatial model of tree α -diversity and tree density for the Amazon. *Biodiversity and Conservation* 12:2255–2277.
- TOLEDO, M. 2010. *Neotropical lowland forests along environmental gradients*. PhD thesis, Wageningen University, the Netherlands.
- TOLEDO, M., POORTER, L., PEÑA-CLAROS, M., ALARCÓN, A., BALCÁZAR, J., CHUVIÑA, J., LEAÑO, C., LICONA, J. C., TER STEEGE, H. & BONGERS, F. 2011a. Patterns and determinants of floristic variation across lowland forests of Bolivia. *Biotropica* doi:10.1111/j.1744-7429.2010.00711.x
- TOLEDO, M., POORTER, L., PEÑA-CLAROS, M., ALARCÓN, A., BALCÁZAR, J., LEAÑO, C., LICONA, J. C., LLANQUE, O., VROOMANS, V., ZUIDEMA, P. & BONGERS, F. 2011b. Climate is a stronger driver of tree and forest growth rates than soil and disturbance. *Journal of Ecology* 99:254–264.
- USLAR, I., MOSTACEDO, B. & SALDIAS, M. 2004. Composición, estructura y dinámica de un bosque seco semideciduo en Santa Cruz, Bolivia. *Ecología en Bolivia* 39:25–43.
- VELARDE, M. J. & MORAES, M. 2008. Densidad de individuos adultos y producción de frutos del asaí (*Euterpe precatória*, *Arecaceae*) en Riberalta, Bolivia. *Ecología en Bolivia* 43:99–110.
- VIEIRA, S., BARBOSA, DE, CAMARGO, P., SELHORST, D., DA SILVA, R., HUTYRA, L., CHAMBERS, J. Q., BROWN, I. F.,

- HIGUCHI, N., DOS SANTOS, J., WOFSY, S. C., TRUMBORE, S. E. & MARTINELLI, L. A. 2004. Forest structure and carbon dynamics in Amazonian tropical rain forests. *Oecologia* 140:468–479.
- VORMISTO, J. 2002. Palms as rainforest resources: how evenly are they distributed in Peruvian Amazonia? *Biodiversity and Conservation* 11:1025–1045.
- VORMISTO, J., SVENNING, J. C., HALL, P. & BALSLEV, H. 2004. Diversity and dominance in palm (Arecaceae) communities in terra firme forests in the western Amazon basin. *Journal of Ecology* 92:577–588.
- WHITE, D. A. & HOOD, C. S. 2004. Vegetation patterns and environmental gradients in tropical dry forests of the northern Yucatan Peninsula. *Journal of Vegetation Science* 15:151–160.

Appendix 1. Pearson correlations between climate variables, soil properties and environmental PCA axes. n = 220 * P ≤ 0.05, ** P ≤ 0.01. Driest = precipitation of the driest months, Temp = temperature, CEC = Cation Exchange Capacity, OM = organic matter.

	Driest	Temp	pH	Ca	Mg	Na	K	CEC	Acidity	P	OM	N	Sand	Silt	Clay	Rainfall axis	Temp. axis	Fertility axis	Texture axis	
Precipitation	0.52**	0.66**	-0.53**	-0.38**	-0.29**	-0.17*	-0.40**	-0.24**	0.56**	0.05	-0.51**	-0.52**	-0.34**	0.28**	0.24**	0.93**	0.26**	-0.35**	0.49**	
Driest		-0.05	0.04	0.19**	0.27**	0.49**	0.29**	0.28**	-0.03	0.39**	-0.09	-0.06	-0.41**	0.58**	0.01	0.74**	-0.63**	0.33**	0.31**	
Temp.			-0.63**	-0.56**	-0.36**	-0.34**	-0.48**	-0.44**	0.57**	-0.28**	-0.64**	-0.62**	-0.11	-0.10	0.30**	0.55**	0.78**	-0.57**	0.36**	
pH				0.85**	0.48**	0.54**	0.49**	0.69**	-0.71**	0.53**	0.48**	0.67**	0.19**	0.13*	-0.47**	-0.43**	-0.48**	0.73**	-0.50**	
Ca					0.60**	0.69**	0.64**	0.94**	-0.55**	0.69**	0.58**	0.72**	-0.10	0.35**	-0.24**	-0.26**	-0.52**	0.91**	-0.22**	
Mg						0.60**	0.67**	0.65**	-0.48**	0.31**	0.31**	0.38**	-0.27**	0.37**	0.01	-0.10	-0.49**	0.73**	0.04	
Na							0.76**	0.73**	-0.40**	0.48**	0.42**	0.53**	-0.32**	0.46**	0.00	0.05	-0.58**	0.83**	0.08	
K								0.69**	-0.48**	0.32**	0.63**	0.63**	-0.33**	0.34**	0.15*	-0.18**	-0.59**	0.84**	0.08	
CEC									-0.31**	0.65**	0.58**	0.68**	-0.33**	0.49**	-0.02	-0.11	-0.50**	0.92**	0.06	
Acidity										-0.28**	-0.28**	-0.41**	-0.31**	0.04	0.47**	0.41**	0.46**	-0.52**	0.61**	
P											0.28**	0.39**	-0.12	0.38**	-0.24**	0.13	-0.39**	0.63**	-0.11	
OM												0.89**	-0.12	0.10	0.08	-0.46**	-0.43**	0.69**	-0.09	
N													-0.07	0.13	-0.04	-0.45**	-0.41**	0.78**	-0.20**	
Sand														-0.80**	-0.72**	-0.42**	0.19**	-0.33**	-0.93**	
Silt															0.15*	0.39**	-0.41**	0.50**	0.61**	
Clay																0.24**	0.17*	-0.03	0.82**	
Rainfall axis																			-0.18**	0.51**
Temp. axis																			0.64**	-0.06

Appendix 2. Pearson correlations between forest structural variables. n = 220, * P ≤ 0.05, ** P ≤ 0.01. BA = basal area, SCD = size class distribution. dbh = diameter at breast height. See Table 1 for parameter details.

	Emergent	Canopy	Subcanopy	Total BA	Tree BA	Palm BA	SCD slope	dbh ₅₀	dbh ₉₉	Total density	Tree density	Palm density	Liana density	Liana infestation
Height _{max}	0.03	0.06	-0.13	0.32**	0.21**	0.59**	0.36**	-0.29**	0.41**	0.31**	0.05	0.54**	-0.14*	-0.25**
Emergent		-0.18**	-0.43**	0.13*	0.10	0.09	-0.07	0.43**	-0.08	-0.02	-0.11	0.13	-0.42**	-0.31**
Canopy			-0.78**	-0.15*	-0.22**	0.20**	0.09	-0.08	-0.05	-0.03	-0.15*	0.19**	-0.25**	-0.23**
Subcanopy				-0.01	0.07	-0.28**	-0.07	-0.20**	0.05	-0.01	0.18**	-0.31**	0.52**	0.43**
Total BA					0.97**	0.44**	0.33**	0.05	0.48**	0.56**	0.41**	0.42**	-0.29**	-0.49**
Tree BA						0.21**	0.29**	0.10	0.45**	0.48**	0.45**	0.18**	-0.22**	-0.41**
Palm BA							0.33**	-0.25**	0.36**	0.44**	-0.05	0.94**	-0.36**	-0.44**
SCD-slope								-0.03	0.78**	-0.27**	-0.45**	0.21**	-0.32**	-0.13
dbh ₅₀									-0.07	-0.41**	-0.34**	-0.24**	-0.25**	-0.01
dbh ₉₉										-0.13*	-0.29**	0.20**	-0.21**	-0.13
Total density											0.85**	0.53**	0.05	-0.35**
Tree density												0.01	0.29**	-0.14*
Palm density													-0.37**	-0.46**
Liana density														0.81**

Appendix 3. Pearson correlation coefficients of environmental variables and environmental axes with forest structural variables. n = 220, * P ≤ 0.05, ** P ≤ 0.01. dbh = diameter at breast height, SCD = Size–class distribution, BA = basal area, Temp. = temperature, CEC = Cation Exchange Capacity, OM = organic matter.

	Height _{max}	Emergent	Canopy	Subcanopy	Total BA	SCD slope	dbh ₅₀	dbh ₉₉	Total density	Tree density	Palm density	Liana density	Liana infestation
Precipitation	0.74**	-0.08	0.26**	-0.21**	0.25**	0.42**	-0.44**	0.41**	0.31**	-0.05	0.86**	-0.19**	-0.28**
Driest month	0.36**	0.01	0.16*	-0.16*	0.32**	0.27**	-0.30**	0.20**	0.36**	0.00	0.62**	-0.34**	-0.28**
Temperature	0.56**	0.05	0.38**	-0.40**	0.14*	0.37**	-0.06	0.30**	0.08	-0.13	0.53**	-0.27**	-0.31**
Dry	-0.74**	0.15*	-0.33**	0.23**	-0.22**	-0.45**	0.44**	-0.40**	-0.30**	0.06	-0.85**	0.19**	0.28**
Drought	-0.47**	-0.05	-0.22**	0.24**	-0.29**	-0.24**	0.24**	-0.18**	-0.34**	0.04	-0.68**	0.34**	0.35**
pH	-0.36**	-0.15*	-0.28**	0.36**	-0.22**	-0.03	0.14*	-0.05	-0.32**	-0.12	-0.52**	0.26**	0.49**
Ca	-0.36**	-0.09	-0.24**	0.29**	-0.07	0.07	0.16*	-0.03	-0.24**	-0.09	-0.43**	0.13	0.33**
Mg	-0.37**	-0.04	-0.07	0.13	-0.07	0.04	0.18*	-0.06	-0.22**	-0.12	-0.33**	0.04	0.26**
Na	-0.19**	0.04	0.05	-0.07	0.03	0.09	0.06	0.05	-0.07	-0.12	0.07	-0.19**	0.01
K	-0.35**	0.01	0.05	-0.03	-0.01	-0.12	0.15*	-0.14*	-0.09	-0.02	-0.29**	-0.12	0.01
CEC	-0.16**	-0.01	-0.21**	0.17**	0.12	0.07	0.14*	0.11	-0.05	-0.05	-0.17*	-0.02	0.13*
Acidity	0.51**	0.16*	0.15*	-0.30**	0.32**	0.12	-0.12	0.20**	0.36**	0.12	0.61**	-0.26**	-0.45**
P	0.03	-0.20**	-0.13*	0.22**	0.10	0.29**	0.17*	0.24**	-0.02	-0.09	0.10	-0.01	0.15
OM	-0.41**	-0.16*	-0.12	0.21**	-0.03	-0.27**	0.06	-0.16*	-0.05	0.13*	-0.42**	0.24**	0.23**
N	-0.38**	-0.19**	-0.15*	0.26**	-0.08	-0.13*	0.06	-0.07	-0.15*	0.06	-0.47**	0.26**	0.34**
Sand	-0.35**	-0.13	-0.05	0.14*	-0.36**	-0.18**	0.14*	-0.22**	-0.27**	-0.07	-0.41**	0.31**	0.39**
Silt	0.21**	0.14*	-0.14*	0.01	0.37**	0.18**	-0.11	0.20**	0.20**	0.08	0.29**	-0.26**	-0.26**
Clay	0.29**	0.06	0.18**	-0.21**	0.14*	0.00	-0.06	0.06	0.15*	0.05	0.28**	-0.17**	-0.31**
Rainfall axis	0.72**	-0.04	0.32**	-0.29**	0.31**	0.43**	-0.39**	0.37**	0.36**	-0.06	0.89**	-0.32**	-0.37**
Temp. axis	0.27**	-0.04	0.16*	-0.14*	-0.09	0.14*	0.03	0.15*	-0.13*	-0.07	-0.10	0.06	-0.02
Fertility axis	-0.28**	-0.08	-0.11	0.16*	0.04	0.03	0.09	0.02	-0.12	-0.05	-0.27**	0.01	0.17**
Texture axis	0.46**	0.17*	-0.16*	-0.28**	0.36**	0.16*	-0.14*	0.21**	0.33**	0.09	0.55**	-0.36**	-0.51**